

1 **Flying the extra mile pays-off: foraging on**  
2 **anthropogenic waste as a time and energy-saving**  
3 **strategy in a generalist bird**

4 Andrea Soriano-Redondo<sup>1,2\*</sup>, Aldina M.A. Franco<sup>3</sup>, Marta Acácio<sup>3</sup>, Bruno  
5 Herlander Martins<sup>1,2,3</sup>, Francisco Moreira<sup>1,4</sup>, Inês Catry<sup>1,2,4</sup>

6  
7 <sup>1</sup> CIBIO/InBIO, Centro de Investigação em Biodiversidade e Recursos Genéticos, Laboratório  
8 Associado, Universidade do Porto, Campus Agrário de Vairão, Rua Padre Armando Quintas,  
9 4485–601 Vairão, Portugal

10 <sup>2</sup> CIBIO/InBIO, Centro de Investigação em Biodiversidade e Recursos Genéticos, Laboratório  
11 Associado, Instituto Superior de Agronomia, Universidade de Lisboa, 1349-017 Lisbon, Portugal

12 <sup>3</sup> School of Environmental Sciences, University of East Anglia, NR4 7TJ Norwich, UK

13 <sup>4</sup> REN Biodiversity Chair, CIBIO/InBIO-UP, Centro de Investigação em Biodiversidade e  
14 Recursos Genéticos, Universidade do Porto, Campus Agrário de Vairão, Rua Padre Armando  
15 Quintas, 4485–601 Vairão, Portugal

16 \* Corresponding author: Andrea Soriano-Redondo [asorianoredondo@outlook.com](mailto:asorianoredondo@outlook.com)

17  
18 **Abstract**

19 Food waste disposal represents a major global source of predictable anthropogenic food  
20 subsidies and is exploited by many organisms. However, the energetic cost-benefits of  
21 foraging on these food subsidies have remained largely unexplored. Here we investigate  
22 the year-round foraging decisions of resident white storks, *Ciconia ciconia*, in Iberia and  
23 assess the energetic and time cost-benefits of foraging on both landfill waste and natural  
24 food sources. To do so, we use GPS and acceleration data from 55 individuals tagged in  
25 southern Portugal between 2016 and 2019. We find that the probability of attending

26 landfill sites was 60% during the non-breeding season and 44% during the breeding  
27 season. Moreover, foraging on landfill waste is a time- and energy-saving strategy;  
28 although birds had to travel 20% further to exploit this resource during the breeding  
29 period, they spent overall 10% less energy than when foraging on natural prey. We show  
30 that this relationship could be mediated by a reduction in foraging time and an increase  
31 in foraging efficiency while exploiting landfill waste. Surprisingly, we did not find any  
32 evidence that landfill specialists experienced any competitive advantage during landfill  
33 exploitation over birds that visit landfills occasionally. These insights are key to predict  
34 how species that rely on landfills can be affected by waste reduction initiatives planned  
35 by the European Union, and implement the necessary management strategies.

36

37 **Keywords:** landfill closure; landfill waste; foraging ecology; predictable anthropogenic  
38 food subsidies; white storks.

39

## 40 **1. Introduction**

41 Humans produce enormous quantities of food waste; estimates suggest that 30-40% of all  
42 food produced is wasted (Parfitt et al., 2010) and deposited in locations where it can be  
43 accessible to wildlife. These predictable anthropogenic food subsidies (PAFS), in the  
44 form of organic waste on landfills, fisheries discards or crop residuals, generate impacts  
45 on animal populations at multiple scales, from the individual to the ecosystem (Oro et al.,  
46 2013). Individuals from numerous animal species have modified their movements,  
47 activity, geographical range, and home range size in response to PAFS (Gilbert et al.,  
48 2016; López-López et al., 2014; Newsome et al., 2015; Patrick et al., 2015; van Donk et  
49 al., 2019); which can affect survival and reproduction rates, and subsequently the  
50 demography of these populations (Plaza and Lambertucci, 2017).

51 Landfill sites, in particular, can potentially sustain high densities of scavenging  
52 individuals (Oro et al., 2013). The spatial and temporal predictability, accessibility, and  
53 nutritional value of anthropogenic food waste can provide abundant food resources for  
54 wildlife, and has been shown to be responsible for the demographic explosion of  
55 generalist animals (such as foxes, rats and gulls), but also for the sustenance of some  
56 endangered species (Plaza and Lambertucci, 2017). Landfill use has been associated with  
57 increased body mass in black vultures (*Coragyps atratus*) (Plaza and Lambertucci, 2018),  
58 kelp gull (*Larus dominicanus*) nestlings (Lenzi et al., 2019), and grizzly bears (*Ursus*  
59 *arctos horribilis*) (Blanchard, 1987). Landfill use has also been linked to higher  
60 reproduction performance; for example, white storks (*Ciconia ciconia*) had larger clutch  
61 sizes (Djerdali et al., 2008), glaucous gulls (*Larus hyperboreus*) experienced higher  
62 fledging rate (Weiser and Powell, 2010), and laughing gull (*Leucophaeus atricilla*) chick  
63 growth and survival was enhanced (Dosch, 1997), compared to conspecifics foraging in  
64 more natural landscapes.

65 According to the optimal foraging theory animals choose to forage in sites that maximize  
66 energy intake while minimizing energy and time expenditure (Stephens and Krebs, 1986;  
67 Ydenberg et al., 1994). Landfill waste offers large quantities of high-energy food  
68 (Patenaude-Monette et al., 2014; van Donk et al., 2019) but is localised in space; hence  
69 trade-offs exist between time spent in landfills sites and energetic costs of travelling,  
70 especially for central-place foragers during the breeding season (Gilbert et al., 2016).  
71 Foraging in landfill sites is likely to exacerbate inter and intraspecific competition  
72 (through agonistic interactions and food-robbing), which increases energetic costs, since  
73 they attract large numbers of individuals and at high densities (Oro et al., 2013). On the  
74 other hand, on natural heterogeneous landscapes, prey is often patchily distributed and  
75 intraspecific competition is likely to be low since individuals tend to forage in smaller

76 groups (Catry et al., 2017). These dynamics could lead to the emergence of individual  
77 foraging strategies and specialisation, with less competitive individuals avoiding landfill  
78 sites. Despite many studies focusing on the fitness benefits for individuals exploiting  
79 landfill waste, the energetic cost-benefits of landfill use have not yet been fully quantified.  
80 Use of landfill resources, together with the global increase of temperature due to climate  
81 change, may have facilitated the establishment of non-migratory white stork populations  
82 in Iberia (Catry et al., 2017). Resident individuals heavily rely on food waste disposal  
83 sites for foraging and no longer complete their annual migrations to and from their sub-  
84 Saharan wintering grounds. In two decades the number of resident white storks in  
85 Portugal has increased from 1,187 individuals (18% of the breeding population) in 1995  
86 to 14,434 (62% of the breeding population) in 2015 (Catry et al., 2017). White storks'  
87 use of landfill resources in Iberia has been investigated, and evidence shows that storks  
88 nesting close to landfill sites heavily relied on them (Gilbert et al., 2016). Moreover,  
89 breeding success for these individuals was higher than for individuals nesting further  
90 away (Gilbert, 2015).

91 Whilst this increasing number of white storks in Iberia is widely attributed to their high  
92 adaptability and behavioral plasticity, new EU directives (1999/31/UE and, more  
93 recently, 2018/850/UE), regulating waste disposal, can greatly revert this trend. These  
94 directives have established new circular economy targets aiming to reduce municipal  
95 waste landfilled to 10% in the next decade. Recent evidence shows that landfill closure  
96 can cause declines on survival, body mass, egg volume and/or clutch size in several gull  
97 species (Payo-Payo et al., 2015; Pons and Migot, 1995; Steigerwald et al., 2015) (but see  
98 (Katzenberger et al., 2019)). Therefore, it is crucial to understand foraging decisions, and  
99 how animals search for and exploit landfill waste, as opposed to natural prey, to predict

100 how animal populations might respond when food waste is removed, and to develop  
101 appropriate conservation and management strategies.

102 Here we investigate (i) the foraging decisions of resident white storks feeding on PAFS  
103 and natural food sources in Iberia; (ii) the energetic and time cost-benefits of these  
104 foraging decisions during the breeding and non-breeding seasons; and (iii) whether birds  
105 highly specialised on landfills gain a competitive advantage while exploiting PAFS, by  
106 increasing their foraging efficiency and decreasing the foraging time necessary to meet  
107 their energetic requirements, over birds that only visit landfills occasionally. These  
108 insights will be key to understand the drivers of landfill use and to predict how storks —  
109 and to some extent other birds found regularly in large numbers at landfills (e.g. egrets,  
110 herons, gulls and some raptors) — may be affected by reductions in PAFS. This  
111 understanding is key for a species that has substantially increased in numbers in the last  
112 decades due to the exploitation of landfills (Catry et al., 2017), and for which density  
113 dependent effects of food depletion are associated with high nestling mortality (Denac,  
114 2006; Zurell et al., 2015).

115 In this study we examine 4 years of GPS and tri-axial acceleration data from resident  
116 white storks in Iberia and determine their behaviour, energy expenditure, foraging  
117 decisions, and landfill use. We use structural equation models (SEM) (Lefcheck et al.,  
118 2016; Lefcheck, 2016) to understand the relationships among foraging movements,  
119 energy expenditure and behaviour. We hypothesize that the choice of food resources  
120 (anthropogenic waste or natural prey) will lead to different foraging strategies, for  
121 example at landfills sites birds will reduce the time devoted to foraging and increase  
122 energy efficiency given the spatiotemporal predictability of organic waste deposition.  
123 Moreover, we predict that foraging strategies might differ between the breeding and non-  
124 breeding season, as breeding birds will have to return to the nest frequently, so their

125 foraging range will be reduced, and landfill sites might then be outside the range. In this  
126 context, fundamental trade-offs between energetic and nutritional requirements in  
127 parental provisioning strategies are expected (Wright et al., 1998). Finally, we predict  
128 that landfill specialists will have competitive advantage over birds that only use landfills  
129 occasionally.

## 130 **2. Material and methods**

### 131 2.1 Bird capture and GPS tracking

132 Our dataset included GPS data with tri-axial acceleration from 55 resident adult white  
133 storks, tagged between 2016 and 2019 in southern Portugal. Resident individuals  
134 overwintered in the Iberian Peninsula and did not cross the Strait of Gibraltar. Storks were  
135 tagged with ‘Flyway 50’ GPS/GSM loggers from Movetech Telemetry (4 different  
136 models varying slightly in weight) and ‘Ornitrack-50’ GPS/GSM loggers from Ornitela.  
137 Adult birds were caught at multiple landfill sites using nylon leg nooses, and at several  
138 breeding colonies using a remotely activated clap net at the nests. Birds were measured  
139 and ringed, and the devices were mounted on the back of the birds as backpacks with a  
140 Teflon harness. The tag and harness together weighted 60-90 g, which represented 1.5-  
141 3.7% of a given bird's body mass at the time of tagging. Most birds were resighted in the  
142 days following tag deployment and no abnormal behaviour or adverse effects due to  
143 tagging were observed. The procedure was approved by the Instituto da Conservação da  
144 Natureza e Floresta (Portugal). The tags were programmed to record 9 consecutive GPS  
145 positions at 1 Hz every 20 minutes, and concurrently a 9 s tri-axial acceleration burst at  
146 1 Hz. We kept the first GPS position recorded, thus location and acceleration matched.  
147 GPS data was visually examined to detect potential outliers, which were subsequently  
148 removed, together with the associated acceleration burst. We identified 75 nest locations  
149 for the tagged birds across the years after visually inspecting the GPS tracks and visiting

150 the sites. The nests were situated between 1.5 and 40.2 km away from the closest landfill  
151 site (mean = 17.3 km).

## 152 2.2 Data selection and processing

153 From each acceleration burst we derived two metrics, ODBA (overall dynamic body  
154 acceleration,  $1\text{ G} = 9.8\text{ m/s}^2$ ), a proxy of energy expenditure invested in locomotion, and  
155 behaviour (Gleiss et al., 2011; Shepard et al., 2008b). ODBA was obtained from tri-axial  
156 acceleration bursts by subtracting the smoothing of total acceleration, using a running-  
157 mean of 4 s, from the total acceleration, as recommended in (Gleiss et al., 2011; Shepard  
158 et al., 2008a, 2008b). To determine the bird behaviour at each burst, we used the tri-axial  
159 acceleration data to train random forest machine-learning algorithms (R package  
160 ‘randomForest’ ver. 4.6 (Liaw and Wiener, 2002)). Movetech Telemetry tags and  
161 Ornitela tags differ in their sensor characteristics (e.g. position of the sensor and axes  
162 orientation) hence we created device specific algorithms. We characterised four  
163 behaviours: foraging, resting (including preening), flapping, and soaring flight (including  
164 gliding and orographic and thermal soaring). To train the algorithm we manually labelled  
165 250 tri-axial acceleration bursts for each behaviour and tag type; 70% of the data was  
166 used for training the algorithm and 30% for testing it. To label the training data we  
167 compiled information from several sources: (i) 9 tags that were programmed to  
168 continuously record GPS and acceleration data which allowed a detailed understanding  
169 of the birds’ movements and behaviour, (ii) video recordings of captive white storks that  
170 were fitted with the tracking devices, and (iii) from the 9 consecutive GPS locations that  
171 allowed to infer birds’ speed and movement during the acceleration burst. The random  
172 forest model had 96% accuracy for Movetech Telemetry tags data and 97% accuracy for  
173 Ornitela tags data. For this study, flapping and soaring flight were aggregated and

174 considered as flying, as both behaviours occurred infrequently, representing only 5.7 %  
175 of locations.

176 Landfill sites were determined through visual inspection of satellite images, and in all  
177 cases, they were confirmed by visits or contact with local authorities. Each GPS location  
178 was classified as either inside or outside a landfill. Subsequently, using the behavioural  
179 and GPS data, we created daily foraging strategy metrics for each individual stork. We  
180 only included in the analyses days with more than 10 daylight GPS positions and  
181 acceleration bursts (median GPS positions per day was 29) and with more than 4 positions  
182 classified as ‘foraging’ to guarantee representative metrics.

183 We built two different datasets for the subsequent analyses. First, to assess individual and  
184 seasonal differences in attendance to landfill sites, we created a ‘daily attendance dataset’.  
185 In this dataset each day was classified as a ‘landfill attendance day’ when at least one  
186 foraging location occurred on a landfill in the corresponding 24-h period, or as a ‘non-  
187 landfill attendance day’ when no foraging locations occurred on landfills. Second, to  
188 understand the foraging movements, behaviour and energy expenditure of feeding on  
189 anthropogenic waste or natural prey we built a ‘daily foraging strategy dataset’ that only  
190 included days where birds displayed either one or the other strategy, thus removing days  
191 where birds foraged on both waste and natural prey. To do so, we created a variable,  
192 ‘foraging site’ with two levels: ‘landfill foraging day’, when 70% of the daily foraging  
193 positions occurred in landfill sites, and ‘non-landfill foraging day’, when 70% of the  
194 foraging positions for a given day occurred outside landfill sites. We chose this threshold  
195 as it ensured birds spent most of the day in one of the two areas, while it minimised the  
196 number of days that had to be removed from the dataset to less than 20%.

197 The activity and energetic trade-offs between foraging on landfills vs natural prey were  
198 explored using several metrics: (i) daily distance travelled (km), calculated as the summed

199 distance between consecutive locations for each day; (ii) daily relative flight time,  
200 calculated by dividing the number of burst classified as flying by the total of bursts  
201 obtained in a day; (iii) mean ODBA (G) as a proxy of daily energy expenditure (Gleiss  
202 et al., 2011), calculated as the mean ODBA of the accelerometer bursts obtained in a day;  
203 (iv) daily relative foraging time, calculated by dividing the number of burst classified as  
204 foraging by the total of burst in a day; and (v) mean foraging ODBA (G), as a proxy for  
205 foraging efficiency (lower mean foraging ODBA values indicate higher foraging  
206 efficiency), calculated as the daily mean ODBA for the foraging bursts. Daily distance  
207 travelled and relative flight time were highly correlated (Pearson correlation = 0.75),  
208 which indicates that the metrics were largely insensitive to the number of fixes; we only  
209 retained daily distance travelled in further analyses.

210 To account for the potential effect of season on the foraging site choice (landfill sites or  
211 non-landfill sites), we located the nesting site of each individual, and using the GPS data,  
212 we defined the beginning of the breeding season as the first three consecutive days that  
213 an individual occupied the nest; and the end of the breeding period as last three  
214 consecutive days the nest was occupied.

215 To explore if birds that visited landfill sites more often had a competitive advantage when  
216 exploiting this resource, we used the ‘daily attendance dataset’ to create a landfill  
217 specialisation index, from 0 (natural prey specialist; never visits landfill sites) to 1 (waste  
218 specialist; only forages on landfill sites), for each individual. We divided the number of  
219 days that the bird attended a landfill by the total numbers of days that the bird was tagged.

### 220 2.3 Statistical analysis

221 To understand seasonal variability in foraging site attendance we fitted a generalised  
222 linear mixed-effects model (GLMM) with foraging site attendance (landfill or non-  
223 landfill) as the response variable and season (breeding or non-breeding) as the

224 explanatory variable, using the *glmmPQL* function with a binomial structure (R package  
225 ‘MASS’ ver. 7.3 (Ripley et al., 2013)); the model included bird ID nested in tag type (5  
226 levels: 4 types of Movetech tags and 1 type of Ornitela tag) as random effects and an  
227 autocorrelation structure of order 1 to account for the potential temporal correlation  
228 between consecutive days. Subsequently, to understand the potential constraints of  
229 reproduction in the choice of foraging site, we used the data from the breeding season to  
230 fit a GLMM with a binomial structure, using the *glmer* function (R package ‘lme4’ ver.  
231 1.1-21 (Bates et al., 2015)). We considered foraging site attendance as the response  
232 variable, and included as explanatory variables, distance from the nest site to the closest  
233 landfill site, and its quadratic term, to account for potential non-linearity in the  
234 relationship; we also included bird ID nested in tag type as random effects. We used the  
235 R package ‘MASS’ when the inclusion of an autocorrelation structure was necessary as  
236 ‘lme4’ does not allow the implementation of GLMMs with that structure.

237 Using the ‘daily foraging strategy dataset’, we explored the effect of landfill use on time  
238 and energy budgets, by implementing 3 linear mixed-effects models (LMM) with daily  
239 distance travelled (km), mean ODBA (G), and mean foraging ODBA (G) as response  
240 variables and foraging site (landfill or non-landfill), season (breeding or non-breeding),  
241 and its interaction, as explanatory variables, using the *lme* function (R package ‘nlme’  
242 ver. 3.1 (Pinheiro et al., 2017)). The model with daily distance travelled as response  
243 variable, included an additional covariate, the number of GPS positions, to account for its  
244 potential effect on the response. We implemented a GLMM with the *glmmPQL*  
245 function for relative foraging time with a binomial structure that included foraging site,  
246 season, and its interaction as fixed effects as well. All four models included bird ID nested  
247 in tag type as random effects and an autocorrelation structure of order 1. Subsequently,  
248 we calculated for each model the coefficient of determination  $R^2$  (Nakagawa et al., 2017).

249 We explored the relative contribution of individuals' foraging decisions into the overall  
250 energy expenditure when using different foraging sites and during different seasons. To  
251 do so, we implemented multigroup analysis for piecewise SEM with foraging site and  
252 season as grouping variables, using the functions *psem* and *multigroup* (R package  
253 'piecewiseSEM' ver. 2.1 (Lefcheck et al., 2016; Lefcheck, 2016)). The SEM included  
254 two sub-models: 1. an LMM with mean ODBA as the response variable, and daily  
255 distance travelled, relative foraging time and mean foraging ODBA as explanatory  
256 variables; and 2. a GLMM with relative foraging time as the response variable, and mean  
257 foraging ODBA as explanatory variables. Both models included random effects and  
258 autocorrelation structure as described above. The global structure of the SEM model was  
259 well supported according to the global goodness-of-fit: Fisher's  $C = 0.098$  with  $p = 0.952$ .

260 To test the role of landfill specialisation on landfill exploitation we selected only days  
261 classified as landfill foraging days. We modelled a LMM with mean foraging ODBA and  
262 a GLMM with relative foraging time as the response variables, and season and  
263 specialisation index as explanatory variables. We used random effects and an  
264 autocorrelation structure as previously specified. Normality of the residuals was checked  
265 for all LMMs performed.

### 266 **3. Results**

267 We obtained two datasets of movement and acceleration data; the 'daily attendance  
268 dataset' contained data of 12,616 stork-days (median  $\pm$  SE;  $162 \pm 28$  days per individual);  
269 and the 'daily foraging strategy dataset' of 10,183 stork-days ( $136 \pm 22$  days per  
270 individual, see detailed information in Table S1 and S2). During both the breeding and  
271 non-breeding season foraging occurred in southern Portugal, but during the non-breeding  
272 season white storks increased their foraging range towards southern Spain (Fig 1). All  
273 individuals but one foraged both on landfill sites and outside of landfill sites. The

274 probability of attending landfill sites varied with season (Estimate = 0.63; SE = 0.07;  $p <$   
275 0.001); it was 60% during the non-breeding season and 44% during the breeding season.  
276 Moreover, during the breeding season, the probability of attending a landfill was  
277 constrained by the distance from the nest to the closest landfill site (Estimate = -0.16; SE  
278 = 0.03;  $p < 0.001$ ) in a linear way (quadratic term was non-significant;  $\chi^2 = 2.882$ ;  $p =$   
279 0.09), with individuals nesting closer to a landfill site foraging there more often (Fig 2).

280 White storks travelled further during the breeding season to visit landfill sites (Estimate  
281  $\pm$  SE;  $27.02 \pm 0.98$  km) than when foraging outside landfill areas ( $22.01 \pm 0.79$  km); while  
282 during the non-breeding season, storks travelled less overall, and birds that used landfills  
283 travelled shorter distances ( $13.11 \pm 0.82$  km) than those feeding outside of landfills ( $15.76$   
284  $\pm 0.80$ , Table 1, Fig 3). The results for mean ODBA, relative foraging time and mean  
285 foraging ODBA were similar. The three parameters were higher during the breeding  
286 season in the days birds foraged outside landfill sites (mean ODBA  $0.10 \pm 0.01$  G; relative  
287 foraging time  $0.37 \pm 0.05$ ; mean foraging ODBA  $0.19 \pm 0.01$  G, Table 1, Fig 3), and  
288 decreased during the non-breeding season and in days that birds foraged in landfill sites,  
289 reaching the lowest values when both conditions occurred (mean ODBA  $0.07 \pm 0.01$  G;  
290 relative foraging time  $0.26 \pm 0.05$ ; mean foraging ODBA  $0.15 \pm 0.01$  G, Table 1, Fig 3).  
291 Thus, mean ODBA, relative foraging time, and mean foraging ODBA decreased 34.6 %,  
292 30.1 %, and 12.4 %, respectively, from the days that birds foraged outside landfills during  
293 the breeding season to days that birds foraged in landfills during the non-breeding season.

294 The multigroup analysis for piecewise SEM indicated that mean ODBA increased when  
295 white storks had to travel further, when they spent a higher proportion of time foraging,  
296 and for storks that presented higher mean foraging ODBA (Table 2). The effect was  
297 stronger when foraging occurred outside of landfills during the breeding season (Fig 4).  
298 The relationship between relative foraging time and mean foraging ODBA differed

299 depending on the foraging site (Table 2). When foraging on outside of landfills, an  
300 increase on mean foraging ODBA led to a significant increase in relative foraging time;  
301 while on landfill sites, an increase on mean foraging ODBA led to slight decrease in  
302 foraging time (Fig 5).

303 Finally, we did not find any relationship between relative foraging time and landfill  
304 specialisation, and between mean foraging ODBA and landfill specialisation when birds  
305 foraged on landfill waste (Table 3).

#### 306 **4. Discussion**

307 This study unravels some of the mechanisms that determine landfill use in a generalist  
308 and opportunistic species. In Iberia, expansion and population growth of white storks is  
309 widely attributed to their high adaptability and behavioural plasticity, which allow them  
310 to efficiently use opportunities provided by anthropized environments such as landfills  
311 (Catry et al., 2017; Elliott et al., 2020). Here, we show that foraging on landfill waste is  
312 a time- and energy-efficient strategy for white storks compared to foraging on natural  
313 habitats. Remarkably, although storks had to travel further to exploit this resource during  
314 the breeding period, they spent overall less energy, in terms of locomotion, than when  
315 foraging on natural prey (Fig 3). Our results indicate that this could be mediated by a  
316 reduction in foraging time and an increase in foraging efficiency (i.e. a decrease in mean  
317 foraging ODBA) while exploiting landfill waste (Fig 3). These findings contrasts with  
318 previous evidence from herring gulls (*Larus argentatus*), which increase energy  
319 expenditure 34 % when foraging on PAFS compared to natural prey (van Donk et al.,  
320 2019). These differences between species are likely to be ascribed to their flight mode;  
321 while herring gulls tend to use flapping flight, which is energetically costly, white storks  
322 soar, which allow them to fly longer distances to reach landfill sites at a cheaper energetic

323 cost, especially at certain times of day when flight conditions (e.g. uplift) are favorable  
324 (Duriez et al., 2014).

325 Nevertheless, we find that white storks experience constraints during the breeding season  
326 and reduce landfill attendance to 44% of days compared to 60% in the non-breeding  
327 period. This reduction is mostly a function of distance to nest location: white storks  
328 nesting further to landfill sites visit them less often than birds nesting closer (Fig 2). These  
329 differences in foraging site preference indicate that the energetic compensation of longer  
330 flights, with increased foraging efficiency and decreased foraging time at landfill sites,  
331 compensate for longer flights only up to a certain limit, after which it is no longer  
332 beneficial to travel further to forage at landfills.

333 The hindrance imposed by reproduction is reflected on all the parameters that we studied;  
334 storks increased distance travelled, energy expenditure, foraging time, and mean foraging  
335 ODBA during the breeding season (Fig 3). These parameters are likely to increase in  
336 response to the spatial restrictions that impose returning to the nest frequently and the  
337 higher energetic demands derived from raising chicks (Johst et al., 2001). On the contrary,  
338 during the non-breeding period, storks are not constrained by nest location and often roost  
339 closer to landfill sites, thus reducing daily distances travelled. Moreover, foraging just to  
340 meet their own energy demands could be leading to a reduction in foraging time and daily  
341 energy expenditure.

342 Our results show that daily energy expenditure is highly influenced by distance travelled,  
343 foraging time and foraging efficiency, with higher energy expenditure with increasing  
344 distance travelled and foraging time and decreasing foraging efficiency (Fig 4). The  
345 slopes of these relationships were maintained when foraging at landfill sites and non-  
346 landfill sites, both during the breeding and non-breeding season. However, in all cases  
347 the intercept was lower when foraging at landfill sites, indicating that there could be other

348 unaccounted factors decreasing the daily energy expenditure on days that birds forage at  
349 landfill sites, such as a reduction on flight energetics, a change of flight mode (soaring vs  
350 flapping), or a decrease on time spent preening.

351 Interestingly, we found that the relationship between relative foraging time and energy  
352 efficiency differed depending on the foraging site. At natural sites, birds with high levels  
353 of energy expenditure (low energy efficiency) during foraging, increased the time spent  
354 foraging. On the contrary, at landfill sites, birds that were less energy-efficient during  
355 foraging spent less time foraging, this effect was stronger during the non-breeding season  
356 (Fig 5). This divergence is likely to stem from the different characteristics of foraging  
357 on landfill waste versus natural prey. Individuals foraging at landfill sites, where large  
358 numbers of white storks congregate, are likely to experience easy access to food but also  
359 density dependence effects, such as direct competition from conspecifics (Oro et al.,  
360 2013). This could lead to an increase of energy expenditure during foraging due to  
361 antagonistic interactions with other individuals, but also to a reduction in foraging time,  
362 since birds could be displaced from optimal foraging sites (Burger, 1981). On natural  
363 areas, however, white storks hunt large invertebrates and small vertebrates and do not  
364 aggregate in such large numbers (Elliott et al., 2020). Thus, an increase of mean foraging  
365 ODBA is likely to be linked to active hunting, which could then favour an increase in  
366 relative foraging time.

367 Surprisingly, we did not find any evidence that landfill specialists experienced any  
368 competitive advantage during landfill exploitation in terms of foraging time and energy  
369 efficiency. Thus, our results suggest that landfill use is mainly driven by distance from  
370 the nest to the nearest landfill site, rather than by the ability of individuals to exploit this  
371 resource. Therefore, it is likely that white storks compete for nests located in close  
372 proximity to landfill sites (Itonaga et al., 2011; Janiszewski et al., 2015).

373 In this study we have quantified one of the aspects required to define optimal foraging  
374 strategies: time allocation and energy expenditure related to locomotion. However, we  
375 could not quantify energy intake, a key aspect for energetic balances. Natural prey is  
376 patchily distributed, and energetic values can be variable. For example, the energetic  
377 value of red-swamp crayfish (*Procambarus clarkii*), an invasive alien species that occurs  
378 in rice fields and is widely consumed by white storks (Ferreira et al., 2019; Negro et al.,  
379 2000; Tablado et al., 2010), is ~ 3 kJ/g (Elvira et al., 1996). On the other hand, landfill  
380 waste is abundant, easily accessible and predictable in space and time, and highly  
381 energetic (up to 10–25 kJ/g (van Donk et al., 2019, 2017)). Recent evidence shows that  
382 white stork nestlings in Iberia whose parents exploit landfill sites present better body  
383 condition and nutritional status than nestlings whose parents feed on natural resources,  
384 which complements our findings (Pineda-Pampliega et al., 2021).

385 However, landfill waste exploitation can also have associated costs, such as higher  
386 exposure to heavy metals, poisons, pathogens, and plastics (Ahlstrom et al., 2018; de la  
387 Casa-Resino et al., 2014; Plaza and Lambertucci, 2017; Tongue et al., 2019). For  
388 example, a recent study of white storks in Iberia found that nestlings from birds foraging  
389 on landfill waste had a higher presence of antibiotic-resistant *Escherichia coli* than  
390 nestlings from birds feeding on natural resources (Pineda-Pampliega et al., 2021); while  
391 another found that nestlings from white storks near landfill sites presented higher blood  
392 levels of lead, mercury, selenium, iron, zinc, and arsenic, which can be toxic at high  
393 concentrations (de la Casa-Resino et al., 2014). Moreover, the occurrence of plastics in  
394 landfills can lead to their ingestion, which can hurt the digestive organs of the birds (Peris,  
395 2003).

396 Overall, our study shows that white storks nesting in the proximity of landfill sites and  
397 foraging there frequently save energy and time, thus exploiting landfill waste can be an

398 advantageous strategy. Increased breeding success and population growth driven by  
399 waste exploitation has been described for other bird species (Bialas et al., 2020; Djerdali  
400 et al., 2008; Gilbert, 2015; Tauler-Ametller et al., 2017), thus a potential reduction of  
401 landfill waste could have dramatic impacts on the population, since alternative food  
402 resources might not be available to meet the energetic requirements of such large number  
403 of individuals (but see (Katzenberger et al., 2019)). In light of the European Union  
404 directives that will lead to a substantial reduction of landfill waste in the next few years,  
405 our results indicate that species that heavily rely on landfill waste will be severely affected  
406 and will have to radically change their behaviour, for example by increasing foraging  
407 time and energy expenditure while foraging, affecting energetic and time balances.  
408 However, it is currently unknown whether alternative foraging resources will be able to  
409 sustain these populations in the future.

#### 410 **Acknowledgements**

411 We are grateful to João Paulo Silva and Phil Atkinson for assistance in programming the  
412 devices, and to Carlos Pacheco, and students and volunteers that help tagging and  
413 monitoring white storks over the years.

#### 414 **Funding**

415 This work was financed by the FEDER Funds through the Operational Competitiveness  
416 Factors Program — COMPETE and by National Funds through FCT – Foundation for  
417 Science and Technology within the scope of the project ‘POCI-01-0145-FEDER-  
418 028176’, by InBIO (UID/BIA/50027/2013 and POCI-01-0145-FEDER-006821) and  
419 FCT/MCTES. ASR, BHM and IC were supported by FCT [grant numbers PTDC/BIA-  
420 ECO/28176/2017, DL57/2016/CP1440/CT0023 and SFRH/BD/145323/2019]. AMAF  
421 was supported by the by the Natural Environment Research Council (NERC) [grant  
422 number NE/ K006312], Norwich Research Park Translational Fund, University of East

423 Anglia Innovation Funds and Earth and Life Systems Alliance funds. MA was supported  
424 by the NERC [grant number NE/N012070/1]. FM was supported by FCT [grant number  
425 IF/01053/2015] and the REN Biodiversity Chair.

#### 426 **Authors' contributions**

427 ASR conceived the study, carried out the statistical analyses, and drafted the manuscript;  
428 AMAF and FM coordinated the study, and critically revised the manuscript; MA  
429 processed data, collected field and behavioural data, and critically revised the manuscript;  
430 BHM collected field and behavioural data, and critically revised the manuscript; IC  
431 conceived the study, coordinated the study and helped draft the manuscript. All authors  
432 gave final approval for publication.

#### 433 **References**

434 Ahlstrom, C.A., Bonnedahl, J., Woksepp, H., Hernandez, J., Olsen, B., Ramey, A.M.,  
435 2018. Acquisition and dissemination of cephalosporin-resistant *E. coli* in migratory  
436 birds sampled at an Alaska landfill as inferred through genomic analysis. *Sci Rep* 8, 1–  
437 11. <https://doi.org/10.1038/s41598-018-25474-w>  
438 Bates, D., Mächler, M., Bolker, B., Walker, S., 2015. Fitting Linear Mixed-Effects  
439 Models Using lme4. *Journal of Statistical Software* 67, 1–48.  
440 <https://doi.org/10.18637/jss.v067.i01>  
441 Bialas, J.T., Dylewski, Ł., Tobolka, M., 2020. Determination of nest occupation and  
442 breeding effect of the white stork by human-mediated landscape in Western Poland.  
443 *Environ Sci Pollut Res* 27, 4148–4158. <https://doi.org/10.1007/s11356-019-06639-0>  
444 Blanchard, B.M., 1987. Size and Growth Patterns of the Yellowstone Grizzly Bear.  
445 *Bears: Their Biology and Management* 7, 99–107. <https://doi.org/10.2307/3872615>  
446 Burger, J., 1981. Feeding Competition between Laughing Gulls and Herring Gulls at a  
447 Sanitary Landfill. *Condor* 83, 328–335. <https://doi.org/10.2307/1367501>  
448 Catry, I., Encarnação, V., Pacheco, C., Catry, T., Tenreiro, P., da Silva, L.P., Moreira,  
449 F., 2017. Recent changes on migratory behaviour of the White stork (*Ciconia ciconia*)  
450 in Portugal: Towards the end of migration. *Airo* 24, 28–35.  
451 de la Casa-Resino, I., Hernández-Moreno, D., Castellano, A., Pérez-López, M., Soler,  
452 F., 2014. Breeding near a landfill may influence blood metals (Cd, Pb, Hg, Fe, Zn) and  
453 metalloids (Se, As) in white stork (*Ciconia ciconia*) nestlings. *Ecotoxicology* 23, 1377–  
454 1386. <https://doi.org/10.1007/s10646-014-1280-0>  
455 Denac, D., 2006. Intraspecific Exploitation Competition as Cause for Density  
456 Dependent Breeding Success in the White Stork. *cowa* 29, 391–394.  
457 [https://doi.org/10.1675/1524-4695\(2006\)29\[391:IECACF\]2.0.CO;2](https://doi.org/10.1675/1524-4695(2006)29[391:IECACF]2.0.CO;2)  
458 Djerdali, S., Tortosa, F.S., Hillstrom, L., Doumandji, S., 2008. Food supply and external  
459 cues limit the clutch size and hatchability in the White Stork *Ciconia ciconia*. *Acta*  
460 *Ornithologica* 43, 145–150. <https://doi.org/10.3161/000164508X395252>  
461 Dosch, J.J., 1997. Diet of nestling Laughing Gulls in Southern New Jersey. *Waterbirds*

462 20, 273–281.

463 Duriez, O., Kato, A., Tromp, C., Dell’Omo, G., Vyssotski, A.L., Sarrazin, F., Ropert-  
464 Coudert, Y., 2014. How Cheap Is Soaring Flight in Raptors? A Preliminary  
465 Investigation in Freely-Flying Vultures. PLOS ONE 9, e84887.  
466 <https://doi.org/10.1371/journal.pone.0084887>

467 Elliott, A., Garcia, E.F.J., Boesman, P.F.D., 2020. White Stork (*Ciconia ciconia*),  
468 version 1.0, in: Birds of the World. Cornell Lab of Ornithology, Ithaca, NY, USA.

469 Elvira, B., Gnicola, G., Almodovar, A., 1996. Pike and red swamp crayfish: a new case  
470 on predator-prey relationship between aliens in central Spain. Journal of Fish Biology  
471 48, 437–446. <https://doi.org/10.1111/j.1095-8649.1996.tb01438.x>

472 Ferreira, E., Grilo, F., Mendes, R., Lourenço, R., Santos, S., Petrucci-Fonseca, F., 2019.  
473 Diet of the White Stork (*Ciconia ciconia*) in a heterogeneous Mediterranean landscape:  
474 the importance of the invasive Red Swamp Crayfish (*Procambarus clarkii*). Airo 26,  
475 27–41.

476 Gilbert, N., 2015. Movement and foraging ecology of partially migrant birds in a  
477 changing world (Doctoral thesis). University of East Anglia.

478 Gilbert, N.I., Correia, R.A., Silva, J.P., Pacheco, C., Catry, I., Atkinson, P.W., Gill,  
479 J.A., Franco, A.M.A., 2016. Are white storks addicted to junk food? Impacts of landfill  
480 use on the movement and behaviour of resident white storks (*Ciconia ciconia*) from a  
481 partially migratory population. Movement Ecology 4, 7. [https://doi.org/10.1186/s40462-](https://doi.org/10.1186/s40462-016-0070-0)  
482 [016-0070-0](https://doi.org/10.1186/s40462-016-0070-0)

483 Gleiss, A.C., Wilson, R.P., Shepard, E.L.C., 2011. Making overall dynamic body  
484 acceleration work: on the theory of acceleration as a proxy for energy expenditure.  
485 Methods in Ecology and Evolution 2, 23–33. [https://doi.org/10.1111/j.2041-](https://doi.org/10.1111/j.2041-210X.2010.00057.x)  
486 [210X.2010.00057.x](https://doi.org/10.1111/j.2041-210X.2010.00057.x)

487 Itonaga, N., Köppen, U., Plath, M., Wallschläger, D., 2011. Declines in breeding site  
488 fidelity in an increasing population of White Storks *Ciconia ciconia*. Ibis 153, 636–639.  
489 <https://doi.org/10.1111/j.1474-919X.2011.01131.x>

490 Janiszewski, T., Minias, P., Wojciechowski, Z., 2015. Selective Forces Responsible for  
491 Transition to Nesting on Electricity Poles in the White Stork *Ciconia ciconia*. arde 103,  
492 39–50. <https://doi.org/10.5253/arde.v103i1.a4>

493 Johst, K., Brandl, R., Pfeifer, R., 2001. Foraging in a Patchy and Dynamic Landscape:  
494 Human Land Use and the White Stork. Ecological Applications 11, 60–69.  
495 [https://doi.org/10.1890/1051-0761\(2001\)011\[0060:FIAPAD\]2.0.CO;2](https://doi.org/10.1890/1051-0761(2001)011[0060:FIAPAD]2.0.CO;2)

496 Katzenberger, J., Tabur, E., Şen, B., İsfendiyaroğlu, S., Erkol, I.L., Opper, S., 2019. No  
497 short-term effect of closing a rubbish dump on reproductive parameters of an Egyptian  
498 Vulture population in Turkey. Bird Conservation International 29, 71–82.  
499 <https://doi.org/10.1017/S0959270917000326>

500 Lefcheck, J., Byrnes, J., Grace, J., 2016. Package ‘piecewiseSEM.’ R package version  
501 1.1.

502 Lefcheck, J.S., 2016. piecewiseSEM: Piecewise structural equation modelling in r for  
503 ecology, evolution, and systematics. Methods in Ecology and Evolution 7, 573–579.  
504 <https://doi.org/10.1111/2041-210X.12512>

505 Lenzi, J., González-Bergonzoni, I., Machín, E., Pijanowski, B., Flaherty, E., 2019. The  
506 impact of anthropogenic food subsidies on a generalist seabird during nestling growth.  
507 Science of The Total Environment 687, 546–553.  
508 <https://doi.org/10.1016/j.scitotenv.2019.05.485>

509 Liaw, A., Wiener, M., 2002. Classification and Regression by randomForest. R News 2,  
510 6.

511 López-López, P., García-Ripollés, C., Urios, V., 2014. Food predictability determines

512 space use of endangered vultures: implications for management of supplementary  
513 feeding. *Ecological Applications* 24, 938–949. <https://doi.org/10.1890/13-2000.1>  
514 Nakagawa, S., Johnson, P.C.D., Schielzeth, H., 2017. The coefficient of determination  
515 R<sup>2</sup> and intra-class correlation coefficient from generalized linear mixed-effects models  
516 revisited and expanded. *Journal of The Royal Society Interface* 14, 20170213.  
517 <https://doi.org/10.1098/rsif.2017.0213>  
518 Negro, J.J., Tella, J.L., Blanco, G., Forero, M.G., Garrido-Fernández, J., 2000. Diet  
519 Explains Interpopulation Variation of Plasma Carotenoids and Skin Pigmentation in  
520 Nestling White Storks. *Physiological and Biochemical Zoology* 73, 97–101.  
521 <https://doi.org/10.1086/316724>  
522 Newsome, T.M., Dellinger, J.A., Pavey, C.R., Ripple, W.J., Shores, C.R., Wirsing, A.J.,  
523 Dickman, C.R., 2015. The ecological effects of providing resource subsidies to  
524 predators. *Global Ecology and Biogeography* 24, 1–11.  
525 <https://doi.org/10.1111/geb.12236>  
526 Oro, D., Genovart, M., Tavecchia, G., Fowler, M.S., Martínez-Abraín, A., 2013.  
527 Ecological and evolutionary implications of food subsidies from humans. *Ecology*  
528 *Letters* 16, 1501–1514. <https://doi.org/10.1111/ele.12187>  
529 Parfitt, J., Barthel, M., Macnaughton, S., 2010. Food waste within food supply chains:  
530 quantification and potential for change to 2050. *Philosophical Transactions of the Royal*  
531 *Society B: Biological Sciences* 365, 3065–3081. <https://doi.org/10.1098/rstb.2010.0126>  
532 Patenaude-Monette, M., Bélisle, M., Giroux, J.-F., 2014. Balancing Energy Budget in a  
533 Central-Place Forager: Which Habitat to Select in a Heterogeneous Environment?  
534 *PLOS ONE* 9, e102162. <https://doi.org/10.1371/journal.pone.0102162>  
535 Patrick, S.C., Bearhop, S., Bodey, T.W., Grecian, W.J., Hamer, K.C., Lee, J., Votier,  
536 S.C., 2015. Individual seabirds show consistent foraging strategies in response to  
537 predictable fisheries discards. *Journal of Avian Biology* 46, 431–440.  
538 <https://doi.org/10.1111/jav.00660>  
539 Payo-Payo, A., Oro, D., Igual, J.M., Jover, L., Sanpera, C., Tavecchia, G., 2015.  
540 Population control of an overabundant species achieved through consecutive  
541 anthropogenic perturbations. *Ecological Applications* 25, 2228–2239.  
542 <https://doi.org/10.1890/14-2090.1>  
543 Peris, S., 2003. Feeding in urban refuse dumps: Ingestion of plastic objects by the White  
544 Stork (*Ciconia ciconia*). *Ardeola: revista ibérica de ornitología* 50.  
545 Pineda-Pampliega, J., Ramiro, Y., Herrera-Dueñas, A., Martínez-Haro, M., Hernández,  
546 J.M., Aguirre, J.I., Höfle, U., 2021. A multidisciplinary approach to the evaluation of  
547 the effects of foraging on landfills on white stork nestlings. *Science of The Total*  
548 *Environment* 145197. <https://doi.org/10.1016/j.scitotenv.2021.145197>  
549 Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D., Heisterkamp, S., Van Willigen, B.,  
550 Maintainer, R., 2017. Package ‘nlme.’ Linear and nonlinear mixed effects models,  
551 version 3.  
552 Plaza, P.I., Lambertucci, S.A., 2018. More massive but potentially less healthy: black  
553 vultures feeding in rubbish dumps differed in clinical and biochemical parameters with  
554 wild feeding birds. *PeerJ* 6, e4645. <https://doi.org/10.7717/peerj.4645>  
555 Plaza, P.I., Lambertucci, S.A., 2017. How are garbage dumps impacting vertebrate  
556 demography, health, and conservation? *Global Ecology and Conservation* 12, 9–20.  
557 <https://doi.org/10.1016/j.gecco.2017.08.002>  
558 Pons, J.-M., Migot, P., 1995. Life-History Strategy of the Herring Gull: Changes in  
559 Survival and Fecundity in a Population Subjected to Various Feeding Conditions.  
560 *Journal of Animal Ecology* 64, 592–599. <https://doi.org/10.2307/5802>  
561 Ripley, B., Venables, B., Bates, D.M., Hornik, K., Gebhardt, A., Firth, D., Ripley,

562 M.B., 2013. Package ‘mass.’ Cran R 538.

563 Shepard, E.L.C., Wilson, R.P., Halsey, L.G., Quintana, F., Laich, A.G., Gleiss, A.C.,

564 Liebsch, N., Myers, A.E., Norman, B., 2008a. Derivation of body motion via

565 appropriate smoothing of acceleration data. *Aquatic Biology* 4, 235–241.

566 <https://doi.org/10.3354/ab00104>

567 Shepard, E.L.C., Wilson, R.P., Quintana, F., Laich, A.G., Liebsch, N., Albareda, D.A.,

568 Halsey, L.G., Gleiss, A., Morgan, D.T., Myers, A.E., Newman, C., Macdonald, D.W.,

569 2008b. Identification of animal movement patterns using tri-axial accelerometry.

570 *Endangered Species Research* 10, 47–60. <https://doi.org/10.3354/esr00084>

571 Steigerwald, E.C., Igual, J.-M., Payo-Payo, A., Tavecchia, G., 2015. Effects of

572 decreased anthropogenic food availability on an opportunistic gull: evidence for a size-

573 mediated response in breeding females. *Ibis* 157, 439–448.

574 <https://doi.org/10.1111/ibi.12252>

575 Stephens, D.W., Krebs, J.R., 1986. *Foraging Theory*. Princeton University Press.

576 Tablado, Z., Tella, J.L., Sánchez-Zapata, J.A., Hiraldo, F., 2010. The Paradox of the

577 Long-Term Positive Effects of a North American Crayfish on a European Community

578 of Predators. *Conservation Biology* 24, 1230–1238. [https://doi.org/10.1111/j.1523-](https://doi.org/10.1111/j.1523-1739.2010.01483.x)

579 [1739.2010.01483.x](https://doi.org/10.1111/j.1523-1739.2010.01483.x)

580 Tauler-Ametller, H., Hernández-Matías, A., Pretus, J.L., Real, J., 2017. Landfills

581 determine the distribution of an expanding breeding population of the endangered

582 Egyptian Vulture *Neophron percnopterus*. *Ibis* 159, 757–768.

583 <https://doi.org/10.1111/ibi.12495>

584 Tongue, A.D.W., Reynolds, S.J., Fernie, K.J., Harrad, S., 2019. Flame retardant

585 concentrations and profiles in wild birds associated with landfill: A critical review.

586 *Environmental Pollution* 248, 646–658. <https://doi.org/10.1016/j.envpol.2019.01.103>

587 van Donk, S., Camphuysen, K.C.J., Shamoun-Baranes, J., Meer, J. van der, 2017. The

588 most common diet results in low reproduction in a generalist seabird. *Ecology and*

589 *Evolution* 7, 4620–4629. <https://doi.org/10.1002/ece3.3018>

590 van Donk, S., Shamoun-Baranes, J., van der Meer, J., Camphuysen, K.C.J., 2019.

591 Foraging for high caloric anthropogenic prey is energetically costly. *Movement Ecology*

592 7, 17. <https://doi.org/10.1186/s40462-019-0159-3>

593 Weiser, E.L., Powell, A.N., 2010. Does Garbage in the Diet Improve Reproductive

594 Output of Glaucous Gulls? *Condor* 112, 530–538.

595 <https://doi.org/10.1525/cond.2010.100020>

596 Wright, J., Both, C., Cotton, P.A., Bryant, D., 1998. Quality vs. Quantity: Energetic and

597 Nutritional Trade-Offs in Parental Provisioning Strategies. *Journal of Animal Ecology*

598 67, 620–634.

599 Ydenberg, R.C., Welham, C.V.J., Schmid-Hempel, R., Schmid-Hempel, P.,

600 Beauchamp, G., 1994. Time and energy constraints and the relationships between

601 currencies in foraging theory. *Behav Ecol* 5, 28–34.

602 <https://doi.org/10.1093/beheco/5.1.28>

603 Zurell, D., Eggers, U., Kaatz, M., Rotics, S., Sapir, N., Wikelski, M., Nathan, R.,

604 Jeltsch, F., 2015. Individual-based modelling of resource competition to predict density-

605 dependent population dynamics: a case study with white storks. *Oikos* 124, 319–330.

606 <https://doi.org/10.1111/oik.01294>

607

608

609

<b>(a) Distance travelled</b>	Estimate	SE	<i>t</i>	<i>p</i>
Fixed effects				
Intercept	20.32	1.50	13.58	<0.001
Season	-13.90	0.81	-17.15	<0.001
Foraging site	-5.001	0.71	-7.07	<0.001
Season:Site	7.65	0.87	8.83	<0.001
GPS positions	0.23	0.04	6.05	<0.001
Random effects				
Bird ID   Tag type	4.98			
Tag type	0.01			
AR(1)	0.50			
Residual variance	15.50			
Marginal R <sup>2</sup>	0.08			
Conditional R <sup>2</sup>	0.17			
<b>(b) Mean ODBA</b>				
Fixed effects				
Intercept	0.09	0.01	10.36	<0.001
Season	-0.02	0.00	-16.24	<0.001
Foraging site	0.01	0.00	12.57	<0.001
Season:Site	0.00	0.00	-1.78	0.0755
Random effects				
Bird ID   Tag type	0.01			
Tag type	0.02			
AR(1)	0.54			
Residual variance	0.03			
Marginal R <sup>2</sup>	0.17			
Conditional R <sup>2</sup>	0.46			
<b>(c) Relative foraging time</b>				
Fixed effects				
Intercept	-0.83	0.06	14.08	<0.001
Season	-0.20	0.03	-5.82	<0.001
Foraging site	0.33	0.03	11.91	<0.001
Season:Site	-0.03	0.03	-0.90	0.3694
Random effects				
Bird ID   Tag type	0.25			
Tag type	0.05			
AR(1)	0.59			
Residual variance	1.58			
Marginal R <sup>2</sup>	0.01			
Conditional R <sup>2</sup>	0.03			
<b>(d) Foraging ODBA</b>				
Fixed effects				
Intercept	0.17	0.01	15.46	<0.001
Season	-0.02	0.002	-9.59	<0.001
Foraging site	0.02	0.002	15.42	<0.001
Season:Site	-0.02	0.002	-8.71	<0.001
Random effects				

Bird ID   Tag type	0.01			
Tag type	0.02			
AR(1)	0.32			
Residual variance	0.03			
Marginal R <sup>2</sup>	0.15			
Conditional R <sup>2</sup>	0.44			

611 Table 1: Estimates from LMMs and GLMM explaining (a) total distance travelled, (b)  
612 mean overall dynamic body acceleration (mean ODBA), (c) relative foraging time, and  
613 (d) mean foraging ODBA. Reference level for season is 'breeding', and for foraging site  
614 is 'landfill'.

615

616

617

618

619

620

621

622

623

624

625

626

627

628

629

630

631

632 Ç

<b>(a) Model-wide Interactions</b>	Predictor	Estimate	SE	<i>p</i>
Response				
Mean ODBA	Distance travelled			<0.001
Mean ODBA	Relative foraging time			<0.001
Mean ODBA	Foraging ODBA			<0.001
Relative foraging time	Foraging ODBA			<0.001
<b>(b) Breeding, non-landfill</b>				
Response				
Mean ODBA	Distance travelled	0.0005	0.0000	<0.001
Mean ODBA	Relative foraging time	0.1049	0.0020	<0.001
Mean ODBA	Foraging ODBA	0.3150	0.0091	<0.001
Relative foraging time	Foraging ODBA	0.8266	0.2827	0.0035
<b>(c) Non-Breeding, non-landfill</b>				
Response				
Mean ODBA	Distance travelled	0.0005	0.0000	<0.001
Mean ODBA	Relative foraging time	0.1016	0.0020	<0.001
Mean ODBA	Foraging ODBA	0.2536	0.0091	<0.001
Relative foraging time	Foraging ODBA	1.6589	0.3157	<0.001
<b>(d) Breeding, landfill</b>				
Response				
Mean ODBA	Distance travelled	0.0006	0.0000	<0.001
Mean ODBA	Relative foraging time	0.0985	0.0068	<0.001
Mean ODBA	Foraging ODBA	0.2218	0.0184	<0.001
Relative foraging time	Foraging ODBA	-0.0627	0.5119	0.9026
<b>(e) Non-Breeding, landfill</b>				
Response				
Mean ODBA	Distance travelled	0.0004	0.0000	<0.001
Mean ODBA	Relative foraging time	0.0691	0.0033	<0.001
Mean ODBA	Foraging ODBA	0.1696	0.0085	<0.001
Relative foraging time	Foraging ODBA	-0.7308	0.2773	0.0085

634 Table 2: Estimates from structural equation models (SEM) explaining the relationship  
635 among total distance travelled, mean overall dynamic body acceleration (ODBA), relative  
636 foraging time, and mean foraging ODBA. Estimates for (a) the whole model, (b) breeding  
637 period outside landfills, (c) non-breeding period outside landfills, (d) breeding period in  
638 landfills, and (e) non-breeding period in landfills.

639

640

641

<b>(a) Relative foraging time</b>	Estimate	SE	<i>t</i>	<i>p</i>
<b>Fixed effects</b>				
Intercept	-1.21	0.11	-11.29	<0.001
Season	-0.03	0.09	-0.28	0.7795
Specialisation index	0.31	0.19	1.65	0.1062
Season:Specialisation index	-0.11	0.14	-0.79	0.4323
<b>Random effects</b>				
Bird ID   Tag type	0.14			
Tag type	0			
AR(1)	0.25			
Residual variance	1.01			
<b>(b) Foraging ODBA</b>				
<b>Fixed effects</b>				
Intercept	0.16	0.01	11.22	<0.001
Season	-0.02	0.01	-2.51	0.0120
Specialisation index	0.01	0.01	0.42	0.6756
Season:Specialisation index	0.01	0.01	0.82	0.4142
<b>Random effects</b>				
Bird ID   Tag type	0.01			
Tag type	0.02			
AR(1)	0.10			
Residual variance	0.03			

642 Table 3: Estimates from GLMM and LMM explaining (a) relative foraging time, and (b)  
643 mean foraging overall dynamic body acceleration (ODBA), while foraging on landfills.  
644 Reference level for season is ‘breeding’.

645

646

647

648

649

650

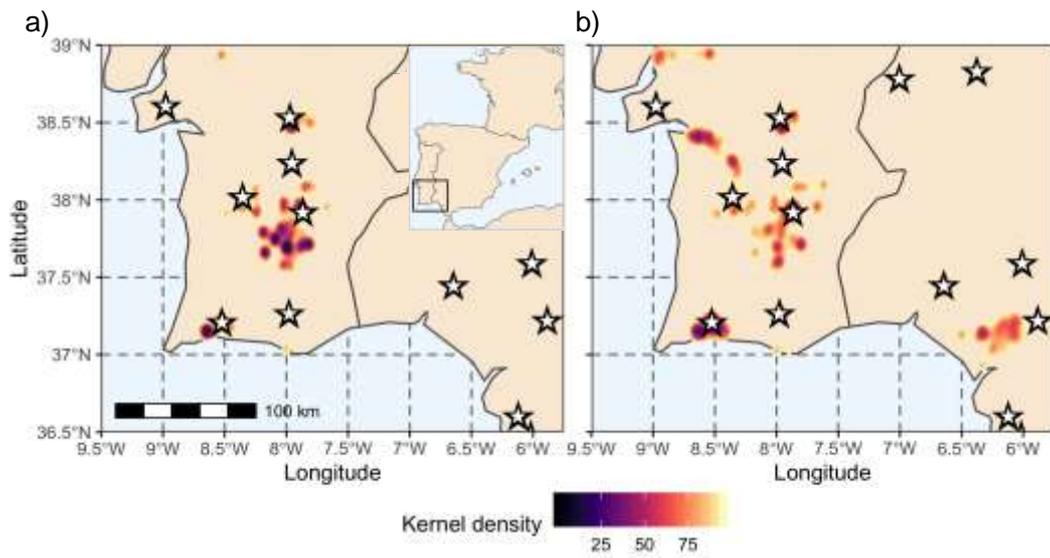
651

652

653

654

655



656

657 Figure 1: Foraging areas of tracked white storks in southern Iberia (Portugal and Spain)

658 between 2016 and 2019 during a) the breeding and b) the non-breeding periods. Stars

659 indicate landfill sites.

660

661

662

663

664

665

666

667

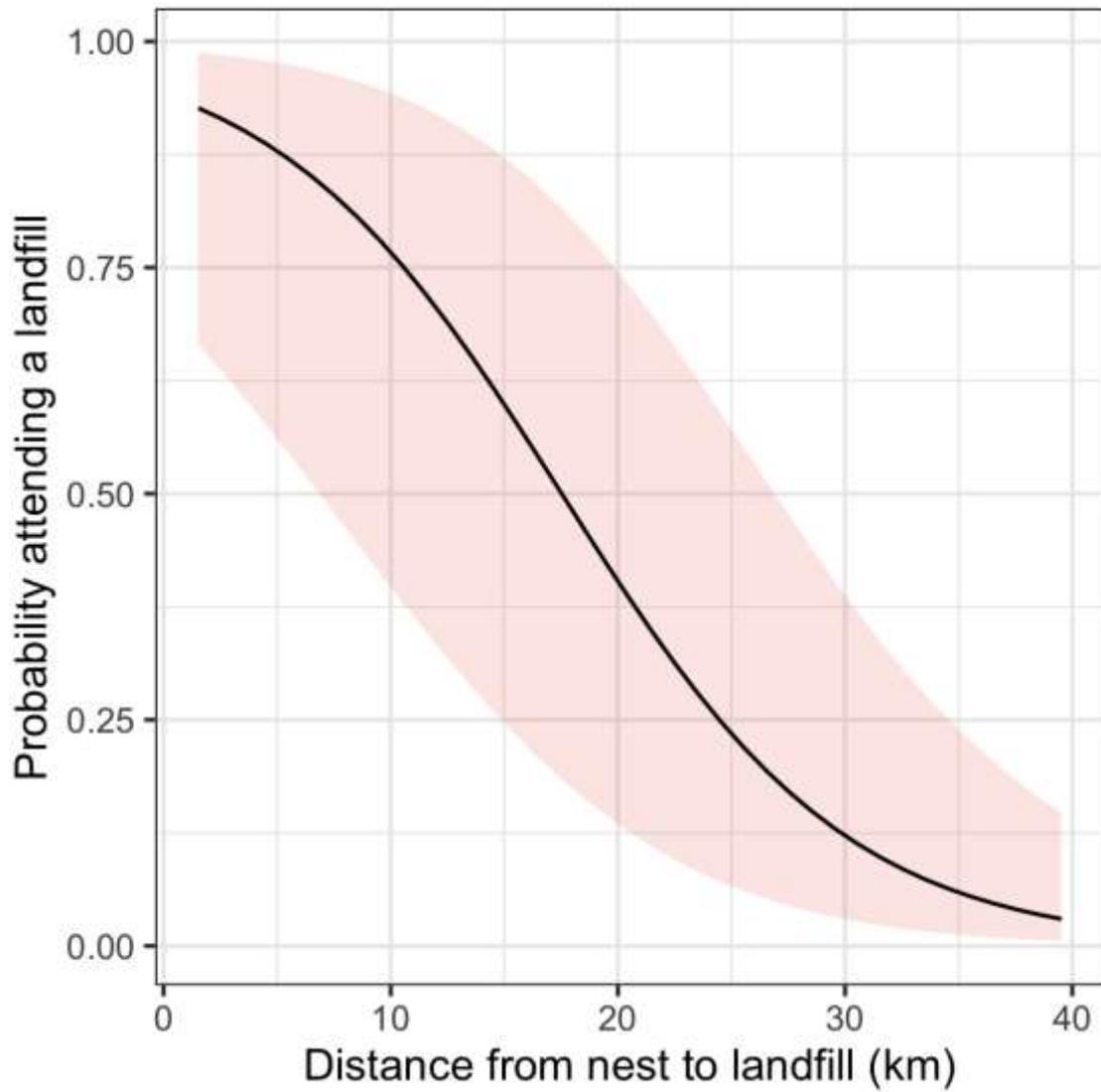
668

669

670

671

672



673

674

675 Figure 2: Estimates from GLMMs explaining the probability of white storks attending a  
676 landfill during the breeding season as a function of the distance from the nesting site to  
677 the closest landfill.

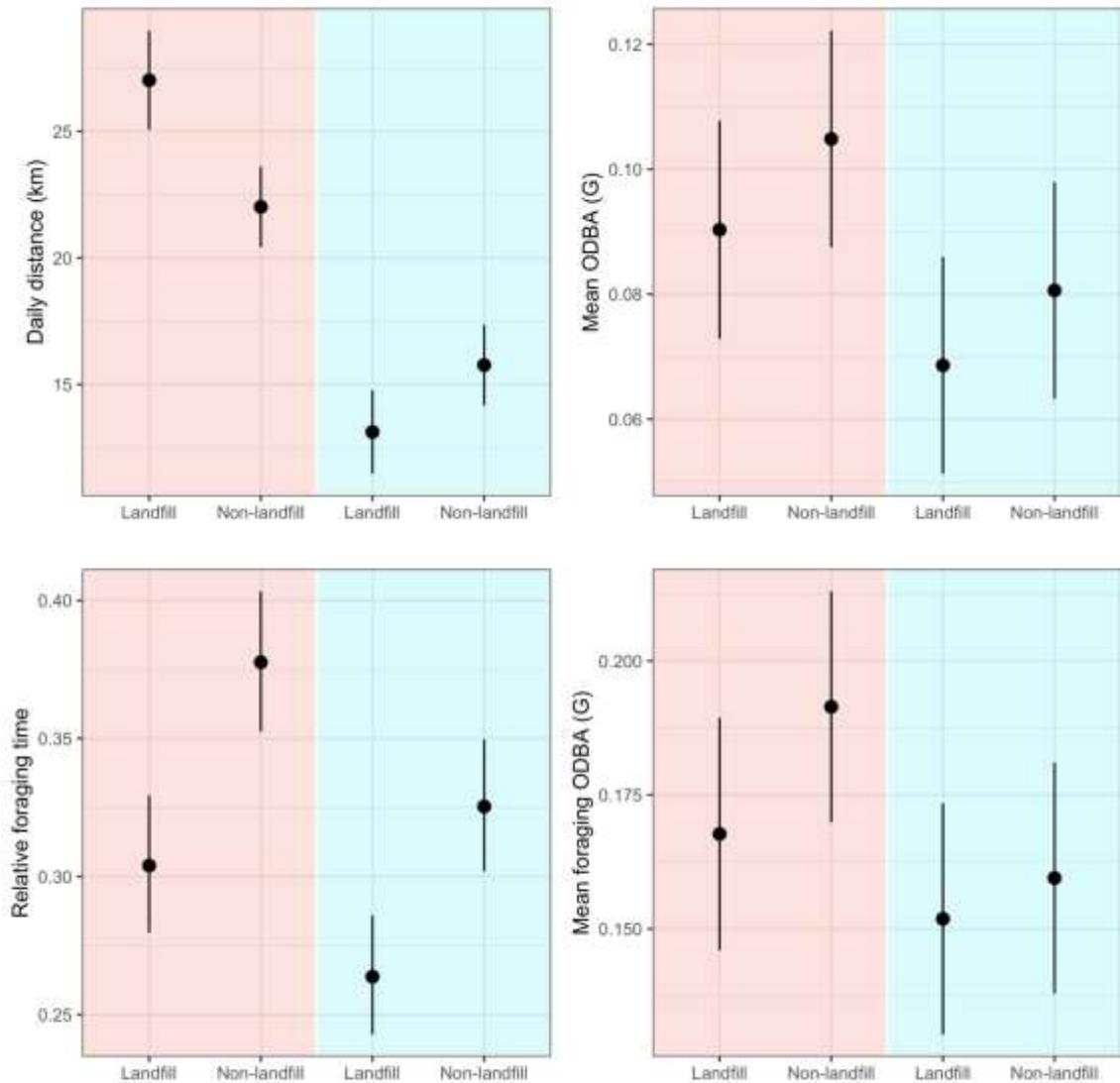
678

679

680

681

682



683

684 Figure 3: Predicted values and 95 % confidence intervals from GLMMs explaining (a)  
 685 daily distance travelled, (b) mean overall dynamic body acceleration (ODBA), (c) relative  
 686 foraging time, and (d) mean foraging ODBA of foraging white storks, as a function of  
 687 season (breeding or non-breeding) and foraging site (landfill or non-landfill).

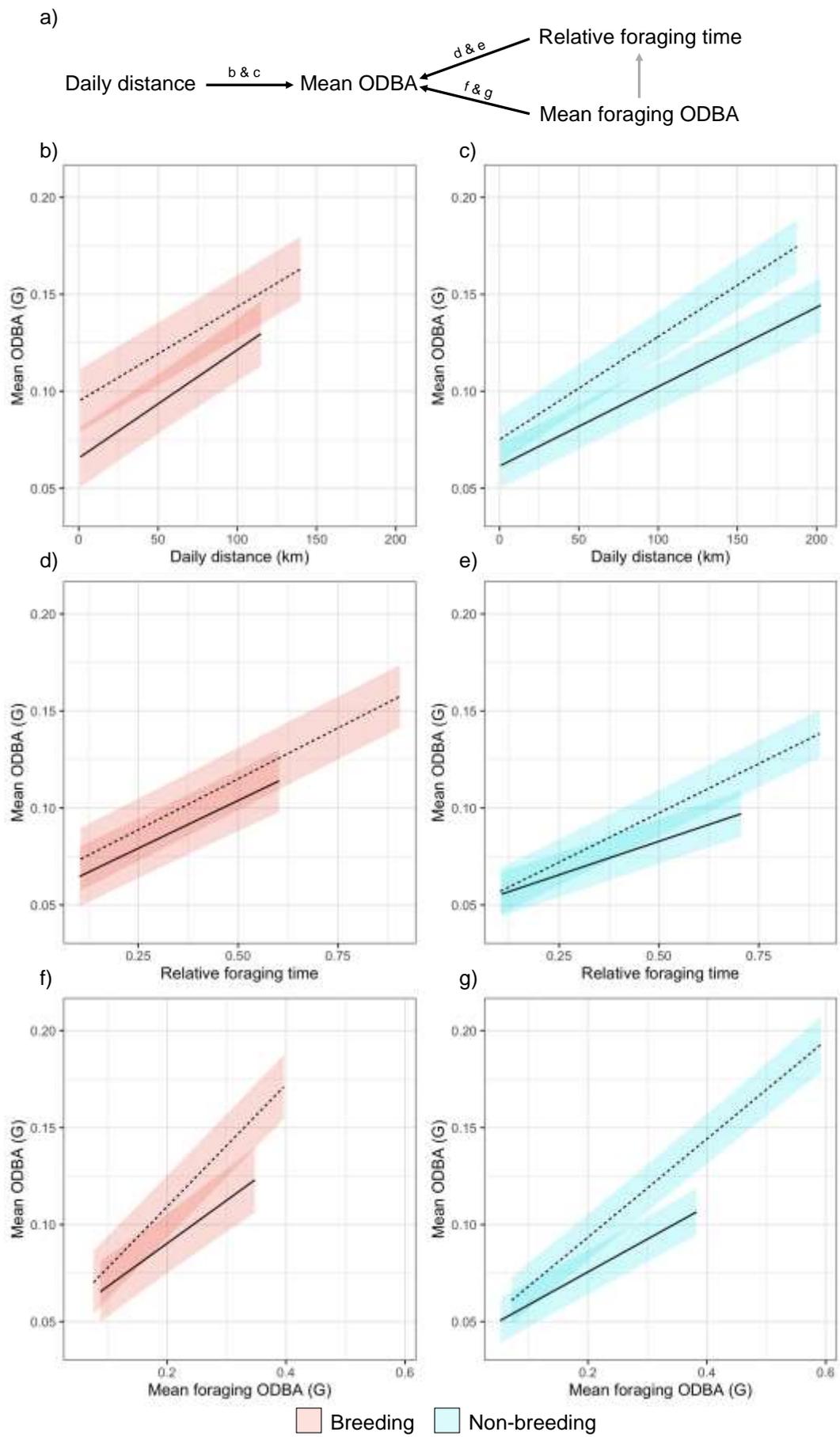
688

689

690

691

692



694 Figure 4: (a) Graphical explanation of the structural equation model (SEM) showing the  
695 relative contribution of white storks' foraging decisions into the mean overall dynamic  
696 body acceleration (ODBA). Black arrows indicate relationships represented in the figure  
697 and the grey arrow relationship not represented in the figure. Mean ODBA as a function  
698 of daily distance travelled for (b) the breeding season and (c) the non-breeding season;  
699 mean ODBA as a function of relative foraging time for (d) the breeding season and (e)  
700 the non-breeding season; and mean ODBA as a function of mean foraging ODBA for (f)  
701 the breeding season and (g) the non-breeding season. Solid line on landfill sites and  
702 dashed line on non-landfill sites.

703

704

705

706

707

708

709

710

711

712

713

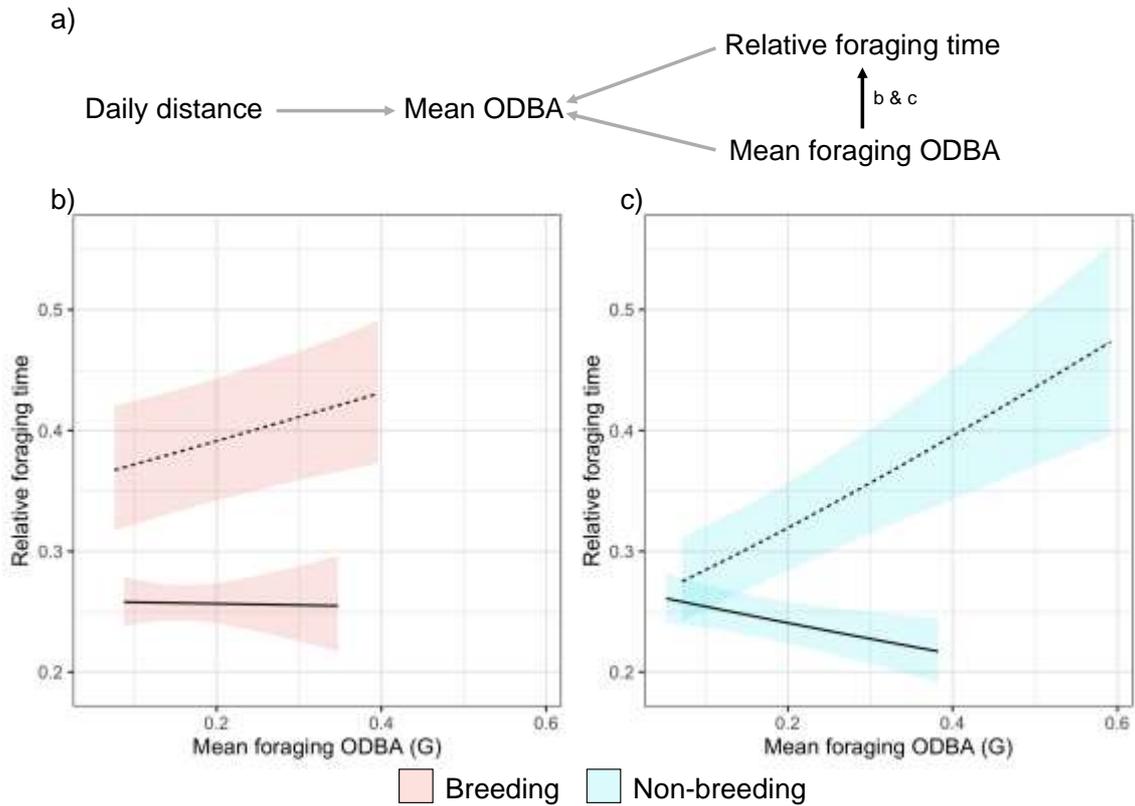
714

715

716

717

718



719

720 Figure 5: (a) Graphical explanation of the structural equation model (SEM) showing the  
 721 relative contribution of white storks' foraging decisions into the mean overall dynamic  
 722 body acceleration (ODBA). Black arrows indicate relationships represented in the figure  
 723 and the grey arrow relationship not represented in the figure. Relative foraging time as a  
 724 function of mean foraging ODBA for (b) the breeding season and (c) the non-breeding  
 725 season. Solid line on landfill sites and dashed line on non-landfill sites.

726

727