

1 **Flying the extra mile pays-off: foraging on**
2 **anthropogenic waste as a time and energy-saving**
3 **strategy in a generalist bird**

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17
18 **Abstract**

19 Food waste disposal represents a major global source of predictable anthropogenic food
20 subsidies and is exploited by many organisms. However, the energetic cost-benefits of
21 foraging on these food subsidies have remained largely unexplored. Here we investigate
22 the year-round foraging decisions of resident white storks, *Ciconia ciconia*, in Iberia and
23 assess the energetic and time cost-benefits of foraging on both landfill waste and natural
24 food sources. To do so, we use GPS and acceleration data from 55 individuals tagged in
25 southern Portugal between 2016 and 2019. We find that the probability of attending

26 landfill sites was 60% during the non-breeding season and 44% during the breeding
27 season. Moreover, foraging on landfill waste is a time- and energy-saving strategy;
28 although birds had to travel 20% further to exploit this resource during the breeding
29 period, they spent overall 10% less energy than when foraging on natural prey. We show
30 that this relationship could be mediated by a reduction in foraging time and an increase
31 in foraging efficiency while exploiting landfill waste. Surprisingly, we did not find any
32 evidence that landfill specialists experienced any competitive advantage during landfill
33 exploitation over birds that visit landfills occasionally. These insights are key to predict
34 how species that rely on landfills can be affected by waste reduction initiatives planned
35 by the European Union, and implement the necessary management strategies.

36

37 **Keywords:** landfill closure; landfill waste; foraging ecology; predictable anthropogenic
38 food subsidies; white storks.

39

40 **1. Introduction**

41 Humans produce enormous quantities of food waste; estimates suggest that 30-40% of all
42 food produced is wasted (Parfitt et al., 2010) and deposited in locations where it can be
43 accessible to wildlife. These predictable anthropogenic food subsidies (PAFS), in the
44 form of organic waste on landfills, fisheries discards or crop residuals, generate impacts
45 on animal populations at multiple scales, from the individual to the ecosystem (Oro et al.,
46 2013). Individuals from numerous animal species have modified their movements,
47 activity, geographical range, and home range size in response to PAFS (Gilbert et al.,
48 2016; López-López et al., 2014; Newsome et al., 2015; Patrick et al., 2015; van Donk et
49 al., 2019); which can affect survival and reproduction rates, and subsequently the
50 demography of these populations (Plaza and Lambertucci, 2017).

51 Landfill sites, in particular, can potentially sustain high densities of scavenging
52 individuals (Oro et al., 2013). The spatial and temporal predictability, accessibility, and
53 nutritional value of anthropogenic food waste can provide abundant food resources for
54 wildlife, and has been shown to be responsible for the demographic explosion of
55 generalist animals (such as foxes, rats and gulls), but also for the sustenance of some
56 endangered species (Plaza and Lambertucci, 2017). Landfill use has been associated with
57 increased body mass in black vultures (*Coragyps atratus*) (Plaza and Lambertucci, 2018),
58 kelp gull (*Larus dominicanus*) nestlings (Lenzi et al., 2019), and grizzly bears (*Ursus*
59 *arctos horribilis*) (Blanchard, 1987). Landfill use has also been linked to higher
60 reproduction performance; for example, white storks (*Ciconia ciconia*) had larger clutch
61 sizes (Djerdali et al., 2008), glaucous gulls (*Larus hyperboreus*) experienced higher
62 fledging rate (Weiser and Powell, 2010), and laughing gull (*Leucophaeus atricilla*) chick
63 growth and survival was enhanced (Dosch, 1997), compared to conspecifics foraging in
64 more natural landscapes.

65 According to the optimal foraging theory animals choose to forage in sites that maximize
66 energy intake while minimizing energy and time expenditure (Stephens and Krebs, 1986;
67 Ydenberg et al., 1994). Landfill waste offers large quantities of high-energy food
68 (Patenaude-Monette et al., 2014; van Donk et al., 2019) but is localised in space; hence
69 trade-offs exist between time spent in landfills sites and energetic costs of travelling,
70 especially for central-place foragers during the breeding season (Gilbert et al., 2016).
71 Foraging in landfill sites is likely to exacerbate inter and intraspecific competition
72 (through agonistic interactions and food-robbing), which increases energetic costs, since
73 they attract large numbers of individuals and at high densities (Oro et al., 2013). On the
74 other hand, on natural heterogeneous landscapes, prey is often patchily distributed and
75 intraspecific competition is likely to be low since individuals tend to forage in smaller

76 groups (Catry et al., 2017). These dynamics could lead to the emergence of individual
77 foraging strategies and specialisation, with less competitive individuals avoiding landfill
78 sites. Despite many studies focusing on the fitness benefits for individuals exploiting
79 landfill waste, the energetic cost-benefits of landfill use have not yet been fully quantified.
80 Use of landfill resources, together with the global increase of temperature due to climate
81 change, may have facilitated the establishment of non-migratory white stork populations
82 in Iberia (Catry et al., 2017). Resident individuals heavily rely on food waste disposal
83 sites for foraging and no longer complete their annual migrations to and from their sub-
84 Saharan wintering grounds. In two decades the number of resident white storks in
85 Portugal has increased from 1,187 individuals (18% of the breeding population) in 1995
86 to 14,434 (62% of the breeding population) in 2015 (Catry et al., 2017). White storks'
87 use of landfill resources in Iberia has been investigated, and evidence shows that storks
88 nesting close to landfill sites heavily relied on them (Gilbert et al., 2016). Moreover,
89 breeding success for these individuals was higher than for individuals nesting further
90 away (Gilbert, 2015).

91 Whilst this increasing number of white storks in Iberia is widely attributed to their high
92 adaptability and behavioral plasticity, new EU directives (1999/31/UE and, more
93 recently, 2018/850/UE), regulating waste disposal, can greatly revert this trend. These
94 directives have established new circular economy targets aiming to reduce municipal
95 waste landfilled to 10% in the next decade. Recent evidence shows that landfill closure
96 can cause declines on survival, body mass, egg volume and/or clutch size in several gull
97 species (Payo-Payo et al., 2015; Pons and Migot, 1995; Steigerwald et al., 2015) (but see
98 (Katzenberger et al., 2019)). Therefore, it is crucial to understand foraging decisions, and
99 how animals search for and exploit landfill waste, as opposed to natural prey, to predict

100 how animal populations might respond when food waste is removed, and to develop
101 appropriate conservation and management strategies.

102 Here we investigate (i) the foraging decisions of resident white storks feeding on PAFS
103 and natural food sources in Iberia; (ii) the energetic and time cost-benefits of these
104 foraging decisions during the breeding and non-breeding seasons; and (iii) whether birds
105 highly specialised on landfills gain a competitive advantage while exploiting PAFS, by
106 increasing their foraging efficiency and decreasing the foraging time necessary to meet
107 their energetic requirements, over birds that only visit landfills occasionally. These
108 insights will be key to understand the drivers of landfill use and to predict how storks —
109 and to some extent other birds found regularly in large numbers at landfills (e.g. egrets,
110 herons, gulls and some raptors) — may be affected by reductions in PAFS. This
111 understanding is key for a species that has substantially increased in numbers in the last
112 decades due to the exploitation of landfills (Catry et al., 2017), and for which density
113 dependent effects of food depletion are associated with high nestling mortality (Denac,
114 2006; Zurell et al., 2015).

115 In this study we examine 4 years of GPS and tri-axial acceleration data from resident
116 white storks in Iberia and determine their behaviour, energy expenditure, foraging
117 decisions, and landfill use. We use structural equation models (SEM) (Lefcheck et al.,
118 2016; Lefcheck, 2016) to understand the relationships among foraging movements,
119 energy expenditure and behaviour. We hypothesize that the choice of food resources
120 (anthropogenic waste or natural prey) will lead to different foraging strategies, for
121 example at landfills sites birds will reduce the time devoted to foraging and increase
122 energy efficiency given the spatiotemporal predictability of organic waste deposition.
123 Moreover, we predict that foraging strategies might differ between the breeding and non-
124 breeding season, as breeding birds will have to return to the nest frequently, so their

125 foraging range will be reduced, and landfill sites might then be outside the range. In this
126 context, fundamental trade-offs between energetic and nutritional requirements in
127 parental provisioning strategies are expected (Wright et al., 1998). Finally, we predict
128 that landfill specialists will have competitive advantage over birds that only use landfills
129 occasionally.

130 **2. Material and methods**

131 2.1 Bird capture and GPS tracking

132 Our dataset included GPS data with tri-axial acceleration from 55 resident adult white
133 storks, tagged between 2016 and 2019 in southern Portugal. Resident individuals
134 overwintered in the Iberian Peninsula and did not cross the Strait of Gibraltar. Storks were
135 tagged with ‘Flyway 50’ GPS/GSM loggers from Movetech Telemetry (4 different
136 models varying slightly in weight) and ‘Ornitrack-50’ GPS/GSM loggers from Ornitela.
137 Adult birds were caught at multiple landfill sites using nylon leg nooses, and at several
138 breeding colonies using a remotely activated clap net at the nests. Birds were measured
139 and ringed, and the devices were mounted on the back of the birds as backpacks with a
140 Teflon harness. The tag and harness together weighted 60-90 g, which represented 1.5-
141 3.7% of a given bird's body mass at the time of tagging. Most birds were resighted in the
142 days following tag deployment and no abnormal behaviour or adverse effects due to
143 tagging were observed. The procedure was approved by the Instituto da Conservação da
144 Natureza e Floresta (Portugal). The tags were programmed to record 9 consecutive GPS
145 positions at 1 Hz every 20 minutes, and concurrently a 9 s tri-axial acceleration burst at
146 1 Hz. We kept the first GPS position recorded, thus location and acceleration matched.
147 GPS data was visually examined to detect potential outliers, which were subsequently
148 removed, together with the associated acceleration burst. We identified 75 nest locations
149 for the tagged birds across the years after visually inspecting the GPS tracks and visiting

150 the sites. The nests were situated between 1.5 and 40.2 km away from the closest landfill
151 site (mean = 17.3 km).

152 2.2 Data selection and processing

153 From each acceleration burst we derived two metrics, ODBA (overall dynamic body
154 acceleration, $1\text{ G} = 9.8\text{ m/s}^2$), a proxy of energy expenditure invested in locomotion, and
155 behaviour (Gleiss et al., 2011; Shepard et al., 2008b). ODBA was obtained from tri-axial
156 acceleration bursts by subtracting the smoothing of total acceleration, using a running-
157 mean of 4 s, from the total acceleration, as recommended in (Gleiss et al., 2011; Shepard
158 et al., 2008a, 2008b). To determine the bird behaviour at each burst, we used the tri-axial
159 acceleration data to train random forest machine-learning algorithms (R package
160 ‘randomForest’ ver. 4.6 (Liaw and Wiener, 2002)). Movetech Telemetry tags and
161 Ornitela tags differ in their sensor characteristics (e.g. position of the sensor and axes
162 orientation) hence we created device specific algorithms. We characterised four
163 behaviours: foraging, resting (including preening), flapping, and soaring flight (including
164 gliding and orographic and thermal soaring). To train the algorithm we manually labelled
165 250 tri-axial acceleration bursts for each behaviour and tag type; 70% of the data was
166 used for training the algorithm and 30% for testing it. To label the training data we
167 compiled information from several sources: (i) 9 tags that were programmed to
168 continuously record GPS and acceleration data which allowed a detailed understanding
169 of the birds’ movements and behaviour, (ii) video recordings of captive white storks that
170 were fitted with the tracking devices, and (iii) from the 9 consecutive GPS locations that
171 allowed to infer birds’ speed and movement during the acceleration burst. The random
172 forest model had 96% accuracy for Movetech Telemetry tags data and 97% accuracy for
173 Ornitela tags data. For this study, flapping and soaring flight were aggregated and

174 considered as flying, as both behaviours occurred infrequently, representing only 5.7 %
175 of locations.

176 Landfill sites were determined through visual inspection of satellite images, and in all
177 cases, they were confirmed by visits or contact with local authorities. Each GPS location
178 was classified as either inside or outside a landfill. Subsequently, using the behavioural
179 and GPS data, we created daily foraging strategy metrics for each individual stork. We
180 only included in the analyses days with more than 10 daylight GPS positions and
181 acceleration bursts (median GPS positions per day was 29) and with more than 4 positions
182 classified as ‘foraging’ to guarantee representative metrics.

183 We built two different datasets for the subsequent analyses. First, to assess individual and
184 seasonal differences in attendance to landfill sites, we created a ‘daily attendance dataset’.
185 In this dataset each day was classified as a ‘landfill attendance day’ when at least one
186 foraging location occurred on a landfill in the corresponding 24-h period, or as a ‘non-
187 landfill attendance day’ when no foraging locations occurred on landfills. Second, to
188 understand the foraging movements, behaviour and energy expenditure of feeding on
189 anthropogenic waste or natural prey we built a ‘daily foraging strategy dataset’ that only
190 included days where birds displayed either one or the other strategy, thus removing days
191 where birds foraged on both waste and natural prey. To do so, we created a variable,
192 ‘foraging site’ with two levels: ‘landfill foraging day’, when 70% of the daily foraging
193 positions occurred in landfill sites, and ‘non-landfill foraging day’, when 70% of the
194 foraging positions for a given day occurred outside landfill sites. We chose this threshold
195 as it ensured birds spent most of the day in one of the two areas, while it minimised the
196 number of days that had to be removed from the dataset to less than 20%.

197 The activity and energetic trade-offs between foraging on landfills vs natural prey were
198 explored using several metrics: (i) daily distance travelled (km), calculated as the summed

199 distance between consecutive locations for each day; (ii) daily relative flight time,
200 calculated by dividing the number of burst classified as flying by the total of bursts
201 obtained in a day; (iii) mean ODBA (G) as a proxy of daily energy expenditure (Gleiss
202 et al., 2011), calculated as the mean ODBA of the accelerometer bursts obtained in a day;
203 (iv) daily relative foraging time, calculated by dividing the number of burst classified as
204 foraging by the total of burst in a day; and (v) mean foraging ODBA (G), as a proxy for
205 foraging efficiency (lower mean foraging ODBA values indicate higher foraging
206 efficiency), calculated as the daily mean ODBA for the foraging bursts. Daily distance
207 travelled and relative flight time were highly correlated (Pearson correlation = 0.75),
208 which indicates that the metrics were largely insensitive to the number of fixes; we only
209 retained daily distance travelled in further analyses.

210 To account for the potential effect of season on the foraging site choice (landfill sites or
211 non-landfill sites), we located the nesting site of each individual, and using the GPS data,
212 we defined the beginning of the breeding season as the first three consecutive days that
213 an individual occupied the nest; and the end of the breeding period as last three
214 consecutive days the nest was occupied.

215 To explore if birds that visited landfill sites more often had a competitive advantage when
216 exploiting this resource, we used the ‘daily attendance dataset’ to create a landfill
217 specialisation index, from 0 (natural prey specialist; never visits landfill sites) to 1 (waste
218 specialist; only forages on landfill sites), for each individual. We divided the number of
219 days that the bird attended a landfill by the total numbers of days that the bird was tagged.

220 2.3 Statistical analysis

221 To understand seasonal variability in foraging site attendance we fitted a generalised
222 linear mixed-effects model (GLMM) with foraging site attendance (landfill or non-
223 landfill) as the response variable and season (breeding or non-breeding) as the

224 explanatory variable, using the *glmmPQL* function with a binomial structure (R package
225 ‘MASS’ ver. 7.3 (Ripley et al., 2013)); the model included bird ID nested in tag type (5
226 levels: 4 types of Movetech tags and 1 type of Ornitela tag) as random effects and an
227 autocorrelation structure of order 1 to account for the potential temporal correlation
228 between consecutive days. Subsequently, to understand the potential constraints of
229 reproduction in the choice of foraging site, we used the data from the breeding season to
230 fit a GLMM with a binomial structure, using the *glmer* function (R package ‘lme4’ ver.
231 1.1-21 (Bates et al., 2015)). We considered foraging site attendance as the response
232 variable, and included as explanatory variables, distance from the nest site to the closest
233 landfill site, and its quadratic term, to account for potential non-linearity in the
234 relationship; we also included bird ID nested in tag type as random effects. We used the
235 R package ‘MASS’ when the inclusion of an autocorrelation structure was necessary as
236 ‘lme4’ does not allow the implementation of GLMMs with that structure.

237 Using the ‘daily foraging strategy dataset’, we explored the effect of landfill use on time
238 and energy budgets, by implementing 3 linear mixed-effects models (LMM) with daily
239 distance travelled (km), mean ODBA (G), and mean foraging ODBA (G) as response
240 variables and foraging site (landfill or non-landfill), season (breeding or non-breeding),
241 and its interaction, as explanatory variables, using the *lme* function (R package ‘nlme’
242 ver. 3.1 (Pinheiro et al., 2017)). The model with daily distance travelled as response
243 variable, included an additional covariate, the number of GPS positions, to account for its
244 potential effect on the response. We implemented a GLMM with the *glmmPQL*
245 function for relative foraging time with a binomial structure that included foraging site,
246 season, and its interaction as fixed effects as well. All four models included bird ID nested
247 in tag type as random effects and an autocorrelation structure of order 1. Subsequently,
248 we calculated for each model the coefficient of determination R^2 (Nakagawa et al., 2017).

249 We explored the relative contribution of individuals' foraging decisions into the overall
250 energy expenditure when using different foraging sites and during different seasons. To
251 do so, we implemented multigroup analysis for piecewise SEM with foraging site and
252 season as grouping variables, using the functions *psem* and *multigroup* (R package
253 'piecewiseSEM' ver. 2.1 (Lefcheck et al., 2016; Lefcheck, 2016)). The SEM included
254 two sub-models: 1. an LMM with mean ODBA as the response variable, and daily
255 distance travelled, relative foraging time and mean foraging ODBA as explanatory
256 variables; and 2. a GLMM with relative foraging time as the response variable, and mean
257 foraging ODBA as explanatory variables. Both models included random effects and
258 autocorrelation structure as described above. The global structure of the SEM model was
259 well supported according to the global goodness-of-fit: Fisher's $C = 0.098$ with $p = 0.952$.

260 To test the role of landfill specialisation on landfill exploitation we selected only days
261 classified as landfill foraging days. We modelled a LMM with mean foraging ODBA and
262 a GLMM with relative foraging time as the response variables, and season and
263 specialisation index as explanatory variables. We used random effects and an
264 autocorrelation structure as previously specified. Normality of the residuals was checked
265 for all LMMs performed.

266 **3. Results**

267 We obtained two datasets of movement and acceleration data; the 'daily attendance
268 dataset' contained data of 12,616 stork-days (median \pm SE; 162 ± 28 days per individual);
269 and the 'daily foraging strategy dataset' of 10,183 stork-days (136 ± 22 days per
270 individual, see detailed information in Table S1 and S2). During both the breeding and
271 non-breeding season foraging occurred in southern Portugal, but during the non-breeding
272 season white storks increased their foraging range towards southern Spain (Fig 1). All
273 individuals but one foraged both on landfill sites and outside of landfill sites. The

274 probability of attending landfill sites varied with season (Estimate = 0.63; SE = 0.07; $p <$
275 0.001); it was 60% during the non-breeding season and 44% during the breeding season.
276 Moreover, during the breeding season, the probability of attending a landfill was
277 constrained by the distance from the nest to the closest landfill site (Estimate = -0.16; SE
278 = 0.03; $p < 0.001$) in a linear way (quadratic term was non-significant; $\chi^2 = 2.882$; $p =$
279 0.09), with individuals nesting closer to a landfill site foraging there more often (Fig 2).

280 White storks travelled further during the breeding season to visit landfill sites (Estimate
281 \pm SE; 27.02 ± 0.98 km) than when foraging outside landfill areas (22.01 ± 0.79 km); while
282 during the non-breeding season, storks travelled less overall, and birds that used landfills
283 travelled shorter distances (13.11 ± 0.82 km) than those feeding outside of landfills (15.76
284 ± 0.80 , Table 1, Fig 3). The results for mean ODBA, relative foraging time and mean
285 foraging ODBA were similar. The three parameters were higher during the breeding
286 season in the days birds foraged outside landfill sites (mean ODBA 0.10 ± 0.01 G; relative
287 foraging time 0.37 ± 0.05 ; mean foraging ODBA 0.19 ± 0.01 G, Table 1, Fig 3), and
288 decreased during the non-breeding season and in days that birds foraged in landfill sites,
289 reaching the lowest values when both conditions occurred (mean ODBA 0.07 ± 0.01 G;
290 relative foraging time 0.26 ± 0.05 ; mean foraging ODBA 0.15 ± 0.01 G, Table 1, Fig 3).
291 Thus, mean ODBA, relative foraging time, and mean foraging ODBA decreased 34.6 %,
292 30.1 %, and 12.4 %, respectively, from the days that birds foraged outside landfills during
293 the breeding season to days that birds foraged in landfills during the non-breeding season.

294 The multigroup analysis for piecewise SEM indicated that mean ODBA increased when
295 white storks had to travel further, when they spent a higher proportion of time foraging,
296 and for storks that presented higher mean foraging ODBA (Table 2). The effect was
297 stronger when foraging occurred outside of landfills during the breeding season (Fig 4).
298 The relationship between relative foraging time and mean foraging ODBA differed

299 depending on the foraging site (Table 2). When foraging on outside of landfills, an
300 increase on mean foraging ODBA led to a significant increase in relative foraging time;
301 while on landfill sites, an increase on mean foraging ODBA led to slight decrease in
302 foraging time (Fig 5).

303 Finally, we did not find any relationship between relative foraging time and landfill
304 specialisation, and between mean foraging ODBA and landfill specialisation when birds
305 foraged on landfill waste (Table 3).

306 **4. Discussion**

307 This study unravels some of the mechanisms that determine landfill use in a generalist
308 and opportunistic species. In Iberia, expansion and population growth of white storks is
309 widely attributed to their high adaptability and behavioural plasticity, which allow them
310 to efficiently use opportunities provided by anthropized environments such as landfills
311 (Catry et al., 2017; Elliott et al., 2020). Here, we show that foraging on landfill waste is
312 a time- and energy-efficient strategy for white storks compared to foraging on natural
313 habitats. Remarkably, although storks had to travel further to exploit this resource during
314 the breeding period, they spent overall less energy, in terms of locomotion, than when
315 foraging on natural prey (Fig 3). Our results indicate that this could be mediated by a
316 reduction in foraging time and an increase in foraging efficiency (i.e. a decrease in mean
317 foraging ODBA) while exploiting landfill waste (Fig 3). These findings contrasts with
318 previous evidence from herring gulls (*Larus argentatus*), which increase energy
319 expenditure 34 % when foraging on PAFS compared to natural prey (van Donk et al.,
320 2019). These differences between species are likely to be ascribed to their flight mode;
321 while herring gulls tend to use flapping flight, which is energetically costly, white storks
322 soar, which allow them to fly longer distances to reach landfill sites at a cheaper energetic

323 cost, especially at certain times of day when flight conditions (e.g. uplift) are favorable
324 (Duriez et al., 2014).

325 Nevertheless, we find that white storks experience constraints during the breeding season
326 and reduce landfill attendance to 44% of days compared to 60% in the non-breeding
327 period. This reduction is mostly a function of distance to nest location: white storks
328 nesting further to landfill sites visit them less often than birds nesting closer (Fig 2). These
329 differences in foraging site preference indicate that the energetic compensation of longer
330 flights, with increased foraging efficiency and decreased foraging time at landfill sites,
331 compensate for longer flights only up to a certain limit, after which it is no longer
332 beneficial to travel further to forage at landfills.

333 The hindrance imposed by reproduction is reflected on all the parameters that we studied;
334 storks increased distance travelled, energy expenditure, foraging time, and mean foraging
335 ODBA during the breeding season (Fig 3). These parameters are likely to increase in
336 response to the spatial restrictions that impose returning to the nest frequently and the
337 higher energetic demands derived from raising chicks (Johst et al., 2001). On the contrary,
338 during the non-breeding period, storks are not constrained by nest location and often roost
339 closer to landfill sites, thus reducing daily distances travelled. Moreover, foraging just to
340 meet their own energy demands could be leading to a reduction in foraging time and daily
341 energy expenditure.

342 Our results show that daily energy expenditure is highly influenced by distance travelled,
343 foraging time and foraging efficiency, with higher energy expenditure with increasing
344 distance travelled and foraging time and decreasing foraging efficiency (Fig 4). The
345 slopes of these relationships were maintained when foraging at landfill sites and non-
346 landfill sites, both during the breeding and non-breeding season. However, in all cases
347 the intercept was lower when foraging at landfill sites, indicating that there could be other

348 unaccounted factors decreasing the daily energy expenditure on days that birds forage at
349 landfill sites, such as a reduction on flight energetics, a change of flight mode (soaring vs
350 flapping), or a decrease on time spent preening.

351 Interestingly, we found that the relationship between relative foraging time and energy
352 efficiency differed depending on the foraging site. At natural sites, birds with high levels
353 of energy expenditure (low energy efficiency) during foraging, increased the time spent
354 foraging. On the contrary, at landfill sites, birds that were less energy-efficient during
355 foraging spent less time foraging, this effect was stronger during the non-breeding season
356 (Fig 5). This divergence is likely to stem from the different characteristics of foraging
357 on landfill waste versus natural prey. Individuals foraging at landfill sites, where large
358 numbers of white storks congregate, are likely to experience easy access to food but also
359 density dependence effects, such as direct competition from conspecifics (Oro et al.,
360 2013). This could lead to an increase of energy expenditure during foraging due to
361 antagonistic interactions with other individuals, but also to a reduction in foraging time,
362 since birds could be displaced from optimal foraging sites (Burger, 1981). On natural
363 areas, however, white storks hunt large invertebrates and small vertebrates and do not
364 aggregate in such large numbers (Elliott et al., 2020). Thus, an increase of mean foraging
365 ODBA is likely to be linked to active hunting, which could then favour an increase in
366 relative foraging time.

367 Surprisingly, we did not find any evidence that landfill specialists experienced any
368 competitive advantage during landfill exploitation in terms of foraging time and energy
369 efficiency. Thus, our results suggest that landfill use is mainly driven by distance from
370 the nest to the nearest landfill site, rather than by the ability of individuals to exploit this
371 resource. Therefore, it is likely that white storks compete for nests located in close
372 proximity to landfill sites (Itonaga et al., 2011; Janiszewski et al., 2015).

373 In this study we have quantified one of the aspects required to define optimal foraging
374 strategies: time allocation and energy expenditure related to locomotion. However, we
375 could not quantify energy intake, a key aspect for energetic balances. Natural prey is
376 patchily distributed, and energetic values can be variable. For example, the energetic
377 value of red-swamp crayfish (*Procambarus clarkii*), an invasive alien species that occurs
378 in rice fields and is widely consumed by white storks (Ferreira et al., 2019; Negro et al.,
379 2000; Tablado et al., 2010), is ~ 3 kJ/g (Elvira et al., 1996). On the other hand, landfill
380 waste is abundant, easily accessible and predictable in space and time, and highly
381 energetic (up to 10–25 kJ/g (van Donk et al., 2019, 2017)). Recent evidence shows that
382 white stork nestlings in Iberia whose parents exploit landfill sites present better body
383 condition and nutritional status than nestlings whose parents feed on natural resources,
384 which complements our findings (Pineda-Pampliega et al., 2021).

385 However, landfill waste exploitation can also have associated costs, such as higher
386 exposure to heavy metals, poisons, pathogens, and plastics (Ahlstrom et al., 2018; de la
387 Casa-Resino et al., 2014; Plaza and Lambertucci, 2017; Tongue et al., 2019). For
388 example, a recent study of white storks in Iberia found that nestlings from birds foraging
389 on landfill waste had a higher presence of antibiotic-resistant *Escherichia coli* than
390 nestlings from birds feeding on natural resources (Pineda-Pampliega et al., 2021); while
391 another found that nestlings from white storks near landfill sites presented higher blood
392 levels of lead, mercury, selenium, iron, zinc, and arsenic, which can be toxic at high
393 concentrations (de la Casa-Resino et al., 2014). Moreover, the occurrence of plastics in
394 landfills can lead to their ingestion, which can hurt the digestive organs of the birds (Peris,
395 2003).

396 Overall, our study shows that white storks nesting in the proximity of landfill sites and
397 foraging there frequently save energy and time, thus exploiting landfill waste can be an

398 advantageous strategy. Increased breeding success and population growth driven by
399 waste exploitation has been described for other bird species (Bialas et al., 2020; Djerdali
400 et al., 2008; Gilbert, 2015; Tauler-Ametller et al., 2017), thus a potential reduction of
401 landfill waste could have dramatic impacts on the population, since alternative food
402 resources might not be available to meet the energetic requirements of such large number
403 of individuals (but see (Katzenberger et al., 2019)). In light of the European Union
404 directives that will lead to a substantial reduction of landfill waste in the next few years,
405 our results indicate that species that heavily rely on landfill waste will be severely affected
406 and will have to radically change their behaviour, for example by increasing foraging
407 time and energy expenditure while foraging, affecting energetic and time balances.
408 However, it is currently unknown whether alternative foraging resources will be able to
409 sustain these populations in the future.

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426 **Authors' contributions**

427 ASR conceived the study, carried out the statistical analyses, and drafted the manuscript;
428 AMAF and FM coordinated the study, and critically revised the manuscript; MA
429 processed data, collected field and behavioural data, and critically revised the manuscript;
430 BHM collected field and behavioural data, and critically revised the manuscript; IC
431 conceived the study, coordinated the study and helped draft the manuscript. All authors
432 gave final approval for publication.

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(a) Distance travelled	Estimate	SE	<i>t</i>	<i>p</i>
Fixed effects				
Intercept	20.32	1.50	13.58	<0.001
Season	-13.90	0.81	-17.15	<0.001
Foraging site	-5.001	0.71	-7.07	<0.001
Season:Site	7.65	0.87	8.83	<0.001
GPS positions	0.23	0.04	6.05	<0.001
Random effects				
Bird ID Tag type	4.98			
Tag type	0.01			
AR(1)	0.50			
Residual variance	15.50			
Marginal R ²	0.08			
Conditional R ²	0.17			
(b) Mean ODBA				
Fixed effects				
Intercept	0.09	0.01	10.36	<0.001
Season	-0.02	0.00	-16.24	<0.001
Foraging site	0.01	0.00	12.57	<0.001
Season:Site	0.00	0.00	-1.78	0.0755
Random effects				
Bird ID Tag type	0.01			
Tag type	0.02			
AR(1)	0.54			
Residual variance	0.03			
Marginal R ²	0.17			
Conditional R ²	0.46			
(c) Relative foraging time				
Fixed effects				
Intercept	-0.83	0.06	14.08	<0.001
Season	-0.20	0.03	-5.82	<0.001
Foraging site	0.33	0.03	11.91	<0.001
Season:Site	-0.03	0.03	-0.90	0.3694
Random effects				
Bird ID Tag type	0.25			
Tag type	0.05			
AR(1)	0.59			
Residual variance	1.58			
Marginal R ²	0.01			
Conditional R ²	0.03			
(d) Foraging ODBA				
Fixed effects				
Intercept	0.17	0.01	15.46	<0.001
Season	-0.02	0.002	-9.59	<0.001
Foraging site	0.02	0.002	15.42	<0.001
Season:Site	-0.02	0.002	-8.71	<0.001
Random effects				

Bird ID Tag type	0.01			
Tag type	0.02			
AR(1)	0.32			
Residual variance	0.03			
Marginal R ²	0.15			
Conditional R ²	0.44			

611 Table 1: Estimates from LMMs and GLMM explaining (a) total distance travelled, (b)
612 mean overall dynamic body acceleration (mean ODBA), (c) relative foraging time, and
613 (d) mean foraging ODBA. Reference level for season is 'breeding', and for foraging site
614 is 'landfill'.

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(a) Model-wide Interactions	Predictor	Estimate	SE	<i>p</i>
Response				
Mean ODBA	Distance travelled			<0.001
Mean ODBA	Relative foraging time			<0.001
Mean ODBA	Foraging ODBA			<0.001
Relative foraging time	Foraging ODBA			<0.001
(b) Breeding, non-landfill				
Response				
Mean ODBA	Distance travelled	0.0005	0.0000	<0.001
Mean ODBA	Relative foraging time	0.1049	0.0020	<0.001
Mean ODBA	Foraging ODBA	0.3150	0.0091	<0.001
Relative foraging time	Foraging ODBA	0.8266	0.2827	0.0035
(c) Non-Breeding, non-landfill				
Response				
Mean ODBA	Distance travelled	0.0005	0.0000	<0.001
Mean ODBA	Relative foraging time	0.1016	0.0020	<0.001
Mean ODBA	Foraging ODBA	0.2536	0.0091	<0.001
Relative foraging time	Foraging ODBA	1.6589	0.3157	<0.001
(d) Breeding, landfill				
Response				
Mean ODBA	Distance travelled	0.0006	0.0000	<0.001
Mean ODBA	Relative foraging time	0.0985	0.0068	<0.001
Mean ODBA	Foraging ODBA	0.2218	0.0184	<0.001
Relative foraging time	Foraging ODBA	-0.0627	0.5119	0.9026
(e) Non-Breeding, landfill				
Response				
Mean ODBA	Distance travelled	0.0004	0.0000	<0.001
Mean ODBA	Relative foraging time	0.0691	0.0033	<0.001
Mean ODBA	Foraging ODBA	0.1696	0.0085	<0.001
Relative foraging time	Foraging ODBA	-0.7308	0.2773	0.0085

634 Table 2: Estimates from structural equation models (SEM) explaining the relationship
635 among total distance travelled, mean overall dynamic body acceleration (ODBA), relative
636 foraging time, and mean foraging ODBA. Estimates for (a) the whole model, (b) breeding
637 period outside landfills, (c) non-breeding period outside landfills, (d) breeding period in
638 landfills, and (e) non-breeding period in landfills.

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(a) Relative foraging time	Estimate	SE	<i>t</i>	<i>p</i>
Fixed effects				
Intercept	-1.21	0.11	-11.29	<0.001
Season	-0.03	0.09	-0.28	0.7795
Specialisation index	0.31	0.19	1.65	0.1062
Season:Specialisation index	-0.11	0.14	-0.79	0.4323
Random effects				
Bird ID Tag type	0.14			
Tag type	0			
AR(1)	0.25			
Residual variance	1.01			
(b) Foraging ODBA				
Fixed effects				
Intercept	0.16	0.01	11.22	<0.001
Season	-0.02	0.01	-2.51	0.0120
Specialisation index	0.01	0.01	0.42	0.6756
Season:Specialisation index	0.01	0.01	0.82	0.4142
Random effects				
Bird ID Tag type	0.01			
Tag type	0.02			
AR(1)	0.10			
Residual variance	0.03			

642 Table 3: Estimates from GLMM and LMM explaining (a) relative foraging time, and (b)
643 mean foraging overall dynamic body acceleration (ODBA), while foraging on landfills.
644 Reference level for season is ‘breeding’.

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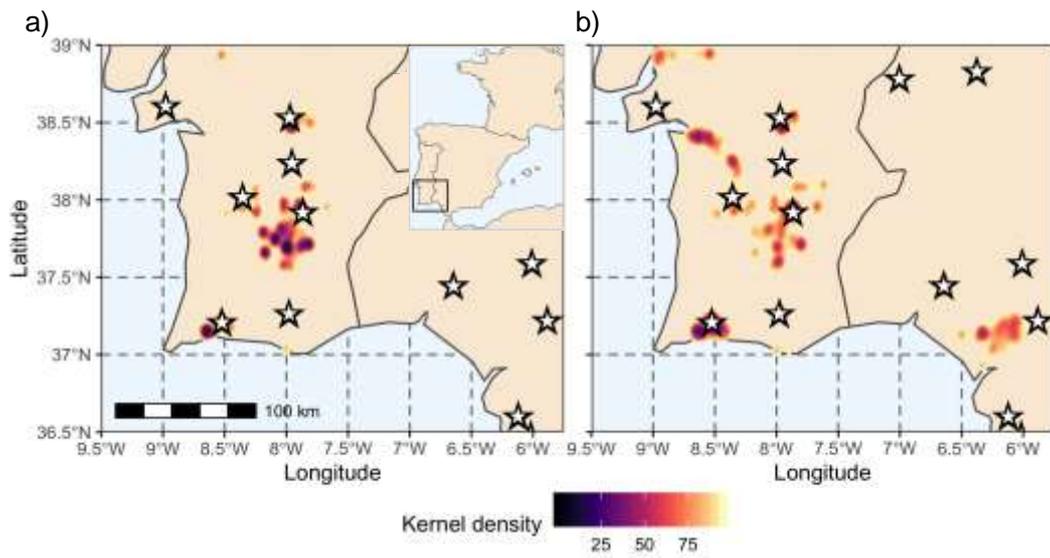
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657 Figure 1: Foraging areas of tracked white storks in southern Iberia (Portugal and Spain)

658 between 2016 and 2019 during a) the breeding and b) the non-breeding periods. Stars

659 indicate landfill sites.

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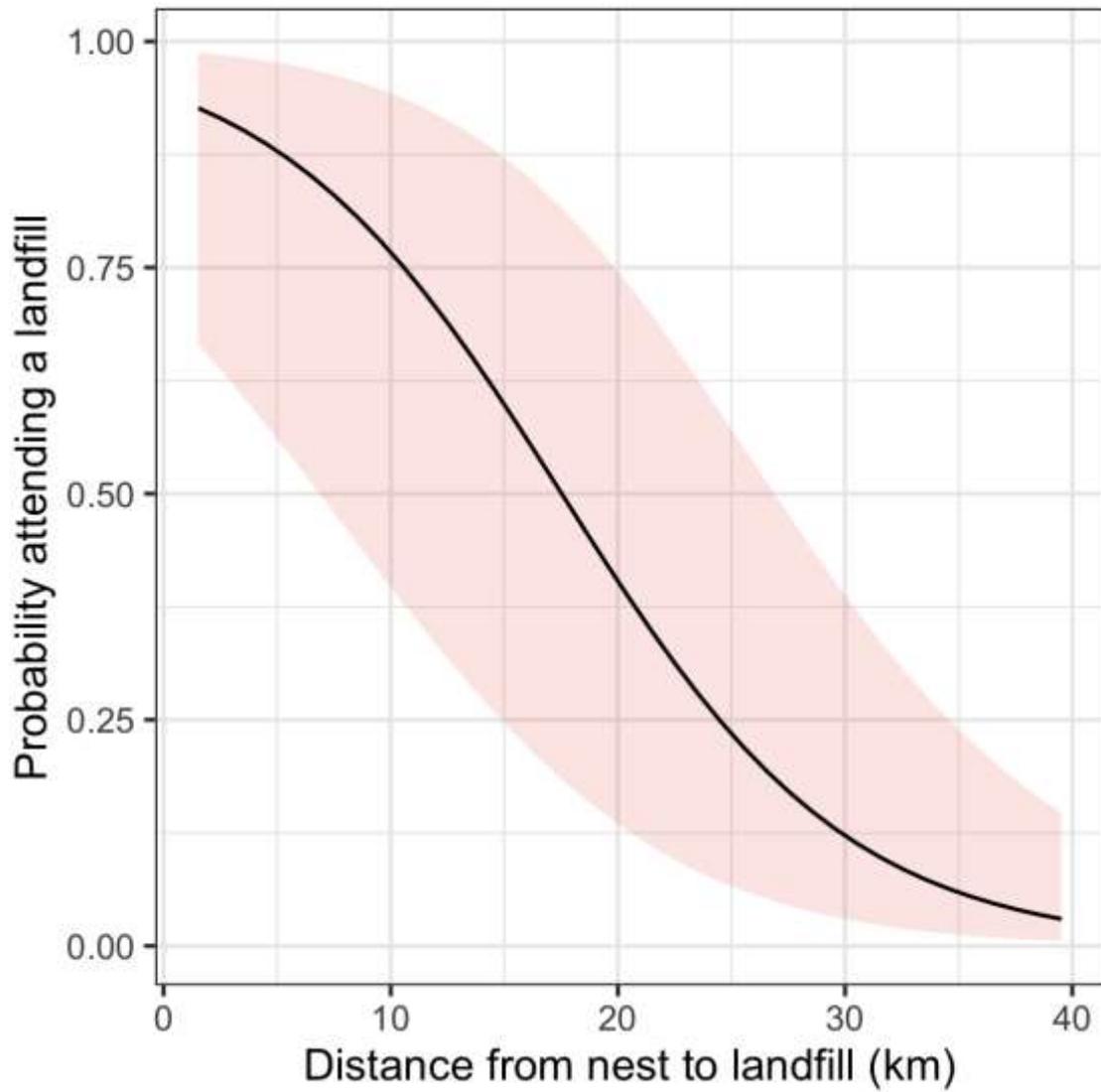
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675 Figure 2: Estimates from GLMMs explaining the probability of white storks attending a
676 landfill during the breeding season as a function of the distance from the nesting site to
677 the closest landfill.

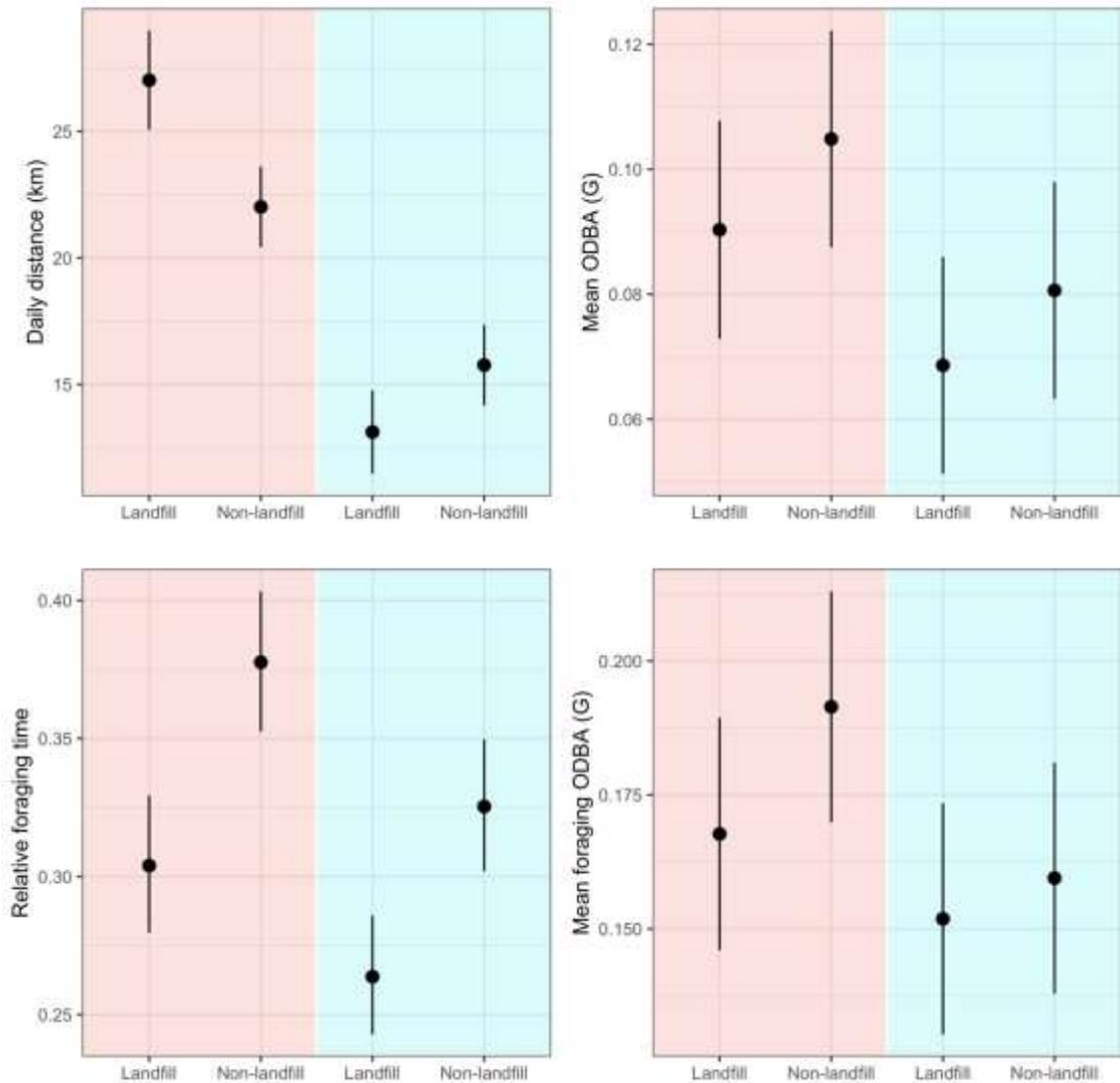
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684 Figure 3: Predicted values and 95 % confidence intervals from GLMMs explaining (a)
 685 daily distance travelled, (b) mean overall dynamic body acceleration (ODBA), (c) relative
 686 foraging time, and (d) mean foraging ODBA of foraging white storks, as a function of
 687 season (breeding or non-breeding) and foraging site (landfill or non-landfill).

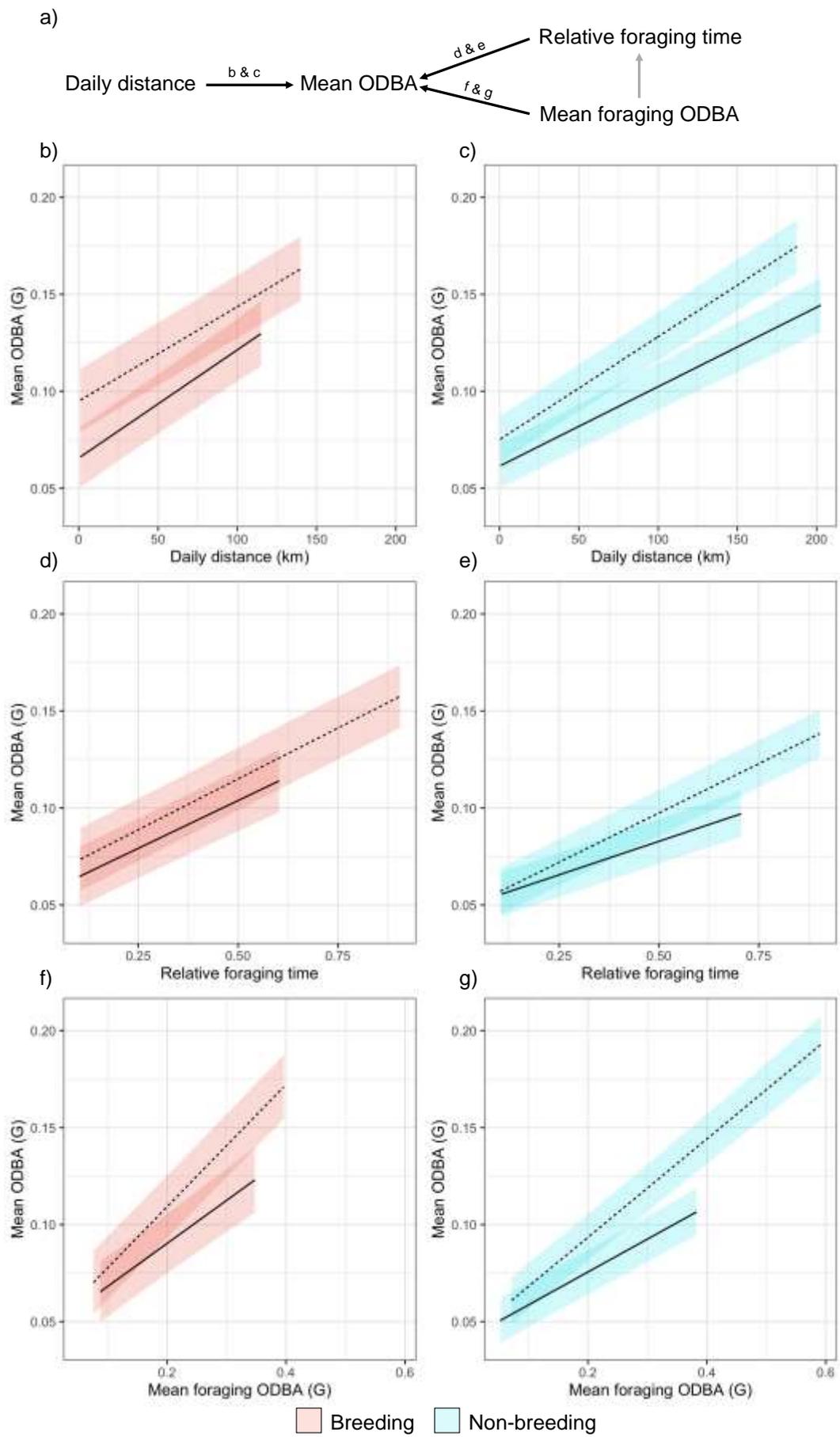
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694 Figure 4: (a) Graphical explanation of the structural equation model (SEM) showing the
695 relative contribution of white storks' foraging decisions into the mean overall dynamic
696 body acceleration (ODBA). Black arrows indicate relationships represented in the figure
697 and the grey arrow relationship not represented in the figure. Mean ODBA as a function
698 of daily distance travelled for (b) the breeding season and (c) the non-breeding season;
699 mean ODBA as a function of relative foraging time for (d) the breeding season and (e)
700 the non-breeding season; and mean ODBA as a function of mean foraging ODBA for (f)
701 the breeding season and (g) the non-breeding season. Solid line on landfill sites and
702 dashed line on non-landfill sites.

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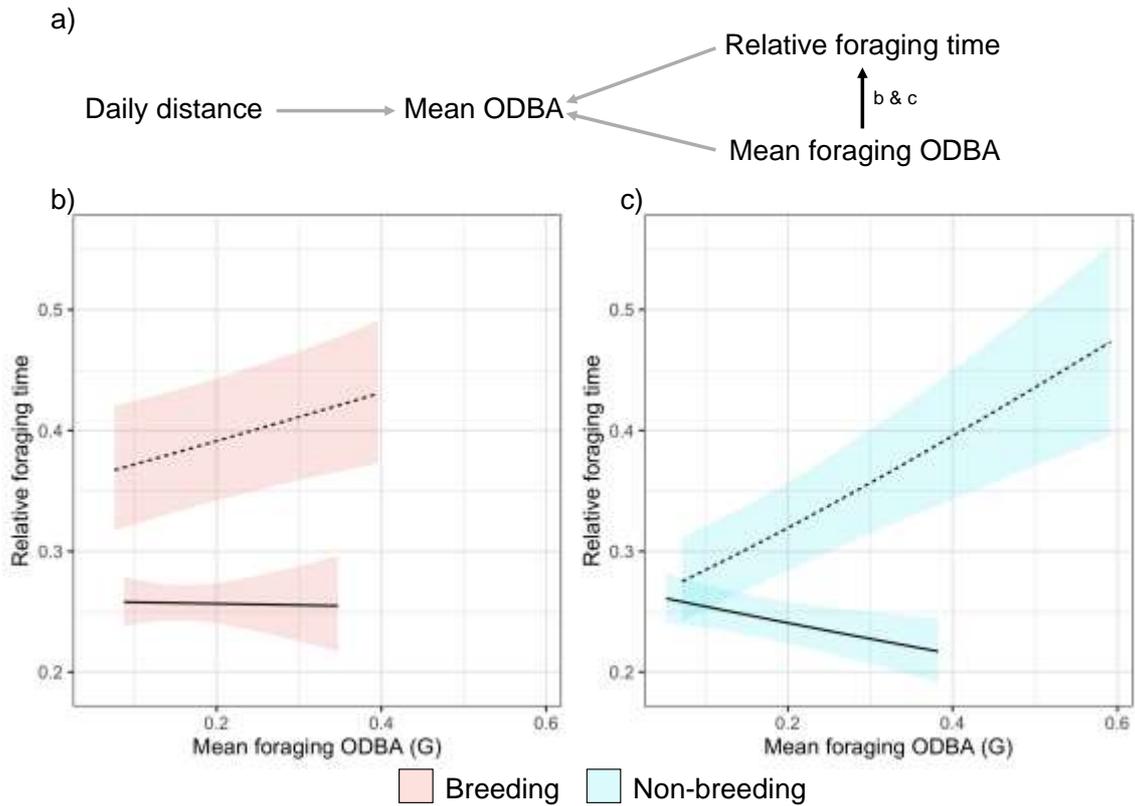
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720 Figure 5: (a) Graphical explanation of the structural equation model (SEM) showing the
 721 relative contribution of white storks' foraging decisions into the mean overall dynamic
 722 body acceleration (ODBA). Black arrows indicate relationships represented in the figure
 723 and the grey arrow relationship not represented in the figure. Relative foraging time as a
 724 function of mean foraging ODBA for (b) the breeding season and (c) the non-breeding
 725 season. Solid line on landfill sites and dashed line on non-landfill sites.

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