

# Smart fertilizers as a strategy for sustainable agriculture

Marcela Calabi-Floody<sup>a</sup>, Jorge Medina<sup>b</sup>, Cornelia Rumpel<sup>c</sup>, Leo M. Condrón<sup>d</sup>, Marcela Hernandez<sup>e</sup>, Marc Dumont<sup>e</sup>, Maria de la Luz Mora<sup>f</sup>

<sup>a</sup>Nano-biotechnology Laboratory, Center of Plant, Soil Interaction and Natural Resources Biotechnology, Scientific and Biotechnological Bioresource Nucleus, BIOREN-UFRO, Universidad de La Frontera, Temuco, Chile.

<sup>b</sup>Departamento de Ciencias Químicas y Recursos Naturales, Scientific and Technological Bioresources Nucleus, BIOREN-UFRO, Universidad de La Frontera, Temuco, Chile.

<sup>c</sup>CNRS, Institute for Ecology and Environmental Sciences IEES (UMR 7618, CNRS-UPMC-UPEC-IRD-INRA), Thiverval-Grignon, France.

<sup>d</sup>Faculty of Agriculture and Life Sciences, PO Box 85084, Lincoln University, Lincoln 7647, Christchurch, New Zealand.

<sup>e</sup>Centre for Biological Sciences, University of Southampton, Southampton, SO17 1BJ, United Kingdom.

<sup>f</sup>Center of Plant, Soil Interaction and Natural Resources Biotechnology, Scientific and Biotechnological Bioresource Nucleus, BIOREN-UFRO, Avenida Francisco Salazar 01145, Universidad de La Frontera, Temuco, Chile.

\*Corresponding author: Marcela Calabi Floody. Telephone: +56 (45) 2596856; fax: +56 (45) 2325053; e-mail: [marcela.calabi@ufrontera.cl](mailto:marcela.calabi@ufrontera.cl)

24 **Abstract**

25

26 In the coming decades there will be increasing pressure on global food systems, and  
27 agriculture will have the challenge to provide food security for a growing world population  
28 without impacting environmental security. Accordingly, it will be necessary to use modern  
29 technologies in agroecosystems in order to supply sufficient food and decrease the negative  
30 impacts on the environment induced by chemical fertilization and by inadequate disposal or  
31 reuse of agricultural wastes. A combination of biotechnology and nanotechnology has the  
32 potential to revolutionize agricultural systems and provide solutions for current and future  
33 problems. These include the development and use of smart fertilizers with controlled  
34 nutrient release, together with bio-formulations based on bacteria or enzymes. This study  
35 was designed to provide a critical review of information related to current food security  
36 issues and the role of smart fertilizer development in future food production. We  
37 concentrate on advances in the development of controlled release biofertilizers and the use  
38 of harvesting residues as coating and carrier materials.

39

40 **Keywords:** biofertilizer; bio-formulations; carrier materials; harvesting residues; food  
41 security; nanofertilizers; slow/controlled release; wheat straw.

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## 81 **1. Introduction**

82 Agricultural land systems (cropland, managed grassland, permanent crops including  
83 agro-forestry and bio-energy crops) cover about 40–50% of the Earth’s land surface (Smith  
84 et al. 2007), on which humanity needs to secure food production. The global population is  
85 expected to increase from 7.2 billion to 9.6 billion by 2050 (UN, 2013), which will increase  
86 food demand and fodder requirements for feedstock. In 2015, the UN adopted 17  
87 sustainable development goals, aiming to eradicate hunger and extreme poverty by 2030,  
88 while at the same time preserving the environment and global climate. This implies  
89 sustainable intensification on existing agricultural land through innovation and  
90 collaboration between multiple sectors (Chabbi et al., 2017). One option to achieve greater  
91 crop production could be the improvement of plant fertilization strategies.

92 Nitrogen (N) and phosphorus (P) are essential nutrients for plant growth and  
93 consequently the application of these nutrients as chemical fertilizers has been growing  
94 since the green revolution in the 1960s and determines crop productivity (Haygarth et al.  
95 2013; Mora et al., 2007; Stewart et al., 2005). Continued fertilizer inputs are essential to  
96 sustain and increase food production. However, there are problems associated with mineral  
97 fertilizer use because of relatively low nutrient uptake by crops in productive systems  
98 (Trenkel, 1997). High fertilization rates lead to N and P losses with negative impacts on  
99 atmospheric greenhouse gas (GHG) concentrations and water quality (Haygarth et al.  
100 2013). There is an urgent need to improve nutrient use efficiency in agricultural systems,  
101 and to manage biogeochemical cycles in a sustainable way (Rumpel et al., 2015a). This  
102 includes the development and application of modern biotechnological tools, such as plant  
103 growth promoting rhizobacteria (PGPR) and diazotrophic N<sub>2</sub>-fixing bacteria as alternatives

104 to conventional fertilization.

105           Many different agricultural commodities are produced worldwide and will be  
106 needed to assure a diverse, healthy nutritious diet. However, global food security will  
107 continue to depend most heavily on staple foods, which are the three cereal crops rice  
108 (*Oryza spp*), wheat (*Triticum spp*) and maize (*Zea mays spp*). Global cereal production has  
109 to increase with the objective to satisfy the growing global demand (He et al., 2014).  
110 Higher production will in turn increase the amounts of harvesting residues (e.g. straw,  
111 stubble) that can be used as biomass feedstock or for animal feeding (Jiang et al., 2012;  
112 Habets et al., 2013). Unfortunately, one of the most important practices globally is the  
113 removal of these residues by *in-situ* burning with significant environmental, economic and  
114 human health impacts (Gupta et al., 2016; Singh et al., 2010). Harvesting residues should  
115 be considered as a resource that can be utilized as organic raw material, which could be  
116 used to improve soil quality and productivity. One way to take advantage of these residues  
117 is their use as composting agents (Medina et al., 2017; Roca-Pérez et al., 2009). Pyrolysis  
118 of harvesting residues to produce biochar has also been suggested as a win-win practice  
119 leading not only to energy generation but also to a soil conditioner with the potential to  
120 increase plant growth and soil carbon (C) sequestration (Abiven et al., 2014). Another  
121 strategy may be the transformation of harvesting residues into raw materials for fertilizer  
122 production.

123           The aim of this review is to present innovations related to smart fertilizer  
124 technology as a response to food security scenarios under growing global population and  
125 the environmental impacts of current agricultural systems. Smart fertilizers may be a  
126 solution to enhance food production and environmental quality. In the sense of a circular  
127 economy, we suggest that these smart fertilizers may be based on the innovative use of

128 harvesting residues.

129

## 130 **2. Food security: Agricultural management and its environmental** 131 **footprints**

### 132 **2.1 World food demand associated with a growing global population**

133 Three cereals (rice, wheat and maize) account for 58% of the annual crop area and  
134 provide about 50% of food calories (Fischer et al., 2009). Rice and wheat are essential  
135 suppliers of energy for the population of developing countries, and maize makes up over  
136 60% of commercial animal feeds (Fischer et al., 2009). Based on projected global  
137 population growth, an annual increase of the world cereal production of 0.9% reaching  
138 3,009 billion tones (3,009,000 Tg) will be required to meet the demand (He et al., 2014).  
139 Average global cereal yields will need to increase from 3.32 t/ha to 4.30 t/ha, with wheat  
140 yields rising from 2.8 t/ha to 3.8 t/ha (He et al., 2014). Agricultural land needs either to be  
141 expanded or used more efficiently in order to increase the global food production by 60-  
142 110% to meet the rising demand of the global population in 2050 (Ray et al., 2013).  
143 Expansion of agricultural land may not be a good option, as non-agricultural land area is  
144 needed for other purposes, for example to provide habitat for endangered species (Balmford  
145 et al., 2005). Thus sustainable intensification leading to production increase on existing  
146 land area seems to be the best option (Godfray et al., 2010).

147 Cereal yields were increased during the green revolution due to genetic  
148 improvement and intensive utilization of mineral fertilizers (Pingali, 2012). However,  
149 yields have not increased further since the 1990s (Brisson et al., 2010; Grassini et al., 2013;  
150 Ray et al., 2012). Amongst the factors responsible for yield stagnation is climate change,

151 principally increases of temperature and prolongation of summer droughts. Depletion of  
152 soil reserves due to soil organic matter (OM) losses and the prolonged and intensive use of  
153 agrochemicals may also have contributed (Baishya, 2015; Lal, 2004; Lipper et al., 2014).

154 It is therefore highly uncertain if current crop improvement and management  
155 practices will be able to achieve food security in the future. Moreover, the future demand  
156 for agricultural products will be further affected by other factors such as the decline in rural  
157 workforce and the requirements of the biofuel market and climate change, which could  
158 have a huge impact on food productivity (FAO, 2009). Moreover, the negative  
159 environmental impacts of the green revolution due to massive fertilizer use call for the  
160 adaptation of more sustainable technologies (Pingali, 2012).

161

## 162 **2.2 Availability of chemical fertilizer to support food production**

163 The food production is strongly dependent on N and P, which are essential and  
164 irreplaceable nutrients for plant growth and to maintain life in the world. For example, long  
165 term data obtained between 1960-2010 estimated that around 48% of crop N was  
166 contributed by inorganic fertilizers, considering maize, rice, and wheat production systems  
167 (Ladha et al., 2016). Nevertheless, N and P present a significant difference in terms of their  
168 availability. In this sense, the supply of N is currently unlimited due to the production of  
169 urea by the Haber Bosch process, (Dawson and Hilton, 2011), which industrially produces  
170 around 100 TgNyr<sup>-1</sup> (Ladha et al., 2016). On the other hand, phosphate rock reserves are  
171 finite and there is a critical concern about the availability and cost of phosphate rock in the  
172 future (Cordell et al., 2009; Dawson and Hilton, 2011; IFDC, 2010; USGS, 2016). In this  
173 sense, Elser and Bennett (2011) stated: “More important than the amount of P in the ground  
174 is how much it will cost to get it out”. Figure 1 shows the principal reserves of rock

175 phosphate, most of which are located in Morocco.

176         Despite the fact that P is a finite resource, continuous inputs are needed to maintain  
177 the productivity of agroecosystems under current scenarios (Ostertag et al., 2016; Valkama  
178 et al., 2016). A recent meta- analysis carried out by Valkama et al. (2016) showed that yield  
179 response to P fertilization varied considerably in grassland systems and initial soil tests for  
180 P do not always predict this behavior. They described that the major sources of variation in  
181 yield responses to annually applied P were soil type specific. Under this context, only about  
182 40% of the variation in yield could be attributed to fertilizer P applications (Valkama et al.,  
183 2016). In many situations P-fertilizer inputs, especially in tropical areas are rapidly fixed by  
184 the soil matrix and not available for plant uptake. Therefore, the emerging global challenge  
185 of the issues associated with P supply is to improve the overall P-use efficiency of plants  
186 (Cordell and White, 2011).

187

### 188 **2.3 Conventional fertilization practices: their environmental consequences**

189         The type of soils and its management has a strong influence on the conventional  
190 fertilizer use efficiency. For example, it has been reported that Cambisols, which are one of  
191 the major soil types under agricultural use with an estimated surface of about 1.5 billion ha  
192 under cropland (FAO, 2001), are deficient in nutrients such as P (Dabin, 1980).

193         Moreover, in some Cambisols from central Africa, the continuous application of N  
194 and P fertilizers in addition with the unbalanced and suboptimal fertilization (e.g., by the  
195 exclusive application of N and P containing fertilizers such as urea and di-amonium  
196 phosphate) for long periods of time, has led to soil nutrient depletion, especially when the  
197 entire crop biomass is removed from land (Tesfay and Gebresamuel, 2016). Moreover,  
198 nowadays, often fertile soils are lost for agricultural production through urbanization.



199 Agriculture, as a result, is conferred to marginal land with low OM and nutrient content  
200 (Ngo et al., 2014). On such soils, the application of mineral fertilizers can cause accelerated  
201 acidification and further nutrient and OM losses (Lal, 2006; Marschner et al., 2002). Other  
202 soil types such as Andisols, which represent 0.84% (around 110 to 124 million ha) of  
203 global soil area (FAO, 2001), are characterized by high OM content (1–25% w/w) and high  
204 capacity to immobilize P. In this sense, their clay fraction makes up 35–60% of the soil,  
205 and is dominated by allophane (Besoain and Sepúlveda, 1985; Escudey et al., 2001).  
206 Agricultural management to obtain productive systems and the allophanic nature of these  
207 soils leads to acidification as a result of the use of urea and other ammonia (NH<sub>3</sub>) fertilizers  
208 (Mora et al., 2007, 2005). For example, about 50% of the Chilean Andisols are acidic with  
209 pH values between 4.5 and 5.5 (Jorquera et al., 2014; Mora et al., 2006, 2002, 1999). Soil  
210 acidity is one soil property contributing to P-fixation (Mora et al., 2004), decreasing its  
211 availability for plant nutrition. More than 50% of P incorporated in these soils is fixed as  
212 organic P (Borie and Rubio, 2003) and may contribute to the residual fraction (Velásquez et  
213 al., 2016). Thus, huge amounts of conventional P-fertilizer need to be applied annually to  
214 maintain available P levels in soil-plant systems.

215 N- and P-fertilizer application at levels exceeding plant requirements due to low acquisition  
216 efficiency leads to significant environmental consequences in many parts of the world due  
217 to N losses, such as: nitrate (NO<sub>3</sub><sup>-</sup>) and phosphate (PO<sub>3</sub><sup>-</sup>) leaching, NH<sub>3</sub> volatilization, and  
218 nitrous oxide (N<sub>2</sub>O) emission (Muñoz et al., 2010; Núñez et al., 2010; Saggar et al., 2013;  
219 Vistoso et al., 2012). Transport of P and N from agricultural soils to surface waters has  
220 been linked to eutrophication of freshwater and estuaries (Entry and Sojka, 2007; Liu et al.,  
221 2013; Riley et al., 2001; Smith and Schindler, 2009). These negative environmental  
222 consequences associated with fertilizer inputs further emphasize the need of technological

223 approaches to improve nutrient management in modern agriculture. In addition, current  
224 agricultural activities contribute up to 20% to the annual atmospheric emissions of GHG,  
225 such as methane (CH<sub>4</sub>) and carbon dioxide (CO<sub>2</sub>) (Lemke et al., 2007). Even higher  
226 contribution was noted for N<sub>2</sub>O (about 60%) (Smith et al., 2007), which is a potent GHG  
227 and catalyst for stratospheric ozone depletion (Yang et al., 2014), with more than 300 times  
228 the global warming potential than CO<sub>2</sub> (Crutzen, 1981; Kennedy et al., 2004). Its emission  
229 is closely related to mineral fertilizer input. In some countries, GHG emissions from  
230 conventional agriculture have increased by 33% between 1984 and 2003 (González et al.,  
231 2009).

232 One way to mitigate these GHG emissions could be the sequestration of C and N in soils in  
233 form of OM (Lal, 2003). Soils contain about three times more C than the above-ground  
234 vegetation, and approximately 75% of the terrestrial C pool and play a key role in the  
235 global C cycle (Calabi-Floody et al., 2015, 2011; Paustian et al., 2016; Le Quéré et al.,  
236 2015; Rumpel et al., 2015b; Schlesinger, 1986). The importance of the sink function of  
237 agricultural soils for GHG depends on the biophysical processes,  
238 incorporation/decomposition of organic residues, fertilizer application and environmental  
239 factors (Muñoz et al., 2010). In this sense, if mismanaged with intensive tillage and  
240 fertilization soil may lose OM and become a source of CO<sub>2</sub>. Agricultural wastes such as  
241 dairy slurry or manure (Muñoz et al., 2010; Rochette et al., 2008; Saggar et al., 2009), and  
242 biomass after harvesting (Garay et al., 2009; Taladriz and Schwember, 2012) may also be a  
243 source of GHG emissions. Improvement of organic residue recycling in agriculture may be  
244 a solution in view of sustainable intensification of agricultural practices and could  
245 contribute to increase soil C storage, thereby improving soil quality and to some extent  
246 mitigating atmospheric GHG concentrations (Chabbi et al., 2017).

247

## 248 **2.4 Management of crop waste after harvesting**

249           Worldwide, the annual production of agricultural residues amounts to 3.7 Pg dry  
250 matter (Bentsen et al., 2014). Straw, roots, shaft and other tissues of corn, wheat and rice as  
251 the main crop residues account for about 40.6%, 24.2% and 15.7% respectively (Medina et  
252 al., 2015). Burning of large amounts of crop residues produced worldwide in open fields  
253 leads to soil fertilization in form of ash input. Residue burning accounts for 27 % (1900 Tg  
254 dry matter) of global biomass burned annually (Crutzen et al., 2016). It is a source of  
255 atmospheric pollution with significant impacts on atmospheric chemistry, global climate  
256 and with great threat to human health (Brühl et al., 2015; Pongpiachan et al., 2015;  
257 Udeigwe et al., 2015). For example, burning wheat residues generates huge amounts of  
258 particulate material less than  $< 2.5 \mu\text{m}$  ( $\text{PM}_{2.5}$ ), GHG (i.e.,  $\text{CH}_4$ ,  $\text{N}_2\text{O}$ ,  $\text{CO}_2$ ), volatile  
259 organic carbon (VOC),  $\text{NH}_3$ , sulfur dioxide ( $\text{SO}_2$ ) and other pollutants (Crutzen et al., 2016;  
260 Sun et al., 2016; Zhang et al., 2015, Koppmann et al., 2005, Li et al., 2008).

261 China is one of the main crop residues producers with around 18% of the total global  
262 production (He et al., 2014; Kung et al., 2015; Zhang et al., 2016, 2015). Studies conducted  
263 by Sun et al. (2016) investigated  $\text{CO}_2$  emissions in China from 1996 to 2013, including the  
264 contribution from combustion of maize, wheat and rice residues. They calculated that these  
265 sources accounted for 22.5% of total emissions. In the USA, over 1.2 million ha of  
266 cropland is burnt annually. This has been estimated to produce 6.1 Tg of  $\text{CO}_2$ , 8.9 Gg of  
267  $\text{CH}_4$ , 232.4 Gg of CO, 10.6 Gg of  $\text{NO}_2$ , 4.4 Gg of  $\text{SO}_2$ , as well as 28.5 Gg of particulate  
268 matter of less than  $10 \mu\text{m}$  in diameter ( $\text{PM}_{10}$ ), and 20.9 Gg of  $\text{PM}_{2.5}$  (Udeigwe et al., 2015).  
269 Although open straw burning occurs primarily in rural areas, it also has a serious impact on  
270 urban air quality under the effect of air circulation (Chen et al., 2015).

271 Taladriz and Schwember (2012) reported that in Chile stubble burning is the most common  
272 management practice with 80 to 90% of wheat stubble being burned. It has been estimated  
273 that straw burning generates nutrient losses by volatilization of 98-100% for N, 20-40% for  
274 P and potassium, and 70-90% for sulfur (Heard et al., 2006), affecting their potential  
275 incorporation into the soil. The associated cost is US\$125 per ha, and considering post-burn  
276 runoff the cost is US\$225 per ha.

277 While the main effect of straw burning is on the atmospheric chemistry and air quality, this  
278 practice has additionally negative consequences on soils affecting the soil OM quantity and  
279 quality, the activity and colonization of topsoil by microorganisms, biological diversity and  
280 nutrient dynamics among other negative environmental implications (Borie et al., 2006,  
281 2010).

282

## 283 **2.5 Potential economic benefits to reusing harvesting residues**

284 Cereal straw, one of the most important harvesting residue globally, is  
285 biodegradable and a natural source of cellulose, hemicellulose and lignin, which in turn  
286 may be used for paper, biofuel and biogas production (Bhatnagar and Sillanpää, 2010;  
287 Hansen et al., 2014; Huang et al., 2007; Liu et al., 2013; Ma et al., 2011; Risberg et al.,  
288 2013; Talebnia et al., 2010; Xie et al., 2011), livestock bedding, animal feed (Dunford and  
289 Edwards, 2010) and for direct incorporation into the soil for nutrient recycling (Aulakh and  
290 Rennie, 1987; Garay et al., 2009; Li et al., 2011; Misselbrook et al., 2012; Taladriz and  
291 Schwember, 2012; Tan et al., 2007). The straw incorporated into the soil is an important  
292 energy and C source, which can positively influence the biological, chemical and physical  
293 properties, improving soil quality and productivity (Mulumba and Lal, 2008; Wei et al.,  
294 2015). The incorporation of crop straw residues may increase N mineralization and

295 available N, as well as organic C and total N stocks in soil (Cassman et al., 1996; Wei et al.,  
296 2015; Zhang et al., 2015). Moreover, increased availability of some nutrients such as P and  
297 K has been also reported for soils in semiarid areas (Wei et al., 2015). Nevertheless, it has  
298 been demonstrated that the excessive incorporation of fresh residues into soils may have  
299 adverse effects on the soil environment and crop yields (Jiang et al., 2012), because many  
300 soil functions require a mature and stable OM (Medina et al., 2015; Plaza and Senesi,  
301 2009). Some adverse effects of straw incorporation include (i) an increase of the  
302 mineralization rate of native soil organic C, (ii) the induction of anaerobic conditions by  
303 mineralization of large amounts of non-stabilized organic C, (iii) the associated extended  
304 O<sub>2</sub>-consumption, (iv) the alteration of soil pH (Senesi and Plaza, 2007), (v) the stimulation  
305 of NH<sub>3</sub> and GHG emissions and (vi) the biotic and abiotic immobilization of N (Luo et al.,  
306 1999; Said-Pullicino et al. 2014; Yan et al., 2005; Zhang et al. 2015). In this sense,  
307 considering its high polymerization degree due to cellulose and lignin content and high C/N  
308 ratio, low decomposition may occur when wheat straw is incorporated into soil (Li et al.,  
309 2011; Taladriz and Schwember, 2012) and lead to N deficiency due to high microbial  
310 demand (i.e. immobilization). Therefore, pretreatment through aerobic or anaerobic  
311 degradation is necessary to initiate wheat straw decomposition, improving its physiological,  
312 biochemical and morphological characteristics, increasing its nutrient content and reducing  
313 its C/N ratio (Pan and Sen, 2013).

314         Wheat straw is one of the best-known fibers and despite its low reutilization, its  
315 industrial potential is increasing. Recent studies suggested that structural modification of  
316 wheat straw by grinding can influence its degradation, porosity and surface area (Silva et  
317 al., 2012; Wang et al., 2016). Moreover, different chemical and biological treatments have  
318 been investigated to modify some properties of these materials for smart fertilizer design

319 (Panthapulakkal and Sain, 2015). Therefore, wheat straw is potentially valuable after a  
320 series of chemical modifications, such as esterification, etherification and copolymerization,  
321 and shows a broad potential for applications in superabsorbent material production (Liu et  
322 al., 2013; Ma et al., 2011; Xie et al., 2012). In recent years, researchers have focused on  
323 superabsorbent and slow-release fertilizers, as summarized below.

324

### 325 **3. New technologies to ensure food security and environmental health for** 326 **the expanding world population**

327 In order to enhance nutrient use efficiency, new types of smart fertilizers with  
328 controlled nutrient release are needed. The development of such fertilizers could be based  
329 on the use of microorganisms (biofertilizers) and/or nanomaterials (nanofertilizers).

330 In this context, nanotechnology is a promising, rapidly evolving field of  
331 interdisciplinary research that has potential to revolutionize food systems (Sastry et al.,  
332 2011). Nanotechnology involves the design, synthesis and use of materials at nanoscale  
333 level, ranging from 1 to 100 nm (EPA, 2007). At this scale, the physical, chemical and  
334 biological properties of materials differ fundamentally from the properties of individual  
335 atoms, molecules or bulk matter (Cao, 2004; Mansoori, 2005). The ability to manipulate  
336 matter at the nanoscale can lead to improved understanding of biological, physical and  
337 chemical processes and to the creation of improved materials, structures, devices and  
338 systems that can be used in agroecosystems (Sastry et al., 2011; Qian and Hinostroza,  
339 2004).

340 The application of nanotechnology to agriculture and food industries is growing  
341 rapidly as shown by increasing numbers of publications and patents (Fig. 2).

342 Nanotechnology has a number of potential benefits ranging from improved food quality and  
343 safety to reduced agricultural inputs and improved processing and nutrition (Veronica et al.,  
344 2015). The development of smart fertilizers based on nanotechnology is a recent  
345 phenomenon (Manjunatha et al., 2016), with an emphasis on controlled release and/or  
346 carrier/delivery systems to synchronize nutrient availability with the plant demands, thus  
347 reducing losses to the environment (Fig. 3) (Bley et al., 2017; Chinnamuthu and Boopathi,  
348 2009; DeRosa et al., 2010).

349

### 350 **3.1 Slow / controlled-release fertilizers**

351 According to Trenkel (1997) slow or controlled-release fertilizers are those  
352 containing a plant nutrient in a form, which either a) delays its availability for plant uptake  
353 and use after application, or b) is available to the plant significantly longer than a reference  
354 ‘rapidly available nutrient fertilizer’ such as ammonium  $\text{NO}_3^-$  or urea, ammonium  
355 phosphate or potassium chloride (AAPFCO, 1995). There is no official differentiation  
356 between slow-release and controlled-release fertilizers. However, the microbially-  
357 decomposed N products, such as urea-formaldehydes, are commonly referred to as slow-  
358 release fertilizers, and coated or encapsulated products as controlled-release fertilizers  
359 (Trenkel, 1997). Delayed availability of nutrients or consistent supply for extended time  
360 periods can be achieved through a number of mechanisms. These include semi-permeable  
361 coatings for controlled solubility of the fertilizer in water, protein materials, occlusion,  
362 chemicals, slow hydrolysis of water soluble compounds of lower molecular weights and  
363 some other unknown means (Naz and Sulaiman, 2016). Other options include utilization of  
364 semi-permeable materials and sensors of chemical or biological origin within the fertilizer  
365 (Fig. 4a). These are advanced materials, whose physical or chemical properties can change

366 in response to an external stimulus such as temperature, pH, and electric or magnetic fields  
367 (Foster, 2013; Mastronardi et al., 2015; Roy and Gupta, 2003).

368 Many coating materials can be used to slow nutrient release, including natural  
369 materials such as clays and nanoclays (e.g., allophane), non-degradable (polysulfone) and  
370 biodegradable polymers (e.g., alginate beads) (Table 1). Several studies have shown that  
371 urea based coatings can have variable efficiencies depending on the material used (Du et  
372 al., 2006; Naz and Sulaiman, 2016; Shaviv, 2005). In addition, they can be highly  
373 expensive in some cases, pollutants, or toxic (i.e., polymer coated urea) and mostly difficult  
374 to degrade with potential environmental impacts (accumulation) as the major concern. As a  
375 consequence, formulation with environmentally safe and biodegradable coating materials  
376 seems to be necessary (Ni et al., 2009). In this sense, natural polymeric carbohydrates  
377 appear as an alternative to non-biodegradable materials acting as permeable or  
378 impermeable membranes with tiny pores in slow release fertilizers (e.g., urea) (Butzen,  
379 2013). These highly degradable materials have also received attention because of their low  
380 cost and low environmental damage due to biodegradability and low accumulation in the  
381 environment (Naz and Sulaiman, 2016). However, these polymers may need some  
382 modifications to be included into the coating design because of their hydrophilic properties  
383 and their weak coating barrier, for example in the case of starch.

384

### 385 **3.2 Bioformulation fertilizer: plant growth promoting and nutrient use efficiency**

386 One group of microorganisms beneficial for plant growth are PGPR, a  
387 heterogeneous group of bacteria that can be found in the rhizosphere, at root surfaces and in  
388 association with roots (Ahmad et al., 2008). These bacteria have several functions,  
389 including production and regulation of phytohormones, release of nutrients to plants (e.g.,



390 P-, N-fixation, siderophores, among others), and control of phytopathogens (production of  
391 antibiotics and siderophores) (Egamberdieva and Adesemoye, 2016; Martínez-Viveros et  
392 al., 2010; Zahid et al., 2015).

393 Phosphobacteria, phytate-mineralizing bacteria, and phosphate-solubilizing bacteria  
394 have been commonly isolated from soil and proposed as inoculants for agricultural  
395 improvement (Jorquera et al., 2008). For example, a large diversity of microbes in  
396 Andisols under pastures and cereal crops are capable of mineralizing phytate (Jorquera et  
397 al., 2008; Martínez-Viveros et al., 2010; Menezes-Blackburn et al., 2014). They may be  
398 used to develop bacterial or enzyme systems as biofertilizers to overcome the limitations of  
399 conventional fertilizers in acidic soils, as well as for developing added value products from  
400 agricultural wastes. To this end, Calabi-Floody et al. (2012) studied the effect of enzyme-  
401 nanoclay complexes on P availability of composted cattle dung and showed that it  
402 increased the inorganic P content. Moreover, Menezes-Blackburn et al. (2014) suggest that  
403 inoculation of cattle manure with enzyme-nanoclay complexes enhances the organic P  
404 cycling and P nutrition of plants grown in P deficient soils.

405 Low N acquisition by plants is a limiting factor in agricultural ecosystems and there  
406 is interest in using N<sub>2</sub>-fixing bacteria as an alternative to conventional fertilization.  
407 Diazotrophic bacteria are capable of converting atmospheric dinitrogen (N<sub>2</sub>) into NH<sub>3</sub>,  
408 which can be used by plants (Ilsam et al., 2009). Among them, a number of free-living soil  
409 bacteria are considered to be PGPR, because of their competitive advantage in C-rich and  
410 N-poor environments (Kennedy et al., 2004). Free-living N<sub>2</sub>-fixing bacteria have been  
411 considered as an alternative to conventional N-fertilizer for promoting plant growth, and  
412 several research studies reported significant increases in grain and shoot biomass yield from  
413 plants inoculated with free-living diazotrophic bacteria (Andrade et al., 2013; Barua et al.,

414 2012; Kennedy et al., 2004; Park et al., 2005). Moreover, Vadakattu and Paterson (2006)  
415 reported that under intensive wheat rotation at Avon, South Australia, free-living N<sub>2</sub>-fixing  
416 bacteria contributed 20 kg N per ha per year, which represented 30–50% of total crop  
417 requirements. This response was attributed to a combination of factors including  
418 enhancement of root development, production of growth regulators, and N fixation  
419 (Naiman et al., 2009). However, it is well known that bacteria directly inoculated in the soil  
420 system could be adversely affected by competition with native microorganisms,  
421 unfavorable physicochemical conditions and fluctuating pH and temperature (Bréant et al.,  
422 2002).

423 Encapsulating microorganisms in carrier materials (bioformulation) is designed to  
424 protect them during storage and from adverse environmental condition (pH, temperature,  
425 etc) (Fig. 4b), thus ensuring a gradual and prolonged release (Bashan, 1986; Kim et al.,  
426 2012). A wide range of microorganisms have been investigated and a framework for  
427 selecting suitable organisms for specific purposes has been developed (Table 2).

428 Materials suitable for immobilization and preservation of bacteria include alginate  
429 gels, synthetic gels (Sol-Gel), polyacrylamide, agar and agarose, polyurethane, vermiculite  
430 and polysaccharides (Bashan, 1998; Liu et al., 2008). In addition, composite materials  
431 based on biodegradable polymer-clay or nanoclays are being studied, including  
432 nanocomposites (Calabi-Floody et al., 2009). For example, encapsulation of free-living  
433 diazotrophic bacteria has been considered as one of the possible alternatives for inorganic  
434 N-fertilizer for promoting plant growth and crop yield (Ivanova et al., 2005).

435

#### 436 **4. Smart fertilizer formulations**

#### 437 **4.1 Nanofertilizers**

438 Nanofertilizers, as smart fertilizers are designed to increase nutrient use efficiency  
439 and consequently reduce adverse effects on the environment compared to application of  
440 conventional mineral fertilizers (Manjunatha et al., 2016; Sharpley et al., 1992; Wurth,  
441 2007). According to Mastronardi et al. (2015) there are three main types of nanofertilizers:  
442 nanoscale fertilizer (synthesized nanoparticles), nanoscale additives (bulk products with  
443 nanoscale additives), and nanoscale coating or host materials (product coated with  
444 nanopolymer or loaded with nanoparticles) (Table 3).

445 Slow-release nanofertilizers and nanocomposites are suitable alternatives to soluble  
446 fertilizers. Nutrients are released at a slower rate during crop growth, thereby reducing loss.  
447 Slow release of nutrients in the environments could be achieved by using zeolites (natural  
448 clays), which acts as a reservoir for nutrients that are released slowly (Manjunatha et al.,  
449 2016). The mineral nutrients required for plant nutrition can be encapsulated inside  
450 nanomaterials such as nanotubes or nanoporous materials, coated with a thin protective  
451 polymer film, or nanoscale particles (DeRosa et al., 2010; Manjunatha et al., 2016).

452 Depending on the application, it is possible to use synthetic or natural nanoparticles  
453 obtained from various sources, including plants, soils and microorganisms (Table 3)  
454 (Calabi-Floody et al., 2011, 2009; Panpatte et al., 2016; Tarafdar et al., 2012). Nanoclays,  
455 which naturally occur in soils, have been considered important tools in modern agriculture  
456 due to their physicochemical properties (Sekhon, 2014). Nanoclays can be used to stabilize  
457 enzymes and thereby increase their catalytic activity for different biotechnological purposes  
458 (Calabi-Floody et al., 2009; Kim et al., 2006; Moelans et al., 2005; Wang, 2006). For  
459 example, Menezes-Blackburn et al. (2011) studied the effect of synthetic allophane,  
460 synthetic iron-coated allophanes and natural montmorillonite as solid supports of phytases

461 and observed that immobilization patterns at different pH values were strongly dependent  
462 on both enzyme and support characteristics. In addition, they concluded that  
463 montmorillonite appeared as a good immobilizing support only for the *Escherichia coli*  
464 phytase, while it was an inhibitor for *Aspergillus niger* phytase activity. Calabi-Floody et  
465 al. (2012) evaluated and implemented the use of natural clays and nanoclays (from  
466 montmorillonite and allophanic clays) as support materials for acid phosphatase (AP) in  
467 nanoclay-cattle dung-AP complexes. They found a clear stabilization of AP by these  
468 materials through encapsulation. The authors reported an increase of both specific activity  
469 (up to ~ 48%) and  $V_{max}$  (up to 38%) of the enzyme. They also observed that AP  
470 immobilized on allophanic nanoclays enhanced the release of inorganic P from cattle dung  
471 compared with free AP. Positive growth responses were found for P nanoparticles applied  
472 to basil (*Ocimum basilicum*) under salt stress (Alipour, 2016) and for synthetic apatite  
473 nanoparticles (solid P nanofertilizer) applied to soybean (*Glycine max*) (Liu and Lal, 2014).

474 Nanocomposites are hybrid materials consisting of a continuous (polymer) phase or  
475 matrix and a dispersed (nanofiller) phase. The dispersion of a small amount (< 10%) of  
476 nanomaterial in the polymer matrix can lead to marked improvement in physical and  
477 mechanical properties (strengths, pH tolerance, storage stability, heat distortion, break  
478 elongation) compared with a single polymer matrix (Calabi-Floody et al., 2009). Currently  
479 research is focused to develop nanocomposites to supply essential nutrients through smart  
480 delivery system (Manjunatha et al., 2016), synchronizing the release of them with the crop  
481 uptake, so preventing undesirable nutrient losses to soil (e.g., leaching and volatilization)  
482 (Bley et al., 2017; DeRosa et al., 2010). Coating and cementing of nano and  
483 subnanocomposites can also be used to regulate the release of nutrients from the fertilizer  
484 capsule (Liu et al., 2006). Therefore, the future development should focus on materials

485 allowing for nutrient release from nanofertilizers triggered by an environmental condition  
486 or simply at specific time (Gruère, 2012). In this context, nanodevices or additives (e.g.,  
487 nanotubes, aptamers, double hydroxide-nanocomposites, urease enzymes, nanosize  
488 titanium dioxide, nanosilica particles) can be associated to nanofertilizers to synchronize  
489 the fertilizer release with plant demand (DeRosa et al., 2010; Foster, 2013).

490

## 491 **4.2 Other smart fertilizers formulations**

492 *4.2.1 Polymers:* Polymers are widely used in agriculture especially for fertilizer  
493 development. Smart polymeric materials have been applied to smart delivery systems of a  
494 wide variety of agrochemicals (Puoci et al., 2008). A broad range of synthetic materials,  
495 such as petroleum based polymers have been used to encapsulate water-soluble fertilizers.  
496 Polysulfone, polyacrylonitrile, polyvinyl chloride, polyurethane and polystyrene are the  
497 main materials currently used for coating (Ibrahim and Jibril, 2005; Lü et al., 2016; Tao et  
498 al., 2011). Jarosiewicz and Tomaszewska (2003) compared the use of the synthetic  
499 polymers polysulfone and polyacrylonitrile and the biodegradable cellulose acetate for the  
500 development of slow release fertilizers. They observed that physical properties of the  
501 coatings can influence the release rate of macronutrients (N, P, and K), which are present in  
502 the core of the coated fertilizers. They found that synthetic non degradable materials had a  
503 slower release rate than cellulose acetate based ones. Tao et al. (2011) studied the use of a  
504 triple polymer fertilizer to encapsulate and enhance the mechanical properties of urea. They  
505 suggested that polyethylene in a first layer, poly (acrylic acid-co-acrylamide) as  
506 superabsorbent in a second layer, and poly (butyl methacrylate) in the third layer improve  
507 the controlled release of urea. They also observed that the incorporation of this triple  
508 polymer fertilizer into soil improved its water holding capacity, which in turn enhanced

509 nutrient uptake and crop yield. Liu et al. (2008) replaced sulfur with dicyclopentadiene to  
510 improve coating properties, moisture resistance, abrasion resistance and mechanical  
511 strength of slow release fertilizers. They found that the mechanical strength of the coating  
512 was directly related to the dicyclopentadiene content, which ameliorated nutrient release  
513 efficiency.

514

515 *4.2.2 Biodegradable polymers:* These materials have increasingly been used as substitutes  
516 of others polymers in agriculture. Devassine et al. (2002) divided them in two main groups  
517 according to their water vapor permeability, namely degradable synthetic polymers with a  
518 small permeability coefficient ( $K < 3000 \text{ cm}^2 \text{ s}^{-1} \text{ Pa}^{-1}$ ) (biopols, polylactic acids and  
519 polycaprolactone), and modified polysaccharides with a higher permeability coefficient  
520 ( $K > 4000 \text{ cm}^2 \text{ s}^{-1} \text{ Pa}^{-1}$ ) (alginates, starchs, agar). Several studies have reported on the  
521 utilization of these degradable polymers for a wide range of nutrients. For example, Perez  
522 and Francois (2016) prepared microspheres with chitosan and chitosan-starch blends using  
523 sodium tripolyphosphate aqueous solution as the crosslinking agent. They observed that  
524 these materials could be used as slow release N and K fertilizers. Zhang et al. (2016) have  
525 reported the development of polymer-coated N-fertilizer using bio-based polyurethane  
526 derived from liquefied locust sawdust as coating material. They found that this fertilizer  
527 was more efficient at supplying N to maize than conventional urea.

528 Biodegradable polymers have also been used in bioformulations, acting as microbial  
529 carriers. These carriers protect microbial inoculants from various stresses and prolong shelf  
530 life (Ardakani et al., 2010; Egamberdieva and Adesemoye, 2016; Kumar et al., 2007). For  
531 example, calcium alginate gel may protect microbial cells with a concomitant increase in  
532 shelf life (Wu et al., 2011). Sodium alginates are widely used for bioformulations (i.e.,

533 bacterial fertilizers) and with pesticides (Liang et al., 2007; Liu et al., 2008; Singh et al.,  
534 2009). Despite the low cost and the environmental friendly properties of biodegradable  
535 polymers, in many cases the properties of these materials need to be blended with synthetic  
536 materials to improve their performance (Puoci et al., 2008)

537

### 538 **4.3 Use of harvesting residues for smart fertilizer formulations**

539 *4.3.1 Lignocellulosic straw as carrier and coating material:* Low-cost materials such as  
540 wheat straw are abundantly available resources in current agricultural systems (Jiang et al.,  
541 2012). These harvesting residues contain lignin, hemicelluloses, and cellulose (Hubbe et al.,  
542 2010). Cellulose fibrils and lignin impart mechanical strength properties (Panthapulakkal  
543 and Sain, 2015). Wheat straw contains surface carboxyl, hydroxyl, ether, amino and  
544 phosphate, which enhance its reactivity and physicochemical properties, useful in the  
545 preparation of adsorbent materials for the treatment of wastewater (Wang et al., 2016) and  
546 slow release fertilizers (Liu et al., 2013; Xie et al., 2011). Moreover, some researchers have  
547 found that wheat straw can be used as reinforcements and/or fillers for nonstructural and  
548 structural composites, (Panthapulakkal and Sain, 2015).

549 Xie et al. (2011) noticed the potential use of wheat straw for the development of  
550 slow release N and boron fertilizers with water-retention properties. The authors prepared  
551 and used the straw as skeletal material in copolymerization with other monomers to form a  
552 superabsorbent material. They introduced inorganic fertilizers (urea and borax) in order to  
553 develop an organo-mineral fertilizer within a core/shell structure. They found that the final  
554 product contained 23.3% N and 0.65% boron with potential slow release characteristics.  
555 Wang et al. (2016) designed a multifunctional strategy for the development of a slow  
556 release compound fertilizer, which was prepared by recovery of  $\text{NH}_4^+$  and  $\text{P}_2\text{O}_4^-$  from

557 aqueous solutions onto an amphoteric straw cellulose adsorbent. The maximum  $\text{NH}_4^+$   
558 adsorption capacity of the material was  $68.4 \text{ mg g}^{-1}$ , whereas the adsorption capacity for  
559  $\text{P}_2\text{O}_4^-$  was  $38.6 \text{ mg g}^{-1}$  at pH 7 and 5, respectively.

560 On the other hand, cellulose obtained from agricultural residues has been also used  
561 in bioformulations as carrier for bacterial inoculants with broad-spectrum antifungal  
562 activity and suppression of fungal pathogens (Egamberdieva and Adesemoye, 2016; Negi et  
563 al., 2005). Albareda et al. (2008) studied the survival of different PGPR strains on various  
564 carrier and liquid formulations and found that compost and perlite were very effective.  
565 However, lignocellulose and compost are subject to rapid decomposition once incorporated  
566 into soil. In order to further improve their properties as slow release fertilizers, they could  
567 be combined with clay minerals or biochar to reduce their decomposition (Barthod et al.,  
568 2016; Ngo et al., 2016).

569

570 *4.3.2 Biochar as carrier and coating material:* Harvesting residues, such as straw may also  
571 be used as feedstock for energy producing pyrolysis systems with biochar generation.  
572 Considering its physicochemical properties, carbonaceous materials like pyrogenic carbon  
573 (biochar) have been widely used as soil ameliorant with several applications in both  
574 laboratory and field studies (Glaser, 2015; Wiedner et al., 2015). Biochar is obtained  
575 through pyrolysis of agricultural or other lignocellulosic biomass at temperatures ranging  
576 from  $350 - 700^\circ\text{C}$  (Glaser et al., 2002; Lee et al., 2013; Lehmann and Joseph, 2009).  
577 Biochar was found to increase the C sequestration potential of soil through its high stability  
578 and the reduction of native soil OM mineralization (Naisse et al., 2015; Ventura et al.,  
579 2015) and to be an excellent microbial habitat (Lehmann et al., 2011). González et al.  
580 (2015) studied the influence of different polymers and biochar produced from oat hull as



581 support material on the release and leaching of N from urea under greenhouse conditions.  
582 They found that urea formulated together with biochar slowed down the release of this  
583 fertilizer. Nevertheless, biochar did not act as a nutrient release retarder agent and may need  
584 other polymeric materials as encapsulating agent to control N leaching from controlled  
585 release fertilizers. Cai et al. (2016) found that biochar produced from corncob, banana stalk  
586 and pomelo peel displayed an excellent retention ability in holding  $\text{NH}_4^+$  associated to the  
587 presence of carboxyl and keto groups when the material was prepared at 200 °C, suggesting  
588 that this material could be used as a slow release carrier for N. Zhao et al. (2016) observed  
589 that the combination of P fertilizers (triple superphosphate and bone meal) pre-mixed with  
590 sawdust and switch grass biomass prior to biochar production was a good strategy for the  
591 production of an effective slow release P fertilizer.

592         Recent studies have also investigated the use of biochar and charcoals as carriers in  
593 combined formulations with beneficial microorganisms. Biochar was a useful carrier for the  
594 bacterial population of *Enterobacter cloacae* (Hale et al., 2015) and *Azospirillum lipoferum*  
595 (AZ 204) (Saranya et al., 2011).

596         The use of biochar as carrier for smart fertilizers could be highly beneficial, as it  
597 combines nutritional benefits for plants with improvement of many other soil functions due  
598 to the addition of biochar itself. In particular, biochar addition to soils has positive effects  
599 on water holding capacity as well as C sequestration. However, biochar properties vary  
600 widely depending on feedstock and production conditions (Wiedner et al., 2013). Thus, the  
601 use of these kinds of formulations presents new challenges related to the optimal  
602 combination of carrier materials and inoculants. Considering the varying properties of  
603 carrier materials here reviewed and the variety of potential utilization for smart fertilizers  
604 designs, more research is needed for their development.

605

## 606 **5. Conclusions and future directions**

607         In order to meet sustainable development goals, agricultural production needs to be  
608 increased and the pollution and GHG emissions related to farming activity need to be  
609 decreased. We suggest that advances in the application of biotechnology and  
610 nanotechnology have the potential to facilitate improved nutrient management and use  
611 efficiency in agroecosystems. Smart fertilizers based on slow/controlled release and/or  
612 carrier delivery systems have been shown to improve crop yields, soil productivity and  
613 lower nutrient loss compared with conventional fertilizers. Several materials such as clays,  
614 nanoclays, non-degradable and degradable polymers, and agricultural wastes are suitable  
615 for the development of smart fertilizers by acting as carrier matrices for nutrients and  
616 bacterial inoculants. Future research should continue to explore and evaluate the  
617 composition, manufacture, and agronomic and environmental performance of various smart  
618 fertilizers, especially those that utilise organic waste materials. We suggest that  
619 lignocellulosic organic waste, such as straw after chemical, physical or thermal  
620 transformations may be an excellent carrier or coating material for fertilizer formulations.  
621 Such organic wastes occurring as harvesting residues in agricultural systems should be used  
622 in the sense of a circular economy to create innovative fertilizers from natural materials,  
623 which are urgently needed to ensure sustainable intensification of agricultural systems.

624

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630

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**Table 1** Carriers and coating materials suitable for the development of smart fertilizers

Materials	Application	References
Brown coal, charcoal and biochar	<i>N retention</i>	Ding et al. (2010)
	<i>Support material in polymer biodegradable formulations of controlled release fertilizer</i>	González et al. (2015)
	<i>Sustained-release fertilizer</i>	Cai et al. (2016)
	<i>Slow release N fertilizer</i>	Rose et al. (2016)
	<i>Biochar–fertilizer composite</i>	Joseph et al. (2013)
	<i>Inoculum carrier on PGPR bioformulations</i>	Hale et al. (2015)
Perlite, vermiculite and bentonite, attapulgit	<i>In superabsorbent composites of controlled released fertilizers</i>	Wu and Liu (2008), Zhang et al. (2006)
	<i>Carrier in bioformulations of bacterial inoculants</i>	Daza et al. (2000); Khavazi et al. (2007); Ardakani et al (2010) Sangeetha (2012) Thompson (1980)
Peat	<i>N controlled fertilizer</i>	Araújo et al. (2017)
	<i>Microbial carrier in bioformulations</i>	Albareda et al (2008)
Alginate beads, calcium alginate gel	<i>Double coated control release fertilizer</i>	Wang et al. (2016)
	<i>Microencapsulated bacterial fertilizer/ Encapsulation of microorganisms in Alginate-clay complexes</i>	Trivedi et al. (2005); Wu et al. (2011) Fravel et al. (1985); Campos et al. (2014)
	<i>Coating materials of controlled release fertilizers</i>	
	<i>Non environmental friendly polymers (Polyurethane, Polyacrylic acid etc)</i>	Tao et al. (2011); Golden et al. (2011); Ibrahim and Jibril (2005) Jarosiewicz and Tomaszewska (2003); Donida and Rocha (2002) 59

<p>Polymeric materials</p>	<p><i>Environmental friendly polymers; Syntethic (Polycaprolactone; ethyl celulose ) and non-modified (Starch, agar)</i></p>	<p>Lan et al. (2011), da Rosa and Rocha (2012) Shaviv (2005); Rose (2002)</p> <p>Devassine et al. (2002); Niu and Li (2012); Yong et al. (2005); Mathews and Narine (2010); Jintakanon et al. (2008); Costa et al. (2013); Azeem et al. (2014)</p>
<p>Saw dust, locust saw dust and wood ashes</p>	<p><i>Polymer-coated N fertilizer using bio-based polyurethane derived from liquefied locust sawdust</i></p> <p><i>Carrier in bioformulations</i></p>	<p>Zhang et al. (2016)</p> <p>Arora et al (2014)</p>
<p>Chitosan and Humics, Modified humic substances</p>	<p><i>N controlled fertilizer</i></p> <p><i>Slow release N from modified humics</i></p> <p><i>Carrier and bioprotector in bioformulations</i></p>	<p>Araújo et al. (2017)</p> <p>Kulikova et al. (2016)</p> <p>Silva et al. (2016); Murphy et al. (2003)</p>
<p>Lignin, cellulosic materials, wheat bran, spent mushroom compost</p>	<p><i>Coating in controlled release fertilizer/Pyrolized lignocellulosic material in slow-release N fertilizer</i></p> <p><i>Carrier in blue green algal biofertilizer/ carrier of inoculant</i></p>	<p>Detroit et al. (1988); Mulder et al. (2011); Li et al. (2017)</p> <p>Bahl and Jauhri (1986); Jackson et al. (1991); Dhar et al. (2007)</p>

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1400 **Table 2** Quality criteria of carriers for the development of smart fertilizers based on  
 1401 microbial inoculants adapted from Sahu and BrahmaPrakash (2016).

Quality criteria of model carriers of Bioformulations	References
High water-holding and water-retention capacity and suitable for as many bacteria as possible/ Cost-effective	<i>Mishra and Dahich (2010)</i>
Free from lump-forming material/ Near sterile or easy to sterilize by autoclaving or by other methods like gamma irradiation/ Nearly neutral pH or easily adjustable and good pH buffering capacity	<i>Keyser et al. (1993)</i>
Available in adequate amounts/ Nontoxic in nature	<i>Bazilah et al. (2011)</i>
For carriers used for seed treatment, should assure the survival of the inoculants on the seed since normally seeds are not immediately sown after seed coating	<i>Muresu et al. (2003)</i>
For carriers that shall be used for seed coating, should have a good adhesion to seeds	<i>Hegde and BrahmaPrakash (1992 )</i>
No heat of wetting/ Easily biodegradable and nonpolluting/ Supports growth and survival of bacteria/ Amenable to nutrient supplement/ Manageable in mixing, curing, and packaging operations	<i>Smith (1992)</i>
Chemically and physically uniform	<i>Bashan (1998)</i>
The inoculant should be nontoxic, biodegradable and nonpolluting, and should minimize environmental risks such as the dispersal of cells to the atmosphere or to the ground water.	<i>Bashan (1998)</i>

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**Table 3.** Nanofertilizers for plant nutrition. Modified from Subramanian et al. (2015)

Nutrients	Absorbent	Size	Reference
Nitrogen (N)	Zeolite	7-10 nm	<i>Mohanraj (2013)</i>
		20-30 nm	<i>Subramanian and Sharmila Rahale (2013)</i>
		60 nm	<i>Selva Preetha (2011)</i>
		87 nm	<i>Manikandan and Subramanian (2014)</i>
		200 nm	<i>Komarneni (2010)</i>
	420 $\mu$ m	<i>Li et al. (2003)</i>	
	Montmorillonite	35-45 nm	<i>Subramanian and Sharmila Rahale (2013)</i>
	50 $\mu$ m	<i>Bortolin et al. (2013)</i>	
	<i>Carbon nanotubes, hydrogels, organic zeolitic complexes</i>	40-80nm	<i>DeRosa et al. (2010); Liu et al. (2006); Leggo (2000); Foster (2013)</i>
Phosphorus (P)	Zeolite	25-30 nm	<i>Subramanian and Sharmila Rahale (2013)</i>
		60nm	<i>Selva Preetha (2011)</i>
		2-3 $\mu$ m	<i>Bansiwal et al. (2006)</i>
	<i>Montmorillonite, bentoninete and apatite</i>	35-40nm	<i>Subramanian and Sharmila Rahale (2013); Liu and Lal. (2014)</i>
Potassium (K)	Zeolite	25-30 nm	<i>Subramanian and Sharmila Rahale (2013)</i>
	Montmorillonite	35-40 nm	<i>Subramanian and Sharmila Rahale (2013)</i>
NPK	<i>Nano-coating of sulfur layer chitosan</i>	78-100 nm	<i>Wilson et al. (2008)</i>
Nanocomposites	<i>Kaolinite</i>	30-80 nm	<i>Xu-mei et al. (2006)</i>
Sulfur (S)	Zeolite	60 nm	<i>Selva Preetha et al. (2014)</i>
		70-93nm	<i>Thirunavukkarasu (2014)</i>
		420 $\mu$ m	<i>Li and Zhang (2010)</i>

Zinc (Zn) iron (Fe) and Borom (B)		25–30 nm	<i>Subramanian and Sharmila Rahale (2013)</i>
	<i>Zeolite</i>	60 nm	<i>Selva Preetha (2011)</i>
		---	<i>Hu et al. (2016)</i>
	<i>Montmorillonite</i>	35-40 nm	<i>Subramanian and Sharmila Rahale (2013)</i>
	<i>Nano Zn and Nano ZnO</i>	35-20 nm	<i>Nair et al. (2010); Mahajan et al. (2011)</i>
PGPR microorganisms and biomolecules as enzymes	<i>Gold nanoparticles</i>	---	<i>Shukla et al. (2015)</i>
	<i>Nanoclays (allophane)</i>	100 nm	<i>Calabi-Floody et al. (2012, 2009); Menezes-Blackburn et al. (2011)</i>

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1405 **Figure captions**

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1407 **Figure 1.** Global phosphate rock reserves, data taken from US Geological Survey estimates  
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1410 **Figure 2.** Results from searches of the Scopus database of scientific documents (grey bars)  
1411 and patents (grey lines) with the Title-abstract-keywords ‘nanotechnology AND  
1412 agriculture’ (<https://www.scopus.com>, accessed March 24, 2017).

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1414 **Figure 3.** Schematic diagram of smart fertilizer effects in the soil-plant system

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1416 **Figure 4.** Schematic representation of smart delivery systems: a) advanced polymeric  
1417 materials, degraded under external stimulus such as temperature, pH, and with water  
1418 permeability to achieve a slow nutrient release; b) microorganism encapsulation

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