

1 High trophic niche overlap in mixed bird colonies relying on 2 artificial nests

3 João Gameiro^{1*}, Aldina M.A. Franco², Teresa Catry³, Jorge M. Palmeirim¹, Inês Catry^{2,4,5}

4 ¹cE3c – Centre for Ecology, Evolution and Environmental Changes, Faculdade de Ciências,
5 Universidade de Lisboa, Lisboa, Portugal

6 ²School of Environmental Sciences, University of East Anglia, Norwich, UK

7 ³Centro de Estudos do Ambiente e do Mar (CESAM), Departamento de Biologia Animal, Faculdade
8 de Ciências da Universidade de Lisboa, Lisbon, Portugal

9 ⁴CIBIO/InBIO, Centro de Investigação em Biodiversidade e Recursos Genéticos, Laboratório
10 Associado, Universidade do Porto, Campus Agrário de Vairão, 4485-661 Vairão, Portugal

11 ⁵CIBIO/InBIO, Centro de Investigação em Biodiversidade e Recursos Genéticos, Laboratório
12 Associado, Instituto Superior de Agronomia, Universidade de Lisboa, Tapada da Ajuda, 1349-017
13 Lisboa, Portugal

14 *jgameiro92@gmail.com

15

16 **ABSTRACT:** Although successful at recovering endangered populations, conservation
17 actions based on nest provisioning seldom consider how they shape the composition of
18 communities and alter interspecific interactions. Specifically, the extent to which dietary
19 overlap within these communities may affect the conservation of target species has rarely
20 been assessed. In Southern Europe, large-scale nest-site provisioning aimed at recovering
21 Lesser Kestrels *Falco naumanni* populations attracted several bird species forming mixed
22 breeding assemblages, likely promoting interspecific competition for resources during
23 breeding. Here we used Stable Isotope Analysis ($\delta^{15}\text{N}$ and $\delta^{13}\text{C}$) to assess inter- and
24 intraspecific dietary segregation in these assemblages and investigate the mechanisms
25 allowing species coexistence. We examined resource partitioning and trophic niche overlap
26 among Lesser Kestrels, Common Kestrel *Falco tinnunculus*, European Roller *Coracias*
27 *garrulus*, Barn Owl *Tyto alba*, Little Owl *Athene noctua*, and Spotless Starling *Sturnus*
28 *unicolor*; and within species between parents and their offspring. Similar isotope ratios and
29 highly overlapped niches, particularly among Lesser Kestrels, Rollers, and Starlings, suggest
30 limited dietary segregation and use of similar prey. Within species, parent-offspring

31 segregation was marked across all species. Our results indicate that species breeding in these
32 assemblages occupy similar ecological niches, despite a potential increase in competition.
33 High-resource availability in the area may guarantee coexistence but the viability of mixed-
34 species groups may be compromised in areas with limited resources, which are predicted to
35 expand with ongoing human and climate-induced changes. Conservation practices based on
36 nest provisioning need to consider the ecological niches of target and sympatric species as
37 well as their interactions.

38 **Keywords:** isotopic niche, dietary segregation, parent-offspring segregation, interspecific
39 competition, mixed-species colonies, artificial nests

40

41

42 **1. Introduction**

43 Human-induced changes in the environment shape the distribution of many species by
44 restricting their access to resources such as food or nest-sites (Wiegand et al. 2005; Teckentrup
45 et al. 2019). Ongoing biodiversity loss has prompted conservation projects and massive
46 provisioning of artificial nest-sites to halt species decline across many taxa and regions (Harper
47 et al. 2005; Catry et al. 2009; Olah et al. 2014; Sutherland et al. 2018), further re-shaping the
48 nesting landscape and altering the composition and structure of communities (Duckworth et al.
49 2017; Catry & Catry 2019). Nest-site provisioning may modify interspecific interactions and
50 trophic dynamics (Duckworth et al. 2017; Catry & Catry 2019), but their consequences are
51 often overlooked by researchers or conservation managers (Mainwaring et al. 2015).

52 According to the niche theory and the competitive exclusion principle, species with similar
53 niches, likely to compete with one another, are not expected to coexist (Hardin 1960; Pianka
54 1981). In the event this new and artificial nesting landscape promotes the aggregation of
55 potential competitor species, artificial nest-sites may turn into ecological traps by reducing
56 long-term individual fitness associated with a reduction of resource availability or increased
57 interspecific aggressive behaviours (Kappeler et al. 2015; Catry & Catry 2019).

58 To avoid competition, especially when resources are limited, coexisting species can segregate
59 in at least one of three main axes: space, time, and diet (Pianka 1981). Segregation in spatio-
60 temporal habitat utilisation occurs when species forage in different locations (like in
61 insectivorous desert bats, Razgour et al. 2011), different periods of the day (like in African
62 carnivore communities, Schuette et al. 2013), or when they have different breeding
63 phenologies (like in sympatric penguin species, Clewlow et al. 2019). Segregation in the
64 dietary axis occurs when species reduce their dietary niche overlap by foraging for alternative
65 food resources (Macarthur & Pianka 1966; Jenkins & Davoren 2020; Mansor et al. 2021).
66 Species may also contract their dietary niches to focus on certain prey types that become

67 temporarily super-abundant, which increases the dietary niche overlap among species without
68 increasing competition (Forero et al. 2004; Charter et al. 2018; Denhard et al. 2020). Dietary
69 niche segregation can also occur within species to reduce intraspecific competition, through
70 dietary sexual or parent-offspring segregation, especially when the potential for competitive
71 interactions is maximum, such as for central place foragers during reproduction (Orłowski et
72 al. 2014; Catry et al. 2016a; Reisinger et al. 2020).

73 Here, we investigate dietary niche overlap in a multi-specific bird community created by
74 large-scale nest-site provisioning aimed at recovering the colonial Lesser Kestrel *Falco*
75 *naumanni* in southern Portugal (Catry et al. 2009). In a treeless landscape with low nesting
76 opportunities, many bird species took advantage of these artificial breeding structures – some
77 with more than 80 cavities, forming mixed-species colonies (Catry & Catry 2019). These
78 include Lesser Kestrels, Common Kestrels *Falco tinnunculus*, European Rollers *Coracias*
79 *garrulus*, Barn Owls *Tyto alba*, Little Owls *Athene noctua*, Jackdaws *Corvus monedula*,
80 Spotless Starling *Sturnus unicolor*, and Feral Pigeons *Columba livia* (Catry & Catry 2019).
81 The dietary habits of most of these species – large arthropods to small mammals – suggest
82 that their use of trophic resources may overlap to a large extent (Jaksić et al. 1982; Motis et
83 al. 1997; Tomé et al. 2008; Catry et al. 2016a; Orihuela-Torres et al. 2017; Catry et al. 2019),
84 especially during the breeding season, when competition for food should peak as parents need
85 to obtain resources to fulfil their own energy requirements and those of their offspring
86 (Orłowski et al. 2014; Antón-Tello et al. 2021). High densities of breeding pairs could lead to
87 prey depletion in the vicinity of the colonies, further increasing inter and intraspecific
88 competition (sensu Ashmole's halo; Birt et al. 1987; Dehnhard et al. 2020; Jenkins and
89 Davoren, 2020).

90 We used Stable Isotope Analysis (SIA) to investigate inter and intraspecific overlap in the
91 diet of six bird species breeding in mixed colonies installed in artificial breeding structures to

92 better understand the mechanisms allowing their coexistence. Specifically, using stable
93 isotope ratios of carbon ($\delta^{13}\text{C}$) and nitrogen ($\delta^{15}\text{N}$), we aimed to: (1) compare isotopic niche
94 breadth and overlap among different species and (2) assess parent-offspring diet segregation
95 of each species within the assemblages. We sampled feathers from nestlings and blood
96 plasma from parent-offspring pairs during the breeding season, thus providing a detailed
97 spectrum of the dietary resources used by these species. We predict species with similar diets
98 and similar trophic positions to show higher dietary overlap, while species at the top and
99 bottom ends of the trophic spectrum to be more segregated from the rest of the assemblage.
100 Altogether, we aimed at understanding the extent on how dietary overlap and niche
101 segregation among potential competitors may affect the conservation of targeted species, an
102 issue seldom acknowledged by conservation projects.

103

104 **2. Methods:**

105 *2.1 Study area and study species*

106 This study was carried out in the Castro Verde Special Protection Area (SPA, Natura 2000),
107 Portugal (37°41'N, 8°05'W). With ca. 85,000 ha, this is the main Portuguese area of agro-
108 steppes – a semi-natural habitat created by low intensity farming practices – and one of the
109 main strongholds for several threatened farmland bird species in Western Europe (Moreira et
110 al. 2007; BirdLife International 2021). Land use within the study area has remained relatively
111 stable in the last decades, in part due to the implementation of agri-environmental policy
112 schemes and funding mechanisms that ensure high-quality foraging habitat for many
113 farmland birds (Catry et al., 2013; Silva et al., 2018). Mixed-species colonies are mostly
114 found in artificial nesting structures provided by LIFE conservation projects target at Lesser
115 Kestrels during the early 2000s (Catry et al. 2009), and to a lesser extent in abandoned

116 farmland buildings, with nests inside wall cavities or under roof tiles. Within these breeding
117 assemblages, we focused on species that prey mainly on small vertebrates and arthropods, as
118 these are the ones with higher potential for having overlapped trophic niches: Barn and Little
119 Owls, Common and Lesser Kestrels, European Rollers (hereafter Rollers) and Spotless
120 Starlings (hereafter Starlings). Feral Pigeons and Jackdaws, due to the relatively large
121 proportion of plant sources in their diets (cereals; Murton & Westwood 1966; Högstedt 1980;
122 Soler et al. 1990), were not included in the study.

123 ***2.2. Sample collection***

124 Feather sampling was carried out during the breeding seasons of 2014, 2016 and 2018 (Table
125 1). Nestlings of Barn Owls (n=12), Little Owls (n=19), Common Kestrels (n=33), Lesser
126 Kestrels (n=36), Rollers (n=36), and Starlings (n=12) were sampled in their nests at the age
127 of 3-4 weeks, when breast feathers are well developed. Approximately 3-5 breast feathers
128 were plucked (keeping the feather base intact) and stored in separate plastic zip bags. Blood
129 sampling of parent-offspring pairs – one parent and one nestling – of Little Owls (n=7),
130 Common Kestrels (n=8), Lesser Kestrels (n=11), Rollers (n=6), and Starlings (n=6) was
131 carried out during the breeding season of 2019 (Table 1). Parents and their offspring were
132 sampled within a close temporal window (6.9 ± 6.1 days) at their nests. Approximately 150
133 μL of blood was collected from the brachial vein and preserved in separate vials. The blood
134 was centrifuged for 10 min at 3400 rev/min within 3 hours of collection, to separate plasma
135 from red blood cells, and plasma samples were frozen before being further prepared for stable
136 isotope analysis (red blood cells were not analysed). We were unable to sample enough Barn
137 Owl adults, and so parent-offspring segregation was not investigated for this species.

138 Additionally, to aid the interpretation of stable isotope data, we sampled key prey types of
139 each species in mixed-species colonies throughout the breeding season of 2018 (Table 2).

140 Prey species were identified and collected inside or near the nests and were kept frozen until
141 processed for isotope analysis. All work involving bird handling and sample collection was
142 approved by the Instituto de Conservação da Natureza e Florestas (ICNF).

143

144 **2.3. *Stable isotope analysis***

145 Stable isotope ratios have been widely used as dietary tracers to assess the different habitats
146 and diet resources used by consumers, providing a proxy for species or individual ecological
147 niches using low invasive methods (Forero et al. 2004; Catry et al. 2019; Dehnhard et al.
148 2020). SIA of nitrogen ($\delta^{15}\text{N}$) and carbon ($\delta^{13}\text{C}$) provides information on the trophic position
149 of an organism in the food chain and on habitat use, respectively, allowing researchers to
150 simultaneously examine the dietary niche segregation/overlap of multiple species or
151 individuals within assemblages and infer on their competition for resources (Inger & Bearhop
152 2008; Alonso et al. 2012; Catry et al. 2016b; Reisinger et al. 2020; Jenkins and Davoren,
153 2020). SIA also provides information on the dietary niche at multiple temporal scales
154 depending on the tissue sampled, as isotope ratios reflect the diet during which that tissue was
155 metabolically active.

156 Here, we use feathers from nestlings of the six species to examine isotopic breadth and
157 overlap among species, as they offer information for the period during which the feather was
158 being formed, adequately characterising the diet of the nestlings during the breeding season
159 (Inger & Bearhop 2008; Weiss et al. 2009). Feathers were washed in double baths of 0.25N
160 sodium hydroxide solution alternated with baths of double distilled water to remove any
161 surface contaminants, and were then air dried and then cut into small fragments (Catry et al.
162 2008). To investigate parent-offspring dietary segregation, we used blood plasma. Blood
163 plasma has a much higher turnover rate, reflecting the averaged diet over much shorter

164 periods (days), and thus allows for a proper comparison between parent and offspring
165 signatures (days; Inger & Bearhop 2008; Reisinger et al. 2020). Blood plasma from adult
166 birds was additionally used to infer niche overlap between species in addition to nestling
167 feathers. Information from nestling plasma was only used to examine parent-offspring
168 segregation, as it does not add information to that obtained with nestling feathers (sample size
169 was smaller and restricted to one year, and plasma reflects information on the birds' diet over
170 a much shorter period). Plasma samples of birds and soft tissues (muscle) of prey were dried
171 in an oven at 60°C for 48h and then reduced to a homogenised powder. Prey samples were
172 then processed for lipid extraction by immersion in a 2:1 chloroform/methanol solution with
173 a solvent volume three to five times larger than the sample volume (Logan et al. 2008).
174 Samples were then mixed for 30s, left undisturbed for ca. 30 min, further centrifuged for 10
175 min at 3400rev/min, and the supernatant containing solvent and lipids were removed. This
176 process was repeated at least three times (until the solvent was clear) and samples were then
177 re-dried at 60°C for 24h to remove any remaining solvent. Between 0.8 to 1.2mg of each
178 sample (feathers, plasma, and prey) were weighted and stored in tin cups for stable carbon
179 and nitrogen isotope assays. Isotopic ratios were determined by continuous-flow isotope ratio
180 mass spectrometry at the “Stable Isotopes and Instrumental Analysis” facility of the faculty
181 of Sciences, University of Lisbon. Results are presented conventionally as δ values in parts
182 per thousand (‰) relative to the IAEA CH7 and Glucose BCR for $\delta^{13}\text{C}$, and USGS25 and
183 USGS35 for $\delta^{15}\text{N}$. The precision of the isotope ratio analysis, calculated using values from
184 six to nine replicates of laboratory standard material (casein) interspersed among samples in
185 every bath analysis, was 0.04-0.05‰ and 0.02-0.04‰ for $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$, respectively.

186 ***2.4 Data analysis***

187 Feather isotopic signatures for each species from all years were pooled together as the
188 relative isotopic position of each species remained similar (Table 1; Fig. S1).

189 To search for differences in isotopic signatures among species, pairwise multivariate analyses
190 of variance were performed on nitrogen and carbon simultaneously. Then analyses of
191 variance were performed on carbon and nitrogen separately, followed by pairwise
192 comparisons between each pair of species. Each set of analysis were performed for both
193 nestling feathers and adult plasma. Non-parametric tests were used for nestlings
194 (PERMANOVA, Kruskal-Wallis, and Wilcoxon rank sum tests), and parametric tests for
195 adults (MANOVA, ANOVA, and Tukey tests), after evaluating for data normality. Prey
196 items with similar isotopic signals were pooled within each taxonomic order (SIA group,
197 Table 2), and then plotted in the isotopic biplot graph to aid the interpretation of consumer
198 signatures using trophic discrimination factors (TEF). Due to lack of reported discrimination
199 factors for all consumer species, we used the mean (\pm SD) value of trophic discrimination
200 factors for feather and plasma samples of other bird species reported in published studies
201 reviewed by Caut et al. (2009) (feathers: $\delta^{13}\text{C} = 2.16 \pm 1.53$, $\delta^{15}\text{N} = 2.84 \pm 1.14$; plasma: $\delta^{13}\text{C}$
202 $= -0.08 \pm 0.85$, $\delta^{15}\text{N} = 2.82 \pm 0.31$).

203 To measure isotopic niche breadth of each species and estimate the degree of niche overlap
204 between species, corrected and Bayesian standard ellipse areas (SEAc and SEAB ,
205 respectively) were calculated using the SIBER package for R (Jackson & Parnell 2020).
206 SEAB were used to quantitatively compare niche breadth among species, using 1 000 000
207 iterations, three chains, a burn-in of 1000 and thinning of 10, using a vague normal prior
208 (Jackson & Parnell 2020). Overlap among standard ellipses (SEAc) for all pairs of species
209 and for parents and offspring within each species were estimated to reflect the degree of
210 trophic resources shared. For each species (i) in one pair (i,j), a value of overlap ($\text{Ov}_{[i]}$) was
211 calculated as the ratio between the area of overlap between the two SEAc ($A_{[i,j]}$) and its own
212 SEAc ($A_{[i]}$), expressed as a proportion ($\text{Ov}_{[i]} = A_{[i,j]}/A_{[i]}$; Catry et al. 2016b). Average overlap
213 per species was estimated as the average of all overlaps involving that species. Additionally,

214 to evaluate parent-offspring segregation within each species, differences in Euclidean
215 distances between the centroids of parents and offspring were calculated and tested using a
216 residual permutation procedure based on Turner et al. (2010). Afterward, differences in $\delta^{15}\text{N}$
217 and $\delta^{13}\text{C}$ between parents and offspring were compared with t-tests or Mann-Whitney U tests
218 for parametric and non-parametric data, respectively. Assumptions of each test were met
219 prior analysis. All analyses were conducted in R statistical environment (R Development
220 Core Team, 2016).

221 **3. Results**

222 *3.1 Stable isotope signatures*

223 Feathers were collected from a total of 148 nestlings and plasma samples from 38 parent-
224 offspring pairs from six different species; mean (\pm SD) stable isotope signatures ($\delta^{13}\text{C}$, $\delta^{15}\text{N}$)
225 for each species are presented in Table 1 and Figure 1. A PERMANOVA on $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ of
226 nestling feathers revealed differences among species ($F_{(5,142)} = 23.618$, $p = 0.001$), both in
227 $\delta^{13}\text{C}$ and in $\delta^{15}\text{N}$ (Kruskal-Wallis tests $\delta^{13}\text{C}$: $\chi^2_{(5,142)} = 15.948$, $p = 0.007$; $\delta^{15}\text{N}$: $\chi^2_{(5,142)} =$
228 87.323 , $p < 0.001$). Post-hoc tests indicated that $\delta^{13}\text{C}$ only differed between Lesser Kestrels
229 and Rollers ($p = 0.02$). Barn Owls and Little Owls fed at the highest trophic level (most
230 enriched $\delta^{15}\text{N}$ values) of the community, significantly higher than all other species (all $p <$
231 0.005). Common Kestrels had more enriched $\delta^{15}\text{N}$ values than Lesser Kestrels and Rollers,
232 (both $p < 0.001$), but similar values to Starlings ($p = 0.257$). Starlings also had more enriched
233 $\delta^{15}\text{N}$ than Lesser Kestrels ($p = 0.016$) and Rollers ($p < 0.001$) and Lesser Kestrels had higher
234 values than Rollers ($p = 0.012$). Regarding the adult community, blood isotopic signatures
235 were only different among species for $\delta^{15}\text{N}$ values (MANOVA: $F_{(4,33)} = 4.97$, $p <$
236 0.001 ; ANOVA $\delta^{13}\text{C}$: $F_{(4,33)} = 2.038$, $p = 0.112$; $\delta^{15}\text{N}$: $F_{(4,33)} = 10$, $p < 0.001$). The trophic rank
237 of adults was similar to that of nestlings. Little Owls and Common Kestrels had the highest

238 $\delta^{15}\text{N}$ values (adult Barn Owls were not sampled), with Little Owls being in a significantly
239 higher position than Lesser Kestrels, Rollers, and Starlings (all $p < 0.001$), while Common
240 Kestrels had only significantly higher $\delta^{15}\text{N}$ values than Starlings (p-values: Lesser Kestrel =
241 0.061, Roller = 0.068, Starling = 0.030). Adult Lesser Kestrels, Rollers and Starlings had
242 similar $\delta^{15}\text{N}$ values (all $p > 0.900$; Fig. 1B). In short, trophic position (according to $\delta^{15}\text{N}$
243 values) of nestlings and adults in the community was Barn Owl = Little Owl > Common
244 Kestrel = Starling > Lesser Kestrel > Roller; and Little Owl = Common Kestrel > Starling =
245 Lesser Kestrel = Roller; respectively.

246 Isotopic signature of main prey items (corrected by TEF) can be found on Figure 1. The 10
247 potential prey groups comprise a broad isotopic space, with mean $\delta^{13}\text{C}$ values ranging from -
248 27.3 ± 0.6 to 23.2 ± 0.3 and mean $\delta^{15}\text{N}$ from 3.4 ± 1.5 to 10.9 ± 3.0 (Table 2; Fig. 1). The
249 relative trophic position of the six species and their relation to prey suggest that Common
250 Kestrels, Lesser Kestrels, Rollers, and Starlings feed their offspring mostly with Orthoptera
251 and to a lesser extent with Coleoptera and vertebrates, while the opposite occurs in Barn and
252 Little Owls (Fig. 1A). Regarding adults, Little Owls and Common Kestrels had closer
253 signatures to vertebrates, while Lesser Kestrels, Rollers, and Starlings maintained a lower
254 trophic position.

255

256

257 ***3.2 Interspecific isotopic niche overlap***

258 As observed for the consumers' isotopic space described above, differences in trophic niches
259 were found mainly between species likely including vertebrates in their diet, i.e., Barn Owls
260 and Little Owls, and those preying mostly on arthropods, such as Lesser Kestrels, Rollers,
261 and Starlings; with Common Kestrels in an intermediate position (Fig. 2). Amongst nestlings,

262 Little Owls had the widest niche breadth ($SEA_B = 4.352$; Table 3), only slightly overlapping
263 with Common Kestrels and Starlings (23% and 12%, respectively; Table S1) but totally
264 overlapping with Barn Owls ($SEA_B = 0.755$; Table 3) (Fig. 2A). Common Kestrels and
265 Starlings also showed a high overlap with each other (63% and 75%) as well as with Little
266 Owls (59% and 37%, respectively; Table S1). Lesser Kestrel nestlings had the smallest niche
267 breadth ($SEA_B = 0.505$; Table 3) which highly overlapped with the dietary niche of Rollers,
268 Common Kestrels, and Starlings (60%, 40% and 60%, respectively; Table S1). Amongst
269 adults, Lesser Kestrels, Rollers, and Starlings showed high niche overlap, while Common
270 Kestrels and Little Owls were clearly segregated (Fig. 2B; Table S1).

271

272

273 ***3.3 Parent-offspring segregation***

274 Differences between adults and nestlings were found in $\delta^{15}N$ values for every species except
275 Little Owls (Little Owl: $w_{(12)} = 38$, $p = 0.097$; Common Kestrel: $w_{(14)} = 68$, $p = 0.001$; Lesser
276 Kestrel $w_{(20)} = 107$, $p = 0.001$; Roller: $t_{(10)} = 2.753$, $p = 0.025$; Starling: $w_{(10)} = 34$, $p = 0.009$).
277 No differences in $\delta^{13}C$ values were found for any species (Little Owl: $t_{(12)} = -1.961$, $p =$
278 0.075 ; Common Kestrel: $t_{(14)} = 0.381$, $p = 0.709$; Lesser Kestrel: $t_{(20)} = 1.256$, $p = 0.228$;
279 Roller: $t_{(10)} = 0.115$, $p = 0.115$; Starling: $t_{(10)} = -0.083$, $p = 0.936$). Despite this, Euclidean
280 distance between centroid location of adults and nestlings was significantly different in all
281 species (all $p < 0.01$; Table S2), indicating a clear parent-offspring segregation in all of them
282 (Fig. 3).

283

284

285 **4. Discussion**

286 Identifying the dietary niche of species within assemblages is crucial to understand the
287 mechanisms allowing the coexistence of potential competitors in a specific area and draw
288 adequate conservation management actions. Here, we used carbon and nitrogen stable isotope
289 ratios from nestlings and from parent-offspring pairs to provide a detailed assessment of the
290 dietary resources used by a group of six bird species breeding in dense mixed-species
291 colonies established as the result of an extensive nest-provisioning program (Catry et al.
292 2009; Catry & Catry, 2019). Overall, our results show a high intraspecific (parent-offspring)
293 but low interspecific dietary segregation in these assemblages, suggesting a high potential for
294 interspecific competition for resources during the breeding season.

295 ***4.1 Interspecific overlap in isotopic niches***

296 Identical isotopic signatures and highly overlapping niches suggest that species breeding in
297 these assemblages rely on similar prey to feed their offspring. As reported in previous studies,
298 the low $\delta^{15}\text{N}$ isotopic signatures of Lesser Kestrel and Roller nestlings likely reflect a diet
299 based on Orthoptera such as grasshoppers and bush-cricket (> 50% and 70% of all prey
300 delivered to Lesser Kestrel and Roller nestlings in the study area, respectively; Catry et al.
301 2016a, 2019). Thus, the high overlap observed between the two species was expected. Little
302 Owl nestlings had the largest isotopic niche breadth in our study, overlapping with Common
303 Kestrels but segregating from Lesser Kestrels and Rollers. Common Kestrels and Little Owls
304 have more generalist diets that include vertebrates but may expand their niche to consume
305 more abundant prey such as beetles and grasshoppers, especially during chick rearing (Tomé
306 et al. 2008; Orihuela-Torres et al. 2017). Remains of birds, amphibians, reptiles, and small
307 mammals, but also of Orthoptera, were frequently found in Little Owl and Common Kestrel
308 nests in the study area during chick rearing, confirming their eclectic diet. Nestlings of Barn

309 Owls had the most enriched nitrogen signatures of the assemblage, but their dietary niche
310 breadth suggests a more specialised, vertebrated-based diet (Jacksić et al. 1982; Riegert et al.
311 2021; authors pers.obs), segregating from all other species except Little Owls. Regarding
312 Starlings, although there is no information on their diet in our study area, in general they rely
313 mostly on small insects such as hoverflies, ants, and on larvae of Coleoptera and Lepidoptera
314 (Motis et al. 1997), which should give Starlings a distinct isotope signature. However, in our
315 study area Starlings were often observed feeding their offspring with grasshoppers,
316 suggesting a similar diet and explaining the observed niche overlap with the remaining
317 community.

318 Isotopic characterisation of the adult community indicated a similar trophic rank, although
319 the information retrieved from adult plasma is probably not representative of the whole
320 breeding season, and results are based on few individuals in a single year. Adult Lesser
321 Kestrels and Rollers showed a high overlap with Starlings, suggesting similar diets, while
322 Common Kestrels and Little Wwls had more enriched $\delta^{15}\text{N}$ values, likely reflecting the
323 consumption of vertebrates.

324 *4.2 Parent-offspring dietary segregation*

325 Despite the lack of significant differences when comparing $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ values between
326 Little Owl adults and nestlings separately, differences in Euclidean distance suggests that all
327 species studied in these mixed colonies (including Little Owls) exhibited clear parent-
328 offspring dietary segregation during the chick-rearing period. During a season of high energy
329 demand, parents are expected to feed their offspring with high energy items, keeping less
330 profitable resources for themselves (Orłowski et al. 2014). While this was already reported
331 for Lesser Kestrels and Rollers (Catry et al. 2016a, 2019), we could not find studies
332 addressing parent-offspring segregation in the other studied species. This study reveals that

333 Little Owls, Common Kestrels and Spotless Starlings also exhibited marked parent-offspring
334 segregation in trophic resources, reinforcing this may be a common evolutionary strategy to
335 avoid intraspecific competition.

336 *4.3 Niche theory applied to mixed-species colonies in an artificial nesting landscape*

337 If on one hand the competitive pressure among species breeding together in dense
338 aggregations can be alleviated through parent-offspring partitioning in their food resources,
339 on the other hand the low dietary segregation among species can lead to an increase of
340 interspecific competitive interactions (Catry & Catry 2019).

341 These results raise questions as to the mechanisms allowing species coexistence. According
342 to the ecological niche theory, species with high dietary overlap may still partition their niche
343 and avoid interspecific competition through spatial or temporal segregation in habitat
344 utilisation (Pianka 1981; Reisinger et al. 2020). In the mixed-species colonies we studied,
345 there is no evidence for temporal or spatial segregation, as all species highly coincide in their
346 breeding phenology and chick rearing period, and all hunt in the vicinity of the colonies as
347 central-place foragers (Casagrande et al. 2008; Catry et al. 2012; Šalek & Lövy 2012; Catry
348 et al. 2017a). Different species could still select for different (micro) habitats or hunting
349 strategies that would foster niche segregation, an aspect that should be investigated in future
350 studies.

351 Alternatively, high niche overlap among sympatric species may reflect high availability of
352 resources (Pianka 1981; Charter et al. 2018; Jenkins & Davoren 2020). In the Castro Verde
353 SPA, traditional farming practices associated with agro-steppes are maintained through
354 specific agri-environmental policy schemes, ensuring high-quality foraging habitats for many
355 farmland species (Catry et al. 2013, 2017a; Silva et al. 2018). Large areas of low-intensity
356 grazed fallows support a high abundance of Orthoptera, and the chick-rearing period of

357 Lesser Kestrels and Rollers (and likely of the whole species assemblage except Barn Owls)
358 coincides with the peak abundance of grasshoppers (Catry et al. 2016a, 2017b, 2019). From
359 mid-May, the abrupt increase in grasshopper availability (Catry et al. 2017b), may allow
360 species to contract their dietary niche to focus on prey that became temporarily super-
361 abundant, increasing their overlap without necessarily increasing interspecific competition
362 (Forero et al. 2004; Jenkins & Davoren 2020). The lack of differences in productivity
363 between Rollers breeding solitary and in mixed-species colonies, or between Lesser Kestrels
364 in small or large colonies (Table S3), suggests that prey depletion around colonies in our
365 study area is unlikely and that food availability is high enough to allow species with similar
366 niches to coexist.

367 ***4.4 Conclusion***

368 Conservation actions based on nest provisioning are often successful at quickly recover
369 endangered populations, but seldom consider how they may shape the composition of
370 communities and alter interspecific interactions (Catry & Catry 2019). In the sequence of the
371 implementation of European Union LIFE European projects focused on the recovery of
372 Lesser Kestrels in Portugal, the provisioning of more than a thousand artificial nest-sites
373 attracted many other avian species. Interference competition and even predation have already
374 been reported among most species breeding in these colonies (Catry & Catry 2019), and the
375 high trophic overlap reported in this study may be another trigger responsible for the
376 observed interspecific antagonistic behaviours. This may particularly affect Lesser Kestrels
377 and Rollers, two species of conservation concern, classified as Vulnerable and Critically
378 endangered in Portugal, respectively, and with decreasing population trends throughout their
379 breeding ranges (Cabral et al. 2005; BirdLife International 2021). Food resources are
380 abundant in the study area, so interspecific competition for food may not be a problem here.
381 It remains unknown whether the high overlap reported here would still occur in areas with

382 limited resource availability or if species would decrease their trophic niche overlap to avoid
383 competition (Pianka 1981). Future studies should address how sympatric species coexist
384 under these scenarios, which are predicted to increase with ongoing human and climate-
385 induced habitat changes (Tcharntke et al. 2005; Marcelino et al. 2020; Raven & Wagner
386 2021). Conservation practices based on nest-site provisioning should nonetheless consider the
387 ecological niche of both target and sympatric species, and ensure suitable foraging habitat
388 with high resource availability in order to secure the long-term viability of threatened species
389 in these new, human-shaped nesting landscapes.”

390

391 **Availability of data:** The data that support the findings of this study are available from the
392 corresponding author upon reasonable request.

393

394 **References**

395 Alonso H, Granadeiro JP, Paiva VH, Dias AS, Ramos JA, Catry P (2012) Parent – offspring dietary
396 segregation of Cory’s shearwaters breeding in contrasting environments. *Mar Biol* 159: 1197–
397 1207. <https://doi.org/10.1007/s00227-012-1900-2>

398 Antón-Tello M, Britto VO, Gil-Delgado JA, Rico E, Dies JI, Monrós JS, Vera P (2021) Unravelling
399 diet composition and niche segregation of colonial waterbirds in a Mediterranean wetland using
400 stable isotopes. *Ibis*: 1-15. <https://doi.org/10.1111/ibi.12928>

401 Birdlife International (2021) IUCN Red List of Birds. <http://www.birdlife.org>. Accessed 07 May 2021

402 Birt VL, Birt TP, Goulet D, Cairns DK, Montevecchi WA (1987) Ashmole’s halo: direct evidence for
403 prey depletion by a seabird. *Mar Ecol Prog Ser* 40: 205-208

404 Casagrande S, Nieder L, Di Minin E, La Fata I, Csermely D (2008) Habitat utilization and prey
405 selection of the kestrel *Falco tinnunculus* in relation to small mammal abundance. *Ital J Zool* 75
406 (4): 401-409. <https://doi.org/10.1080/11250000802085526>

407 Catry I, Alcazar R, Franco AMA Sutherland W (2009) Identifying the effectiveness and constraints of
408 conservation interventions: A case study of the endangered lesser kestrel. *Biol Conserv* 142(11):
409 2782–2791. <https://doi.org/10.1016/j.biocon.2009.07.011>

410 Catry I, Franco AMA, Rocha P, Alcazar R, Reis S, Cordeiro A, Ventim R, Teodósio J, Moreira F
411 (2013) Foraging Habitat Quality Constrains Effectiveness of Artificial Nest-Site Provisioning in
412 Reversing Population Declines in a Colonial Cavity Nester. *PLoS ONE*, 8(3): 1–10.
413 <https://doi.org/10.1371/journal.pone.0058320>

414 Catry I, Catry T, Alho M, Franco AMA, Moreira F (2016a) Sexual and parent-offspring dietary
415 segregation in a colonial raptor as revealed by stable isotopes. *J Zool* 299(1): 58–67.
416 <https://doi.org/10.1111/jzo.12324>

417 Catry I, Marcelino J, Franco AMA, Moreira F (2017a) Landscape determinants of European roller
418 foraging habitat: implications for the definition of agri-environmental measures for species
419 conservation. *Biodivers Conserv* 26: 553-566. <https://doi.org/10.1007/s10531-016-1241-4>

420 Catry I, Sampaio A, Silva MC, Moreira F, Franco AMA, Catry T (2019) Combining stable isotope
421 analysis and conventional techniques to improve knowledge of the diet of the European Roller
422 *Coracias garrulus*. *Ibis* 161: 272-285. <https://doi.org/10.1111/ibi.12625>

423 Catry T, Ramos J, Le Corre M, Kojadinovic J, Bustamante P (2008) The role of stable isotopes and
424 mercury concentrations to describe seabird foraging ecology in tropical environments. *Mar Biol*
425 155 (6): 637-647. <https://doi.org/10.1007/s00227-008-1060-6>

426 Catry T, Catry I (2019) Nest-site provisioning re-shapes species interactions within bird assemblages.
427 *Ibis* 161: 699–704. <https://doi.org/10.1111/ibi.12731>

428 Catry T, Lourenço PM, Lopes RJ, Carneiro C, Alves JA, Costas J, Rguibi-Idrissi H, Bearhop S,
429 Piersma T, Granadeiro JP (2016b) Structure and functioning of intertidal food webs along an avian
430 flyway: a comparative approach using stable isotopes. *Funct Ecol* 30(3): 468-478.
431 <https://doi.org/10.1111/1365-2435.12506>

432 Catry T, Moreira F, Alcazar R, Rocha PA, Catry I (2017b) Mechanisms and fitness consequences of
433 laying decisions in a migratory raptor. *Behav Ecol* 28 (1): 222-232.
434 <https://doi.org/10.1093/beheco/arw150>

435 Caut S, Angulo E, Courchamp F (2009) Variation in discrimination factors ($\Delta^{15}\text{N}$ and $\Delta^{13}\text{C}$): the
436 effect of diet isotopic values and applications for diet reconstruction. *J Appl Ecol* 46: 443-453.
437 <https://doi.org/10.1111/j.1365-2664.2009.01620.x>

438 Charter M, Izhaki I, Roulin A (2018) The relationship between intra-guild diet overlap and breeding
439 in owls in Israel. *Popul Ecol* 60(4): 397-403. <https://doi.org/10.1007/s10144-018-0633-6>

440 Clewlow HL, Takahashi A, Watanabe S, Votier SC, Downie R, Ratcliffe N (2019) Niche partitioning
441 of sympatric penguins by leapfrog foraging appears to be resilient to climate change. *J Anim Ecol*
442 88 (2): 223-235. <https://doi.org/10.1111/1365-2656.12919>

443 Dehnhard N, Achurch H, Clarke J, Michel LN, Southwell C, Sumner MD, Eens M, Emmerson L
444 (2020) High inter- and intraspecific niche overlap among three sympatrically breeding, closely
445 related seabird species: Generalist foraging as an adaptation to a highly variable environment? *J*
446 *Anim Ecol* 89(1): 104-119. <https://doi.org/10.1111/1365-2656.13078>

447 Duckworth RA, Hallinger KK, Hall N, Potticary AL (2017) Switch to a novel breeding resource
448 influences coexistence of two passerine birds. *Front Ecol Evol* 5 (72): 1-11.
449 <https://doi.org/10.3389/fevo.2017.00072>

450 Forero MG, Bortolotti GR, Hobson KA, Donazar JA, Bertelotti M (2004) High trophic overlap within
451 the seabird community of Argentinean Patagonia: a multiscale approach. *J Anim Ecol* 78: 789-
452 801. <https://doi.org/10.1111/j.0021-8790.2004.00852.x>

453 Gameiro J, Silva JP, Franco AMA, Palmeirim JM (2020) Effectiveness of the European Natura 2000
454 network at protecting Western Europe's agro-steppes. *Biol Conserv* 248: 1-9.
455 <https://doi.org/10.1016/j.biocon.2020.108681>

456 Hardin G (1960) The competitive exclusion principle. *Science* 131 (3409): 1292-1297

457 Harper MJ, McCarthy MA, van der Ree R (2005) The use of nest boxes in urban natural vegetation
458 remnants by vertebrate fauna. *Wildl Res* 32: 509-516. <https://doi.org/10.1071/WR0410>

459 Högstedt G (1980) Resource partitioning in Magpie *Pica pica* and jackdaw *Corvus monedula* during
460 the breeding season. *Ornis Scand* 11(2): 110–115

461 Inger R, Bearhop S (2008) Applications of stable isotope analyses to avian ecology. *Ibis* 150(3): 447–
462 461. <https://doi.org/10.1111/j.1474-919X.2008.00839.x>

463 Jackson A, Parnell A (2020) SIBER: Stable Isotope Bayesian Ellipses in R. Retrieved from
464 <https://cran.r-project.org/web/packages/SIBER/index.html>

465 Jaksić FM, Seib RL, Herrera CM (1982) Predation by the Barn owl (*Tyto alba*) in Mediterranean
466 habitats of Chile, Spain and California; a comparative approach. *Am Midl Nat* 107: 152–159.
467 <https://doi.org/10.1007/BF00545658>

468 Jenkins EJ, Davoren GK (2020) Seabird species- and assemblage-level isotopic niche shifts associated
469 with changing prey availability during breeding in coastal Newfoundland. *Ibis*.
470 <https://doi.org/10.1111/ibi.12873>

471 Kappeler PM, Cremer S, Nunn CL (2015) Sociality and health: Impacts of sociality on disease
472 susceptibility and transmission in animal and human societies. *Philos Trans R Soc B* 370 (1669).
473 <https://doi.org/10.1098/rstb.2014.0116>

474 Logan JM, Jardine TD, Miller TJ, Bunn SE, Cunjak RA, Lutcavage ME (2008) Lipid corrections in
475 carbon and nitrogen stable isotope analyses: Comparison of chemical extraction and modelling
476 methods. *J Anim Ecol* 77(4): 838–846. <https://doi.org/10.1111/j.1365-2656.2008.01394.x>

477 MacArthur RH, Pianka ER (1966) On optimal use of a patchy environment. *Am Nat* 100 (916): 603-
478 609

479 Mainwaring MC (2015) The use of man-made structures as nesting sites by birds: A review of the
480 costs and benefits. *J Nat Conserv* 25: 17-22. <https://doi.org/10.1016/j.jnc.2015.02.007>

481 Marcelino J, Silva JP, Gameiro J, Silva A, Rego FC, Moreira F, Catry I (2020) Extreme events are
482 more likely to affect the breeding success of lesser kestrels than average climate change. *Sci Rep*
483 10(1): 1–11. <https://doi.org/10.1038/s41598-020-64087-0>

484 Moreira, F, Leitão PJ, Morgado R, Alcazar R, Cardoso A, Carrapato C, Delgado A, Gerales P,
485 Gordinho L, Henriques I, Lecoq M, Leitão D, Marques AT, Pedroso R, Prego I, Reino L, Rocha P,
486 Tomé R, Osborne PE (2007) Spatial distribution patterns, habitat correlates and population
487 estimates of steppe birds in Castro Verde. *Airo* 17, 5-30

488 Motis A, Estrada J, Oro D (1997) Nestling diet of the Spotless starling *Sturnus unicolor* and the
489 European Starling *Sturnus vulgaris* in a sympatric breeding area. *Ornis Fenn* 74(4): 179–185

490 Murton RK, Westwood NJ (1966) The foods of the rock dove and feral pigeon. *Bird Study* 13(2):
491 130–146. <https://doi.org/10.1080/00063656609476116>

492 Olah G, Vigo G, Heisohn R, Brightsmith DK (2014) Nest site selection and efficacy of artificial nests
493 for breeding success of Scarlet macaws *Ara macao macao* in lowland Peru. *J Nat Conserv* 22: 176-
494 185. <https://doi.org/10.1016/j.jnc.2013.11.003>

495 Orihuela-Torres A, Perales P, Rosado D, Pérez-García JM (2017). Feeding ecology of the Common
496 Kestrel *Falco tinnunculus* in the south of Alicante (SE Spain). *Rev Cat Ornitol* 33: 10–16.

497 Orłowski G, Karg J, Rłowski GO, Usiecki SR, Arg JK (2014) Partial Dietary Segregation between
498 Adult and Nestling Bluethroats *Luscinia svecica*. *Acta Ornithol.*49(1),:107–118.
499 <https://doi.org/10.3161/000164514X682931>

500 Pianka ER (1981) Competition and niche theory. In RM May (Ed.) *Theoretical ecology* (pp. 167–
501 196). Blackwell, Oxford.

502 R Core Team (2016) A Language and environment for statistical computing. Retrieved from Vienna:
503 R Foundation for statistical computing website: <https://cran.rproject.org/>

504 Razgour O, Korine C, Saltz D (2011) Does interspecific competition drive patterns of habitat use in
505 desert bat communities? *Oecologia* 167: 493-502. <https://doi.org/10.1007/s00442-011-1995-z>

506 Raven PH, Wagner DL (2021) Agricultural intensification and climate change are rapidly decreasing
507 insect biodiversity. *PNAS* 118(2): 1-6. <https://doi.org/10.1073/pnas.2002548117>

508 Reisinger RR, Carpenter-Kling T, Connan M, Cherel Y, Pistorius PA (2020) Foraging behaviour and
509 habitat-use drives niche segregation in sibling seabird species: Niche segregation in giant petrels.
510 *R Soc Open Sci* 7(9). <https://doi.org/10.1098/rsos.200649>

511 Riegert J, Šindelář, J, Zárbybnická M, Horáček I (2021). Large-scale spatial patterns of small-mamma
512 communities in the Mediterranean region revealed by Barn owl diet. *Sci Rep* 11: 4985.
513 <https://doi.org/10.1038/s41598-021-84683-y>.

514 Šalek M, Lövy M (2012) Spatial ecology and habitat selection of little owl *Athene noctua* during the
515 breeding season in central European farmland. *Bird Conserv Int* 22: 328-338.
516 <https://doi.org/10.1017/S0959270911000268>

517 Schuette P, Wagner AP, Wagner ME, Creel S (2013) Occupancy patterns and niche partitioning
518 within a diverse carnivore community exposed to anthropogenic pressures. *Biol Conserv* 158: 301-
519 312. <https://doi.org/10.1016/j.biocon.2012.08.008>

520 Silva JP, Correia R, Alonso H, Martins RC, D'Amico M, Delgado A, Sampaio H, Godinho C,
521 Moreira F (2018) EU protected area network did not prevent a country wide population decline in
522 a threatened grassland bird. *Peer J* 6: 1-13. <https://doi.org/10.7717/peerj.4284>

523 Soler M, Alcata N, Soler JJ (1990) Alimentación de la Grahilla *Corvus monedula* en tres zonas del sur
524 de España. *Doñana Acta Vertebrata* 17, 17-48

525 Sutherland WJ, Dicks LV, Ockendon N, Petrovan SO, Smith RK (2018) What works in conservation.
526 Open Book Publishers, Cambridge. <https://doi.org/10.11647/OBP.0131>

527 Teckentrup L, Kramer-Schadt S, Jeltsch F (2019) The risk of ignoring fear: underestimating the
528 effects of habitat loss and fragmentation on biodiversity. *Landsc Ecol* 34: 2851-2868.
529 <https://doi.org/10.1007/s10980-019-00922-8>

530 Tomé R, Catry P, Bloise C, Korpimäki E (2008) Breeding density and success, and diet composition of
531 little owls *Athene noctua* in steppe-like habitats in Portugal. *Ornis Fenn.*85: 22-32.

532 Tschardt T, Klein AM, Kruess A, Steffan-Dewenter I, Thies C (2005) Landscape perspectives on
533 agricultural intensification and biodiversity – ecosystem service management. *Ecol Lett* 8: 857-
534 874. <https://doi.org/10.1111/j.1461-0248.2005.00782.x>

535 Turner TF, Collyer ML, Krabbenhoft TJ (2010) A general hypothesis-testing framework for stable
536 isotope ratios in ecological studies. *Ecology* 91: 2227-2233. <https://doi.org/10.1890/09-1454.1>

537 Wiegand T, Revilla E, Moloney KA (2005) Effects of habitat loss and fragmentation on population
538 dynamics. *Conserv Biol* 19 (1): 108-121. <https://doi.org/10.1016/j.actao.2017.06.006>

539 Weiss F, Furness RW, McGill RAR, Strange IJ, Masello JF, Quillfeldt P (2009). Trophic segregation
540 of Falkland Islands seabirds: Insights from stable isotope analysis. *Polar Biol* 32(12): 1753–1763.
541 <https://doi.org/10.1007/s00300-009-0674-6>

542

543 **Tables**

544 **Table 1:** Carbon ($\delta^{13}\text{C}$) and nitrogen ($\delta^{15}\text{N}$) stable isotope signatures (mean \pm SD) of six bird
545 species breeding in mixed colonies in Castro Verde, Southern Portugal. Nestling feathers and
546 blood (plasma) from parent-offspring pairs were collected during the breeding seasons of
547 2014-2018 (pooled) and 2019, respectively. n = sample size.

| Feathers (nestlings) | | | Blood | | |
|-----------------------|-----------------------|-----|-----------------------|-----------------------|-----|
| $\delta^{13}\text{C}$ | $\delta^{15}\text{N}$ | n | $\delta^{13}\text{C}$ | $\delta^{15}\text{N}$ | n |

| | | | | | | | |
|----------------|-------------------|------------------|----|----------|-------------------|------------------|----|
| Barn owl | -23.42 ± 0.58 | 10.64 ± 0.45 | 12 | - | - | - | - |
| Little owl | -23.81 ± 1.07 | 10.08 ± 1.37 | 19 | Nestling | -25.21 ± 0.45 | 10.03 ± 1.11 | 7 |
| | | | | Adult | -25.63 ± 0.34 | 11.15 ± 0.74 | 7 |
| Common kestrel | -23.64 ± 0.55 | 8.88 ± 1.05 | 33 | Nestling | -25.28 ± 0.34 | 8.77 ± 0.63 | 8 |
| | | | | Adult | -25.20 ± 0.44 | 10.07 ± 0.47 | 8 |
| Lesser kestrel | -23.78 ± 0.25 | 7.76 ± 0.66 | 36 | Nestling | -25.55 ± 0.19 | 7.93 ± 0.76 | 11 |
| | | | | Adult | -25.38 ± 0.40 | 8.94 ± 0.91 | 11 |
| Roller | -23.52 ± 0.42 | 7.34 ± 0.69 | 36 | Nestling | -25.42 ± 0.26 | 7.30 ± 0.67 | 6 |
| | | | | Adult | -25.17 ± 0.23 | 8.78 ± 1.14 | 6 |
| Starling | -23.84 ± 0.52 | 8.54 ± 0.89 | 12 | Nestling | -25.48 ± 0.26 | 6.62 ± 1.03 | 6 |
| | | | | Adult | -25.49 ± 0.20 | 8.61 ± 1.02 | 6 |

549 **Table 2:** Carbon ($\delta^{13}\text{C}$) and nitrogen ($\delta^{15}\text{N}$) stable isotope signatures (mean \pm SD) of main
 550 prey sampled in Castro Verde, Southern Portugal. Taxa within coleoptera and orthoptera
 551 were grouped into SIA groups according to their similarity in isotopic signatures. n = sample
 552 size.

| | $\delta^{13}\text{C}$ | $\delta^{15}\text{N}$ | n | SIA group |
|----------------------------|-----------------------|-----------------------|----|-----------|
| Small mammals | | | | |
| <i>Apodemus sp.</i> | -25.6 \pm 0.8 | 7.0 \pm 1.9 | 7 | |
| <i>Mus sp.</i> | -25.3 \pm 0.4 | 8.1 \pm 0.9 | 3 | |
| Soricidae | -24.9 \pm 0.3 | 8.2 \pm 1.6 | 2 | |
| Birds | | | | |
| <i>Coturnix sp.</i> | -26.1 \pm 0.3 | 7.2 \pm 0.5 | 2 | |
| <i>Sturnus unicolor</i> | -25.3 \pm 0.6 | 8.5 \pm 1.2 | 3 | |
| Reptiles | | | | |
| <i>Chalcides striatus</i> | -25.3 \pm 0.5 | 8.2 \pm 0.8 | 8 | |
| Chilopoda | | | | |
| <i>Scolopendra sp.</i> | -25.9 \pm 0.8 | 9.0 \pm 0.7 | 8 | |
| Coleoptera | | | | |
| Carabidae und. | -25.5 \pm 0.5 | 10.9 \pm 3.0 | 3 | 1 |
| <i>Silpha sp.</i> | -25.5 \pm 0.7 | 10.8 \pm 3.4 | 3 | 1 |
| Tenebrionide und. | -26.3 \pm 0.3 | 9.7 \pm 0.2 | 2 | 1 |
| <i>Amphimillon nigrum</i> | -26.6 \pm 0.8 | 6.6 \pm 0.9 | 10 | 2 |
| <i>Carabus rugosus</i> | -26.7 \pm 0.8 | 5.9 \pm 1.0 | 2 | 2 |
| <i>Chrysolina bankii</i> | -27.1 \pm 0.2 | 7.0 \pm 1.0 | 3 | 2 |
| <i>Pterostichus ebenus</i> | -25.9 \pm 0.1 | 7.0 \pm 0.2 | 2 | 2 |
| <i>Sepidium sp.</i> | -26.8 \pm 0.8 | 6.5 \pm 0.7 | 8 | 2 |
| <i>Bubas bison</i> | -27.3 \pm 0.6 | 8.9 \pm 0.4 | 5 | 3 |
| <i>Netocia sp.</i> | -23.2 \pm 0.3 | 5.6 \pm 3.7 | 2 | 4 |
| Protaetia sp. | -23.9 | 7.4 | 1 | 4 |
| Orthoptera | | | | |
| Acrididae | -26.4 \pm 0.7 | 3.8 \pm 1.0 | 12 | 5 |
| Tettigonidae | -25.3 \pm 0.7 | 3.4 \pm 1.5 | 14 | 5 |
| <i>Gryllotalpa sp.</i> | -27.2 \pm 0.7 | 6.4 \pm 1.0 | 7 | 6 |

553

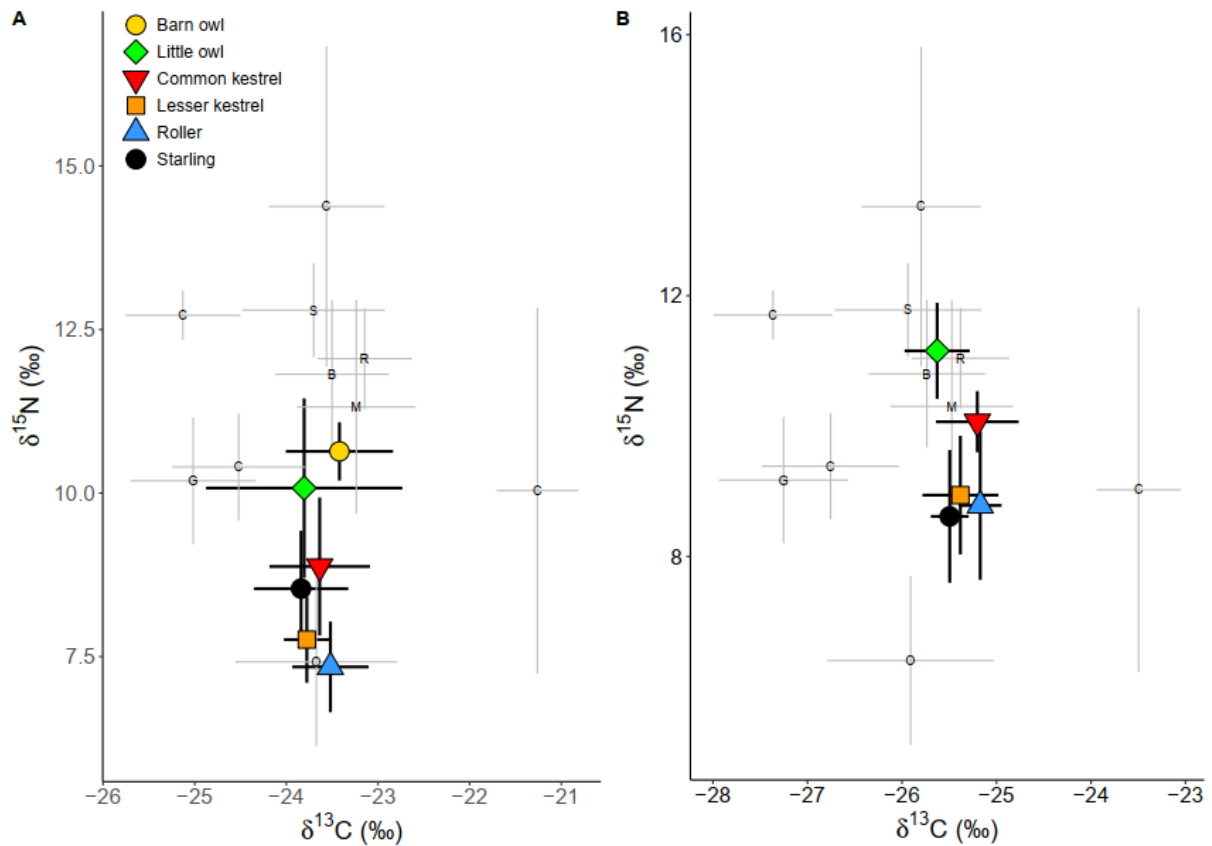
554

555 **Table 3:** *Isotopic niche breadth of nestlings (feathers) and adults (plasma) from bird species*
 556 *breeding in mixed colonies. Bayesian standard ellipse areas (SEAb, with 95% credible*
 557 *intervals) and sample size-corrected standard ellipse areas (SEAc) are shown. n = sample*
 558 *size.*

| Consumers | n | SEAb (95% credible interval) | SEAc |
|-----------------------------|----------|---|-------------|
| Nestlings (feathers) | | | |
| Barn Owl | 12 | 0.755 (0.439-1.451) | 0.899 |
| Little Owl | 19 | 4.352 (2.824-7.175) | 4.848 |
| Common Kestrel | 33 | 1.763 (1.265-2.543) | 1.872 |
| Lesser Kestrel | 36 | 0.505 (0.369-0.718) | 0.535 |
| Roller | 36 | 0.830 (0.602-1.173) | 0.871 |
| Starling | 12 | 1.320 (0.770-2.540) | 1.575 |
| Adults (plasma) | | | |
| Little owl | 7 | 0.688 (0.333-1.682) | 0.922 |
| Common kestrel | 8 | 0.558 (0.290-1.297) | 0.736 |
| Lesser kestrel | 11 | 1.058 (0.590-2.064) | 1.254 |
| Roller | 6 | 0.578 (0.274-1.592) | 0.785 |
| Starling | 6 | 0.532 (0.248-1.446) | 0.764 |

559

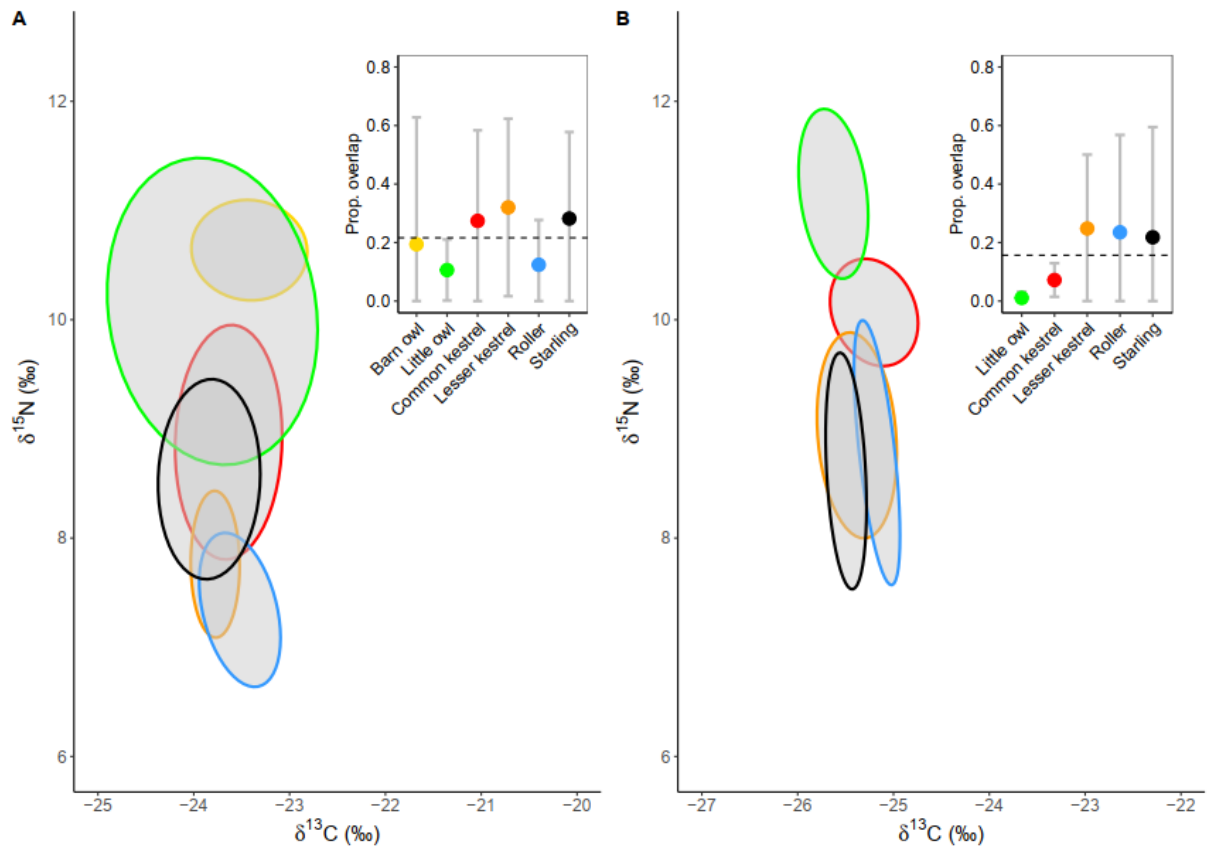
560 **Figure legends**



561

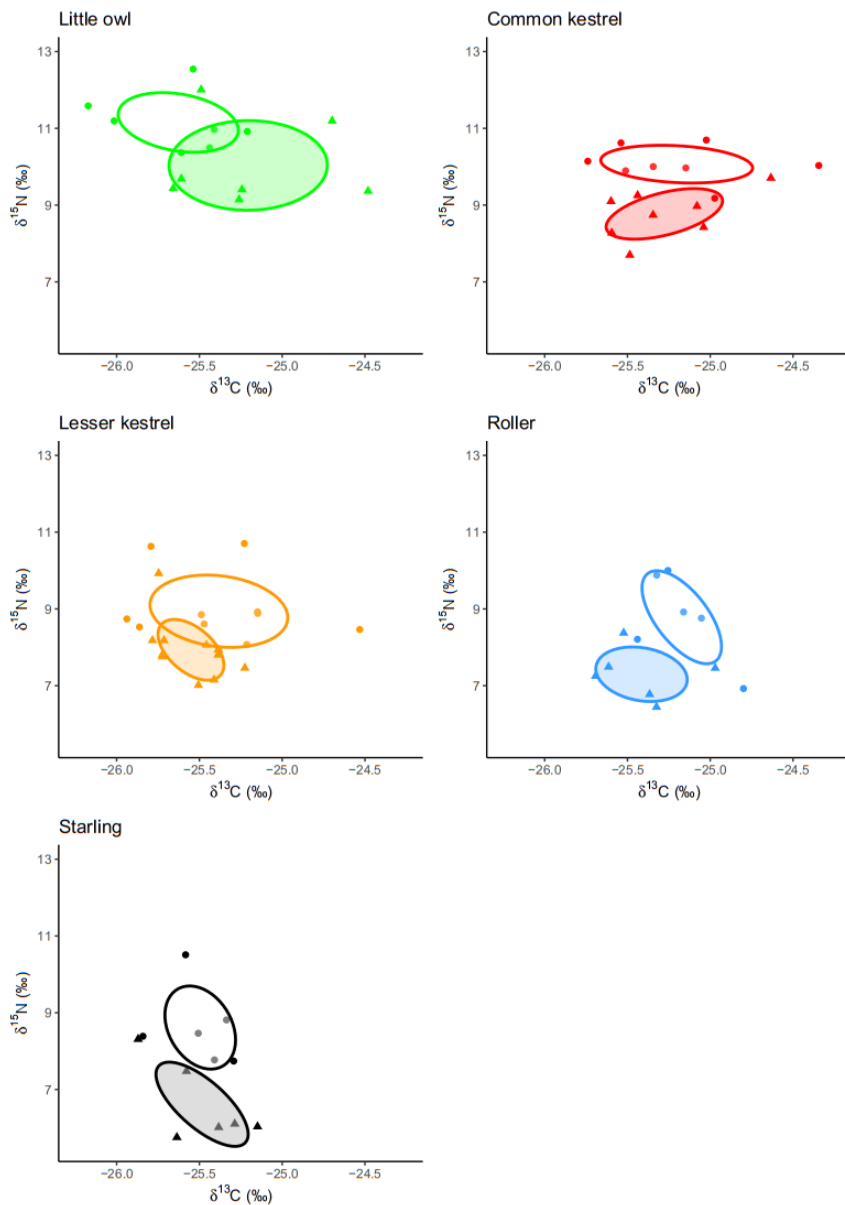
562 **Fig. 1:** Carbon ($\delta^{13}C$) and nitrogen ($\delta^{15}N$) stable isotope signatures (mean \pm SD) of nestlings
 563 (A) and adults (B) from six bird species breeding in mixed colonies established in an
 564 artificial nesting landscape in Southern Portugal. Isotope signatures of nestlings and adults
 565 were obtained from feathers (collected in 2014, 2016, and 2018, pooled) and blood plasma
 566 (collected in 2019), respectively. Signatures of main prey items are shown. M, small
 567 mammals; B, birds; R, reptiles; S, scolopendra; C, coleoptera; O, orthoptera; G,
 568 *Gryllotalpa*. Stable isotope ratios of prey items were obtained from muscle/soft tissue and
 569 corrected for the trophic discrimination factor (based on Caut et al. 2009, see methods).
 570 Because different tissues were used for nestlings and adults, direct comparisons in trophic
 571 position between the two age groups should be avoided.

572



573

574 **Fig. 2:** *Overlap in isotopic niches (corrected standard ellipses; SEAC) among nestlings (A)*
 575 *and adults (B) from 6 bird species breeding in mixed colonies established in an artificial*
 576 *breeding landscape in Southern Portugal. Isotope signatures of nestlings and adults were*
 577 *obtained from feathers and blood plasma, respectively. Inset plot (upper right corner) shows*
 578 *the mean (\pm SD) niche overlap for each species and the mean overlap of the community*
 579 *(horizontal dashed line).*



580

581 **Fig. 3** *Overlap in isotopic niches (corrected standard ellipses; SEAc) between parents (clear)*

582 *and their offspring (filled) of bird species breeding in mixed colonies established in an*

583 *artificial nesting landscape in Southern Portugal. Stable isotope ratios ($\delta^{13}\text{C}$ and $\delta^{15}\text{N}$) were*

584 *measured using blood plasma from 7 Little Owl, 8 Common Kestrel, 11 Lesser Kestrel, 6*

585 *Roller, and 6 Starling parent-offspring pairs (one parent and one offspring in each pair).*

586 *Circles and triangles show individual signatures of parents and nestlings, respectively.*