

### Key Points:

- Global CH<sub>2</sub>Cl<sub>2</sub> emissions increased by ~85% between 2006 and 2017, mostly due to increasing emissions from Asia
- Global C<sub>2</sub>Cl<sub>4</sub> emissions decreased in the same period by ~25%, mainly due to reduced emissions from Europe and North America
- Posterior CH<sub>2</sub>Cl<sub>2</sub> and C<sub>2</sub>Cl<sub>4</sub> emissions provide good agreement with surface and aircraft observational data

### Supporting Information:

- Supporting Information S1

### Correspondence to:

T. Claxton and R. Hossaini,  
t.claxton@lancaster.ac.uk;  
r.hossaini@lancaster.ac.uk

### Citation:

Claxton, T., Hossaini, R., Wilson, C., Montzka, S. A., Chipperfield, M. P., Wild, O., et al. (2020). A synthesis inversion to constrain global emissions of two very short lived chlorocarbons: dichloromethane, and perchloroethylene. *Journal of Geophysical Research: Atmospheres*, 125, e2019JD031818. <https://doi.org/10.1029/2019JD031818>

Received 11 OCT 2019

Accepted 4 MAR 2020

Accepted article online 8 APR 2020

# A Synthesis Inversion to Constrain Global Emissions of Two Very Short Lived Chlorocarbons: Dichloromethane, and Perchloroethylene

Tom Claxton<sup>1</sup> , Ryan Hossaini<sup>1</sup> , Chris Wilson<sup>2,3</sup> , Stephen A. Montzka<sup>4</sup> , Martyn P. Chipperfield<sup>2,3</sup> , Oliver Wild<sup>1</sup> , Ewa M. Bednarz<sup>1</sup>, Lucy J. Carpenter<sup>5</sup> , Stephen J. Andrews<sup>5</sup> , Sina C. Hackenberg<sup>5,19</sup> , Jens Mühle<sup>6</sup> , David Oram<sup>7</sup>, Sunyoung Park<sup>8</sup> , Mi-Kyung Park<sup>8</sup>, Elliot Atlas<sup>9</sup> , Maria Navarro<sup>9,20</sup>, Sue Schauffler<sup>10</sup> , David Sherry<sup>11</sup>, Martin Vollmer<sup>12</sup> , Tanja Schuck<sup>13</sup> , Andreas Engel<sup>13</sup> , Paul B. Krummel<sup>14</sup> , Michela Maione<sup>15</sup>, Jgor Arduini<sup>15</sup>, Takuya Saito<sup>16</sup>, Yoko Yokouchi<sup>16,21</sup> , Simon O'Doherty<sup>17</sup> , Dickon Young<sup>17</sup> , and Chris Lunder<sup>18</sup>

<sup>1</sup>Lancaster Environment Centre, Lancaster University, Lancaster, UK, <sup>2</sup>School of Earth and Environment, University of Leeds, Leeds, UK, <sup>3</sup>National Centre for Earth Observation, University of Leeds, Leeds, UK, <sup>4</sup>National Oceanic and Atmospheric Administration, Boulder, Colorado, USA, <sup>5</sup>Wolfson Atmospheric Chemistry Laboratories, Department of Chemistry, University of York, York, UK, <sup>6</sup>Scripps Institution of Oceanography, University of California, San Diego, CA, USA, <sup>7</sup>National Centre for Atmospheric Science, School of Environmental Sciences, University of East Anglia, Norwich, UK, <sup>8</sup>Kyungpook Institute of Oceanography, Kyungpook National University, Daegu, South Korea, <sup>9</sup>Rosenstiel School of Marine and Atmospheric Science, University of Miami, Miami, FL, USA, <sup>10</sup>National Centre for Atmospheric Research, Boulder, CO, USA, <sup>11</sup>Nolan Sherry & Associates, London, UK, <sup>12</sup>Laboratory for Air Pollution and Environmental Technology, Empa, Swiss Federal Laboratories for Materials Science and Technology, Dübendorf, Switzerland, <sup>13</sup>Institute for Atmospheric and Environmental Sciences, Goethe University Frankfurt, Frankfurt, Germany, <sup>14</sup>Climate Science Centre, CSIRO Oceans and Atmosphere, Aspendale, Australia, <sup>15</sup>Department of Pure and Applied Sciences, University of Urbino, Urbino, Italy, <sup>16</sup>National Institute for Environmental Studies, Tsukuba, Japan, <sup>17</sup>School of Chemistry, University of Bristol, Bristol, UK, <sup>18</sup>Norwegian Institute for Air Research, Kjeller, Norway, <sup>19</sup>now at Eurofins Dr. Specht International GmbH, Am Neuländer Gewerbepark 2, Hamburg, <sup>20</sup>Deceased 19th December 2017, <sup>21</sup>Deceased 1st January 2020

**Abstract** Dichloromethane (CH<sub>2</sub>Cl<sub>2</sub>) and perchloroethylene (C<sub>2</sub>Cl<sub>4</sub>) are chlorinated very short lived substances (Cl-VSLS) with anthropogenic sources. Recent studies highlight the increasing influence of such compounds, particularly CH<sub>2</sub>Cl<sub>2</sub>, on the stratospheric chlorine budget and therefore on ozone depletion. Here, a multiyear global-scale synthesis inversion was performed to optimize CH<sub>2</sub>Cl<sub>2</sub> (2006–2017) and C<sub>2</sub>Cl<sub>4</sub> (2007–2017) emissions. The approach combines long-term surface observations from global monitoring networks, output from a three-dimensional chemical transport model (TOMCAT), and novel bottom-up information on prior industry emissions. Our posterior results show an increase in global CH<sub>2</sub>Cl<sub>2</sub> emissions from 637 ± 36 Gg yr<sup>-1</sup> in 2006 to 1,171 ± 45 Gg yr<sup>-1</sup> in 2017, with Asian emissions accounting for 68% and 89% of these totals, respectively. In absolute terms, Asian CH<sub>2</sub>Cl<sub>2</sub> emissions increased annually by 51 Gg yr<sup>-1</sup> over the study period, while European and North American emissions declined, indicating a continental-scale shift in emission distribution since the mid-2000s. For C<sub>2</sub>Cl<sub>4</sub>, we estimate a decrease in global emissions from 141 ± 14 Gg yr<sup>-1</sup> in 2007 to 106 ± 12 Gg yr<sup>-1</sup> in 2017. The time-varying posterior emissions offer significant improvements over the prior. Utilizing the posterior emissions leads to modeled tropospheric CH<sub>2</sub>Cl<sub>2</sub> and C<sub>2</sub>Cl<sub>4</sub> abundances and trends in good agreement to those observed (including independent observations to the inversion). A shorter C<sub>2</sub>Cl<sub>4</sub> lifetime, from including an uncertain Cl sink, leads to larger global C<sub>2</sub>Cl<sub>4</sub> emissions by a factor of ~1.5, which in some places improves model-measurement agreement. The sensitivity of our findings to assumptions in the inversion procedure, including CH<sub>2</sub>Cl<sub>2</sub> oceanic emissions, is discussed.

**Plain Language Summary** The 1987 Montreal Protocol banned production for dispersive uses of major ozone-depleting gases, such as chlorofluorocarbons, due to their role in depletion of the stratospheric ozone layer. In consequence, the ozone layer is expected to recover in coming decades, as stratospheric chlorine from banned substances slowly declines. However, chlorinated very short lived substances (Cl-VSLS), not controlled by the Montreal Protocol, represent a small, but growing, source of atmospheric chlorine that could potentially slow ozone recovery. It is thus important that the magnitude of emissions of

©2020. The Authors.

This is an open access article under the terms of the Creative Commons Attribution License, which permits use, distribution and reproduction in any medium, provided the original work is properly cited.

these compounds, their spatial distribution, and changes with time are quantified. Here, we combined observations of Cl-VSLS, prior estimates of their emissions, and a chemical transport model to produce an optimized set of emission estimates on a region-by-region basis between 2006 and 2017. We show that industrial emissions of dichloromethane, the most abundant Cl-VSLS, increased by ~84% within this period, predominately due to an increase in Asian emissions, while European and North American emissions decreased. Over 2007–2017, emissions of perchloroethylene, a less abundant Cl-VSLS, decreased, particularly in Europe and North America. We show that our new emission estimates lead to better agreement with observational data compared to previous estimates.

## 1. Introduction

Halogenated very short lived substances (VSLS) are organic compounds with annual mean atmospheric lifetimes at the planetary surface of ~6 months or less (Engel et al., 2018). These lifetimes are short compared to the principal gases synonymous with ozone depletion, such as chlorofluorocarbons (CFCs), which were banned under the terms of the 1987 Montreal Protocol and its later amendments. However, despite their short lifetimes, over the last two decades, a wealth of research has shown that VSLS of both natural and anthropogenic origin can reach the stratosphere, where they contribute to stratospheric bromine and chlorine and thus ozone depletion (e.g., Claxton et al., 2019; Fernandez et al., 2014; Hossaini et al., 2017; Laube et al., 2008; Sturges et al., 2000; Wales et al., 2018; ). Brominated VSLS (e.g., bromoform and dibromomethane) are predominately of natural oceanic origin (e.g., Quack & Wallace, 2003; Ziska et al., 2013), while chlorinated VSLS (Cl-VSLS) have significant anthropogenic sources (e.g., Engel et al., 2018; McCulloch et al., 1999). At present, these compounds account for a small, but growing, portion of atmospheric chlorine, and they are not controlled by the Montreal Protocol. In 2016, Cl-VSLS were estimated to provide 115 (75–160) ppt of chlorine to the stratosphere, which represents 3.5% of total chlorine in the stratosphere from all sources (Engel et al., 2018; Hossaini et al., 2019).

The most abundant Cl-VSLS, dichloromethane ( $\text{CH}_2\text{Cl}_2$ ), is of particular interest owing to an observed rapid increase in its global concentration since the mid-2000s (Hossaini et al., 2017, 2019; Leedham Elvidge et al., 2015). As a versatile solvent,  $\text{CH}_2\text{Cl}_2$  has a range of industrial applications and roughly 90% of total emissions have been estimated to be anthropogenic (Montzka, Reimann, et al., 2011). Annual global  $\text{CH}_2\text{Cl}_2$  emissions have been estimated at ~1,000 Gg  $\text{yr}^{-1}$  in 2016, with a global mean surface mole fraction of 33–39 ppt observed from monitoring networks, a factor of ~2 larger compared to the early part of the century (Engel et al., 2018). Biogenic  $\text{CH}_2\text{Cl}_2$  sources have also been hypothesized from the ocean (Jones & Carpenter, 2005; Ooki & Yokouchi, 2011) and from mangrove forests (Kolusu et al., 2018), though the magnitudes of these sources are poorly constrained and are expected to be small. A less abundant Cl-VSLS is perchloroethylene,  $\text{C}_2\text{Cl}_4$ , which is almost solely anthropogenic and historically has found use, for example, in dry-cleaning applications. Unlike  $\text{CH}_2\text{Cl}_2$ , the abundance of  $\text{C}_2\text{Cl}_4$  has continually decreased over the last few decades (Carpenter et al., 2014; Simpson et al., 2004), due to phasing out in favor of less-toxic alternatives. In 2016, the global mean  $\text{C}_2\text{Cl}_4$  mole fraction was 1.1–1.2 ppt, with global emissions estimated at 83–103 Gg  $\text{yr}^{-1}$  (Engel et al., 2018).

Claxton et al. (2019) recently quantified the ozone-depletion potential (ODP) of several Cl-VSLS, highlighting a strong dependence of the ODP on the location of emission. They reported ODP ranges for  $\text{CH}_2\text{Cl}_2$  and  $\text{C}_2\text{Cl}_4$  of 0.0097–0.0208 and 0.0057–0.0198, respectively, with emissions from Southern Asia having the largest ODPs. This is significant for Cl-VSLS, as Asian emissions (a) likely account for a large fraction of present-day global total emissions, having grown in importance over the last decade (Fang et al., 2019; Leedham Elvidge et al., 2015; Oram et al., 2017), and (b) may continue to increase in coming years (Feng et al., 2018). On the above basis, it is important that the geographical distribution and strength of Cl-VSLS emissions are investigated and that accurate, up-to-date inventories are available as input for global modeling studies. Such modeling studies examining the stratospheric input of Cl-VSLS have thus far relied on simple surface mixing ratio boundary conditions to constrain surface abundances of  $\text{CH}_2\text{Cl}_2$  and other compounds based on measurements in the remote atmosphere. While these are observationally based and have been implemented so that time trends and latitudinal gradients are captured (Hossaini et al., 2019), zonal variability is not represented by the approach. This includes any

potential colocation of large surface emissions with regions of efficient transport pathways to the upper troposphere/stratosphere, such as from continental East Asia (e.g., Ashfold et al., 2015), which are likely relevant to determining accurate ODPs (Claxton et al., 2019).

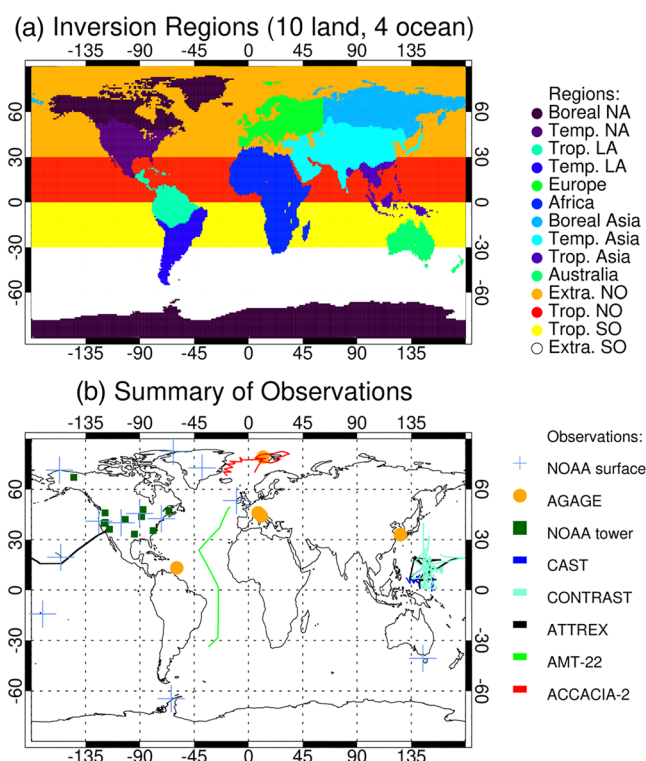
Despite a growing interest in Cl-VSLS, there have been few recent studies examining their emissions at the global scale. Keene et al. (1999) established the Reactive Chlorine Emissions Inventory (RCEI) framework in which global emissions were estimated using a bottom-up approach for a wide range of chlorocarbons. Within that framework, industrial emissions of  $583 \pm 32 \text{ Gg yr}^{-1} \text{ CH}_2\text{Cl}_2$  and  $366 \pm 20 \text{ Gg yr}^{-1} \text{ C}_2\text{Cl}_4$  were estimated (McCulloch et al., 1999). These values, based on analysis relevant to the 1990s, likely underestimate present  $\text{CH}_2\text{Cl}_2$  emissions and overestimate  $\text{C}_2\text{Cl}_4$  emissions, based on recent trends (Engel et al., 2018). Khalil et al. (1999) added to the RCEI framework by estimating total oceanic emissions of  $191 \text{ Gg yr}^{-1} \text{ CH}_2\text{Cl}_2$  and  $19 \text{ Gg yr}^{-1} \text{ C}_2\text{Cl}_4$ . However, Cl-VSLS fluxes from the ocean are highly spatially variable (e.g., Kolusu et al., 2016) and a significantly lower  $\text{CH}_2\text{Cl}_2$  source ( $<90 \text{ Gg yr}^{-1}$ ) has been inferred in later work (Trudinger et al., 2004). Furthermore, while some evidence for in situ  $\text{CH}_2\text{Cl}_2$  production (related to biological activity) has been reported (Ooki & Yokouchi, 2011), the ocean may also take up atmospheric  $\text{CH}_2\text{Cl}_2$  and re-emit it elsewhere (Moore, 2004). This possibly confounds the interpretation of observational results that were used to infer the magnitude of natural emissions in earlier work. In addition, Lobert et al. (1999) estimated a biomass burning  $\text{CH}_2\text{Cl}_2$  source of  $59 \text{ Gg yr}^{-1}$ , though evidence for the existence of this source is missing from more recent analyses (Lawson et al., 2015; Leedham Elvidge et al., 2015; Mühle et al., 2007; Simpson et al., 2011).

There are two core objectives of this study: first, to investigate global and regional changes in  $\text{CH}_2\text{Cl}_2$  and  $\text{C}_2\text{Cl}_4$  emission magnitudes and distributions on a multiannual timescale; second, to generate and evaluate a set of up-to-date global emissions for both compounds, suitable for use as input to atmospheric models. To accomplish this, we performed a global synthesis inversion to optimize Cl-VSLS emissions over the period 2006–2017. Briefly, this approach combines long-term observations from global monitoring networks, prior information on emissions, and a chemical transport model. The paper is structured as follows. The 3-D chemical transport model is described in section 2. The inversion procedure is outlined in section 3, including both the theory and a description of the different observations used. Our main inversion results, including various sensitivity analyses, are presented in section 4. These include the addition of ocean sources of  $\text{CH}_2\text{Cl}_2$  and an added Cl sink of  $\text{C}_2\text{Cl}_4$ . Conclusions and recommendations for future work are given in section 5.

## 2. Description of the TOMCAT Chemical Transport Model

TOMCAT is an offline 3-D Chemistry Transport Model (CTM) (Chipperfield, 2006; Monks et al., 2016) that has been widely used to investigate tropospheric chemistry and transport, including several VSLS-focused studies (e.g., Hossaini et al., 2010, 2019; Claxton et al., 2019). The CTM is forced by six-hourly wind, temperature, and humidity fields taken from the European Centre for Medium-Range Weather Forecasts (ECMWF) ERA-Interim meteorological reanalyses (Dee et al., 2011). The TOMCAT configuration used had a horizontal resolution of  $2.8^\circ \times 2.8^\circ$ , with a vertical resolution of 60 levels, up to an altitude of  $\sim 64 \text{ km}$ . Our model configuration also employs a simplified tropospheric chemistry scheme, reading an offline monthly varying field of the tropospheric hydroxyl radical (OH) concentration (Spivakovsky et al., 2000; Huijnen et al., 2010). The OH field was used in the Atmospheric Tracer Transport Model Intercomparison Project (TransCom) study of  $\text{CH}_4$  (Patra et al., 2011) and leads to an average methyl chloroform lifetime (1992–2007) of  $4.71 (\pm 0.18)$  years in TOMCAT, in reasonable agreement with recent estimates of  $\sim 5$  years obtained from inverse methods (e.g., Rigby et al., 2013). Although the model OH field here is fixed in time, we note that evidence for interannual OH variability, for instance, due to ENSO activity, exists (e.g., Montzka, Krol, et al., 2011; Prinn et al., 2005; Turner et al., 2018).

Both  $\text{CH}_2\text{Cl}_2$  and  $\text{C}_2\text{Cl}_4$  are subject to OH oxidation and photolysis sinks in the model. An additional inversion experiment (see ensuing discussion) was performed for  $\text{C}_2\text{Cl}_4$  in which the competing three-bodied loss reaction of  $\text{C}_2\text{Cl}_4$  with Cl atoms was also included. The inclusion of this reaction in models has been shown to be important to reproduce atmospheric  $\text{C}_2\text{Cl}_4$  observations in the upper troposphere (Hossaini et al., 2019; Rudolph et al., 1996). In this case, the model assumes a fixed tropospheric mean Cl concentration of



**Figure 1.** (a) Map showing the 14 regions (10 land and 4 ocean) used in the inversion. NA = North America, LA = Latin America, NO = Northern Ocean, SO = Southern Ocean, Extra. = Extratropical, Trop. = Tropical, Temp. = Temperate. (b) Summary of the various observations used in this study: weekly flasks at NOAA surface sites (blue pluses); on-site high-frequency measurements at AGAGE surface sites (orange circles); approximately daily flasks at NOAA tall tower sites (green squares). Flight campaigns: CAST (purple); ATTREX (black); CONTRAST (light blue). Ocean campaigns: AMT-22 (green); ACCACIA-22 (red).

$1.3 \times 10^3$  atoms  $\text{cm}^{-3}$  globally, based on model estimates from Hossaini et al. (2016). In practice, the spatial distribution of tropospheric Cl would be nonuniform, and given this uncertainty, this model run is treated as a sensitivity. Reaction rate constants were taken from the 2015 Jet Propulsion Laboratory (JPL) report (Burkholder et al., 2015). For the purposes of this study which investigates source gas emissions, product gas chemistry was not required.

### 3. Description of the Inversion Technique

#### 3.1. Synthesis Inversion

The “synthesis inversion” technique optimizes model prior emissions of a given compound by minimizing differences between modeled and observed mixing ratios (e.g., Baker et al., 2006). This top-down technique is well established and has been used to investigate surface emissions of several compounds, including  $\text{CH}_4$  (McNorton et al., 2018),  $\text{CO}_2$  (Law et al., 2008; Wang et al., 2018),  $\text{CO}$  (Pétron et al., 2002), and  $\text{H}_2$  (Bousquet et al., 2011). Here, we apply the technique to  $\text{CH}_2\text{Cl}_2$  and  $\text{C}_2\text{Cl}_4$  to optimize their emissions for 12-year (2006–2017) and 11-year (2007–2017) periods, respectively, over which a wide range of tropospheric observations are available (section 3.2). Prior surface  $\text{CH}_2\text{Cl}_2$  and  $\text{C}_2\text{Cl}_4$  emissions (section 3.4) were aggregated over a possible 14 source regions (Figure 1(a)). Boundaries for these source regions (10 land and four ocean), which are continental in scale, are adapted from previous TransCom inversion studies (e.g., Baker et al., 2006). The four ocean regions are defined by the following latitude bands: Extratropical Northern Ocean (30–90°N), Tropical Northern Ocean (0–30°N), Tropical Southern Ocean (0–30°S), and Extratropical Southern Ocean (30–90°S). Given the large uncertainty surrounding oceanic  $\text{CH}_2\text{Cl}_2$  emissions (see discussion in section 3.2.3), for this compound, two different inversions were performed as part of our sensitivity analysis. The first did not include any oceanic  $\text{CH}_2\text{Cl}_2$  emission (i.e., it assumed industry sources only), while the second also considered emissions from the ocean regions.

Within each source region, the distribution of emission is fixed (see section 3.4), and the inversion optimizes the total emission from each region on an annual basis.

The technique is based on minimizing the cost function,  $J$ :

$$J(\mathbf{x}) = \frac{1}{2}(\mathbf{x} - \mathbf{x}_b) \cdot \mathbf{B}^{-1} \cdot (\mathbf{x} - \mathbf{x}_b) + \frac{1}{2}(\mathbf{y} - \mathbf{G} \cdot \mathbf{x}) \cdot \mathbf{R}^{-1} \cdot (\mathbf{y} - \mathbf{G} \cdot \mathbf{x}) \quad (1)$$

where  $\mathbf{x}$  is an emission estimate,  $\mathbf{x}_b$  are the prior emissions,  $\mathbf{B}$  is the covariance matrix for the errors in emissions,  $\mathbf{y}$  are the observations,  $\mathbf{R}$  is the covariance matrix for the errors in observations, and  $\mathbf{G}$  is the normalized model output concentration Jacobian matrix. It maps the emission field on to the observation vector  $\mathbf{y}$  via the transport model.

The cost function is at a minimum at  $\mathbf{x} = \mathbf{x}_a$ , where  $\mathbf{x}_a$  is given as (Tarantola & Valette, 1982):

$$\mathbf{x}_a = \mathbf{x}_b + [\mathbf{G}^T \cdot \mathbf{R}^{-1} \cdot \mathbf{G} + \mathbf{B}^{-1}]^{-1} \cdot \mathbf{G}^T \cdot \mathbf{R}^{-1} \cdot (\mathbf{y} - \mathbf{G} \cdot \mathbf{x}_b) \quad (2)$$

Since all the other quantities are known, the posterior emissions for each of the 14 regions analyzed in the inversion can be solved on a year-by-year basis. Note that our justification for estimating annual emissions (e.g., as opposed to monthly resolved) is based on several factors that are outlined in section 3.2.1 below. This solution of  $\mathbf{x}_a$  gives the best match to the observations, while reducing the likelihood of straying



unrealistically from the prior emissions  $\mathbf{x}_b$ . A successful inversion is indicated by a significant reduction in the posterior emission errors compared to the prior emission errors.

### 3.2. Observations

#### 3.2.1. Surface Observations of $\text{CH}_2\text{Cl}_2$ and $\text{C}_2\text{Cl}_4$

Most of the  $\text{CH}_2\text{Cl}_2$  and  $\text{C}_2\text{Cl}_4$  observational data considered in this study come from remote surface sites, as summarized in Tables 1 and 2. We consider monthly mean measurements of both compounds over the 12-year period obtained from a total of 29 unique surface locations, 19 used as input into the inversion, and 10 held back for independent verification. These data are from the National Oceanic and Atmospheric Administration (NOAA) and Advanced Global Atmospheric Gases Experiment (AGAGE) long-term monitoring networks, which have been described extensively in the literature (e.g., Montzka et al., 2018; Prinn et al., 2018). AGAGE network monthly mean measurements include pollution events, while NOAA measurements are mostly obtained at remote sites. Observations obtained from the sites in Table 1 were used directly in the inversion. Between the two networks, a reasonable level of geographical coverage is achieved (see Figure 1(b)). Critically, this includes sites in each of the main industrialized regions where Cl-VSLs emissions are expected to be greatest, such as the continental USA (four sites), Europe (four sites), and East Asia (one site). A conversion factor of 1.1038 was applied to the AGAGE  $\text{CH}_2\text{Cl}_2$  record to account for a known calibration difference between the NOAA-2003 and AGAGE SIO-14 calibration scales of ~10% (Carpenter et al., 2014; Engel et al., 2018). Note that four measurement sites are shared between the two networks; for this study, we use both measurements; however, we convert AGAGE data to NOAA calibration scales. For  $\text{C}_2\text{Cl}_4$ , NOAA and AGAGE use NOAA-2003 and NOAA-2003B calibration scales, respectively, which have been found to agree to within <1%.

We additionally considered NOAA measurements of both compounds in 2015 from the USA-based tall tower network (Table 2). These data were not assimilated in the inversion but rather were used to provide an independent assessment of the prior versus posterior emissions over the USA (at 10 sites).

The availability and abundance of Cl-VSLs measurement data was a principal factor in our decision to estimate annual mean emissions as opposed to monthly resolved emissions. The 19 unique observational sites (Table 1) provide a maximum of 228 monthly mean measurements in a given year. Solving emissions for 14 different regions would, in a monthly resolved inversion, require 168 ( $14 \times 12$ ) model Cl-VSLs tracers for each year of our study period. This number of tracers (168) is comparable to the number of observations we have available to us in a year (maximum of 228 monthly means, assuming no missing data) and would lead to a less well constrained inversion process, as each month's emissions would only be constrained on average by 1.4 observations. In addition, we believe that the large computational expense of running with such a large number of tracers is not warranted on the basis of (1) our study is primarily interested in long-term interannual emission trends and (2) the seasonal cycle of Cl-VSLs is found to be reproduced well using our non-seasonal posterior emissions (see sections 4.6 and 4.7). Finally, we note that there is little information in the literature with which to inform any prior emission seasonality in our model. Furthermore, for  $\text{CH}_2\text{Cl}_2$ , no significant seasonal variation in industrial emissions has been reported (McCulloch & Midgley, 1996).

#### 3.2.2. Aircraft Observations of $\text{CH}_2\text{Cl}_2$ and $\text{C}_2\text{Cl}_4$

We also considered measurements of both Cl-VSLs from three different flight campaigns: the 2014 Co-ordinated Airborne Studies in the Tropics (CAST) mission (Andrews et al., 2016; Harris et al., 2017), the 2014 Convective Transport of Active Species in the Tropics (CONTRAST) mission (Pan et al., 2017), and the 2014 Airborne Tropical Tropopause Experiment (ATTREX) mission (Navarro et al., 2015). The locations of these campaigns are shown in Figure 1(b). The CAST mission (January–February) centered around Guam in the tropical West Pacific and made extensive measurements in the marine boundary layer during 22 flights, with vertical profiles extending up to ~10 km. Likewise, the CONTRAST (January–February) and ATTREX (January–March) missions also sampled the tropical West Pacific in a region centered around Guam. However, these campaigns sampled air from higher altitudes, with ATTREX extending into the lower stratosphere. Data from these three flight campaigns are not used as input to the inversion; instead they are used as independent observations to test the posterior results (in the relevant months of 2014).

**Table 1**  
Summary of Surface Observational Sites Used as Input to the Inversion (Arranged North to South)

Code	Station name, location	Lat (°)	Lon (°)	Elevation (m)	Network
ALT	Alert, Canada	82.5	−62.5	190.0	NOAA
ZEP	Zeppelin, Svalbard, Norway	78.9	11.9	490.0	AGAGE
SUM	Summit, Greenland	72.6	−38.4	3209.5	NOAA
BRW	Barrow, AK, USA	71.3	−156.6	11.0	NOAA
MHD	Mace Head, Ireland	53.3	−9.9	5.0	NOAA, AGAGE
JFJ	Jungfraujoch, Switzerland	46.3	8.0	3580.0	AGAGE
LEF	Park Falls, WI, USA	45.9	−90.3	472.0	NOAA
CMN	Monte Cimone, Italy	44.2	10.7	2165.0	AGAGE
HFM	Harvard Forest, MA, USA	42.5	−72.2	340.0	NOAA
THD	Trinidad Head, CA, USA	41.1	−124.2	107.0	NOAA, AGAGE
NWR	Niwot Ridge, CO, USA	40.1	−105.6	3523.0	NOAA
GSN	Gosan, Jeju, South Korea	33.3	126.2	89.0	AGAGE
MLO	Mauna Loa, HI, USA	19.5	−155.6	3397.0	NOAA
KUM	Cape Kumukai, HI, USA	19.5	−154.8	3.0	NOAA
RPB	Ragged Point, Barbados	13.2	−59.5	42.0	AGAGE
SMO	Tutuila, American Samoa	−14.2	−170.6	42.0	NOAA, AGAGE
CGO	Cape Grim, Australia	−40.7	144.7	94.0	NOAA, AGAGE
PSA	Palmer Station, Antarctica	−64.9	−64.0	10.0	NOAA
SPO	South Pole, Antarctica	−90.0	−24.8	2810.0	NOAA

### 3.2.3. Ocean Emission Data

As noted in section 1, the ocean is a potential source of  $\text{CH}_2\text{Cl}_2$  and  $\text{C}_2\text{Cl}_4$ . However, there are large uncertainties and several important confounding issues that require attention. Given that oceanic emissions have been proposed to be relatively more important for  $\text{CH}_2\text{Cl}_2$  than  $\text{C}_2\text{Cl}_4$  (Keene et al., 1999), we focus most of the following discussion on  $\text{CH}_2\text{Cl}_2$ , which provides a rationale for performing an inversion with and without an ocean  $\text{CH}_2\text{Cl}_2$  source.

First, there is very limited observational data with which to draw any firm conclusions regarding the strength of any oceanic emission. Khalil et al. (1999) estimated a total oceanic  $\text{CH}_2\text{Cl}_2$  source of  $\sim 196 \text{ Gg yr}^{-1}$  distributed in four latitude bands:  $30\text{--}90^\circ\text{N}$  ( $\sim 24 \text{ Gg yr}^{-1}$ ),  $0\text{--}30^\circ\text{N}$  ( $\sim 50 \text{ Gg yr}^{-1}$ ),  $0\text{--}30^\circ\text{S}$  ( $\sim 50 \text{ Gg yr}^{-1}$ ), and  $30\text{--}90^\circ\text{S}$  ( $\sim 72 \text{ Gg yr}^{-1}$ ). Khalil et al. (1999) acknowledged that the data available to them to calculate fluxes, including measured seawater and atmosphere concentrations of  $\text{CH}_2\text{Cl}_2$  (and  $\text{C}_2\text{Cl}_4$ ), were limited. The calculated fluxes were thus deemed to be “extremely uncertain” and later work inferred a significantly smaller upper limit to total ocean  $\text{CH}_2\text{Cl}_2$  emissions ( $< 90 \text{ Gg yr}^{-1}$ ) based on analysis of firm air samples (Trudinger et al., 2004).

**Table 2**  
Summary of Surface Observational Sites Available in 2015 from the NOAA Tall Tower Network (Arranged North to South), not used as Input in the Inversion

Code	Station name, location	Lat (°)	Lon (°)	Elevation (m)	Network
CRV	CARVE, AK, USA	65.0	−147.6	611.4	NOAA
AMT	Argyle, ME, USA	45.0	−68.7	53.0	NOAA
MBO	Mt. Bachelor, OR, USA	44.0	−121.7	2731.0	NOAA
WBI	West Branch, IA, USA	41.7	−91.4	241.7	NOAA
BAO	Boulder, CO, USA	40.1	−105.0	1,584.0	NOAA
WGC	Walnut Grove, CA, USA	38.3	−121.5	0.0	NOAA
STR	Sutro Tower, CA, USA	37.8	−122.5	254.0	NOAA
MWO	Mt. Wilson, CA, USA	34.2	−118.1	1,728.0	NOAA
SCT	Beech Island, SC, USA	33.4	−81.8	115.2	NOAA
WKT	Moody, TX, USA	31.3	−97.3	251.0	NOAA

Second, in addition to a paucity of measurements, observational results and expectations suggest the possibility for very large spatiotemporal variability in ocean  $\text{CH}_2\text{Cl}_2$  fluxes. For example, based on data collected during a cruise in the tropical Atlantic, Kolusu et al. (2016) calculated a mean  $\text{CH}_2\text{Cl}_2$  flux of  $81 (\pm 82) \text{ nmol m}^{-2} \text{ day}^{-1}$ . Given this large variability, short-term observational studies likely lack sufficient spatial and seasonal coverage to provide adequate estimates of annual net emissions over large domains. Extrapolation to infer regional or global emission totals, while common practice, can be problematic. Extrapolating the Kolusu et al. (2016) flux to a tropical ocean band gives  $\sim 236 (\pm 237) \text{ Gg yr}^{-1}$  and to the entire ocean gives a total of  $\sim 915 (\pm 468) \text{ Gg yr}^{-1}$ . This is a similar order to our prior global emission of  $1,011.5 \text{ Gg yr}^{-1}$ , that includes both land and ocean sources.

Third, another major confounding issue related to the above and relevant to drawing inference on the nature of any ocean  $\text{CH}_2\text{Cl}_2$  source related to in situ production was discussed by Moore (2004). Due to

seasonal changes in  $\text{CH}_2\text{Cl}_2$  ocean solubility (a decrease in warmer waters) and large seasonal changes in the  $\text{CH}_2\text{Cl}_2$  concentration, summertime measurements may show ocean supersaturations that are unrelated to in situ production. In consequence, seasonally resolved data (or analyses accounting for temporary fluxes arising from physical effects, e.g., Ooki & Yokouchi, 2011) are needed to determine the degree to which fluxes derived from measured ocean-water saturation are a result of in situ production or simply seasonal changes in solubility and atmospheric concentration. Moore (2004) also provided strong evidence that dissolved  $\text{CH}_2\text{Cl}_2$  persists for extended periods (possibly years to decades) in intermediate and deep ocean waters. In consequence, observed  $\text{CH}_2\text{Cl}_2$  supersaturations in seawater may be caused by its transport from colder waters at higher latitudes. Based on the above, the inferred oceanic  $\text{CH}_2\text{Cl}_2$  source reported in previous studies (Keene et al., 1999; Khalil et al., 1999) may reflect re-equilibration processes and does not necessarily provide evidence for marine production.

A plausible mechanism by which  $\text{CH}_2\text{Cl}_2$  may be produced in the ocean has been proposed and involves the photolysis and subsequent reaction of biogenic precursors, such as  $\text{CH}_2\text{ICl}$ , in seawater (Jones & Carpenter, 2005). To our knowledge, the only observational study that provides some evidence of marine  $\text{CH}_2\text{Cl}_2$  production (related to phytoplankton) is that of Ooki and Yokouchi (2011). That study accounted for the physical factors discussed above to derive a marine  $\text{CH}_2\text{Cl}_2$  in situ source from the Indian Ocean (between  $10^\circ\text{S}$  and  $40^\circ\text{S}$ ) of  $0.29\text{--}0.43 \mu\text{g m}^{-2} \text{day}^{-1}$ . When extrapolated zonally across the globe, a  $\text{CH}_2\text{Cl}_2$  source of  $10\text{--}15 \text{ Gg yr}^{-1}$  was derived for this latitude band. In summary, considering the uncertainties mentioned above, we performed inversions with and without ocean  $\text{CH}_2\text{Cl}_2$  sources.

For the inversion performed allowing net  $\text{CH}_2\text{Cl}_2$  emissions from the ocean, we compare posterior emissions from our inversion to novel measurements from two recent ship cruises: (a) AMT-22 (Atlantic Meridional Transect, RRS *James Cook*) and (b) ACCACIA-2 (Aerosol-Cloud Coupling And Climate Interactions in the Arctic, JR288, RRS *James Clark Ross*). These campaigns took place in October/November 2012 and July/August 2013, respectively. AMT-22 covered a track through the Atlantic Ocean from  $45^\circ\text{N}$  to  $30^\circ\text{S}$  and ACCACIA-2 covered the North Atlantic/Arctic Oceans from  $70^\circ\text{N}$  to  $80^\circ\text{N}$ , including a navigation around the archipelago of Svalbard, Norway (Figure 1(b)). Along these cruise tracks, sea-to-air flux estimates of  $\text{CH}_2\text{Cl}_2$  (only) were derived based on in situ automated measurements of  $\text{CH}_2\text{Cl}_2$  concentrations in surface seawater (from the ships' clean underway seawater supply inlets; nominal depth 5–6 m) and in air from a continuously pumped air inlet (Hackenberg et al., 2017). Details of the GC-MS measurement systems are given in Andrews et al. (2015). The  $\text{CH}_2\text{Cl}_2$  sea-to-air flux was calculated following the approach of Johnson (2010) but would reflect the combination of both physical effects and any in situ production as discussed above. Average fluxes within the latitude limits of our ocean regions (Figure 1 and section 3.1) were calculated and an estimate of the global ocean emission from each latitude band was obtained through a simple extrapolation. These integrated fluxes are a starting point to compare to our posterior ocean emissions for  $\text{CH}_2\text{Cl}_2$  in section 4.4.

### 3.3. Observation Errors

The covariance matrix for errors in observations (e.g.,  $\mathbf{R}$  in equation 1) is made up of various error sources. Our approach to quantifying these follows the framework described by Xiao (2008), which considers (1) “sampling frequency” errors, (2) “measurement” errors, and (3) “mismatch” errors (Chen & Prinn, 2006). Each of these terms are used to define the total observational error and are detailed in turn below.

#### 3.3.1. Sampling Frequency

The first error source arises due to the sampling frequency of the observational networks. That is, how well the observed monthly mean  $\text{CH}_2\text{Cl}_2$  or  $\text{C}_2\text{Cl}_4$  mole fractions are described by a finite number of measurements (Xiao, 2008). For each site, and each month, the total sampling frequency error,  $\sigma_{sf}$ , for an observational monthly mean is given as

$$\sigma_{sf} = \sqrt{\frac{\sigma_{mon}^2}{m}} \quad (3)$$

where  $\sigma_{mon}^2$  is the variance of the reported mole fractions over the month and  $m$  is the number of observations in that month. For AGAGE surface sites, where measurements are obtained at relatively high frequency (order of 200 measurements per month), the sampling frequency error is calculated according to

equation 3. As it is difficult to assess the independence of successive measurements, equation 3 assumes uncorrelated observations. This might lead to an underestimation in  $\sigma_{sf}$ , but this is likely to be small compared to the overall error. For the NOAA surface sites, mole fractions are obtained based on paired flask samples obtained approximately weekly (i.e., relatively low frequency). Therefore, following Xiao (2008) sampling frequency errors for the NOAA data points were generated from the TOMCAT model, using 30-min averaged output at each of the NOAA locations.

### 3.3.2. Measurement Error

A second source of error arises from errors in the measurements. These can result from instrument precision or other uncertainties in the measuring techniques, such as calibration imperfections. Every observation will have a measurement error, although these are often difficult to fully estimate. In terms of precisions, the AGAGE network reports 0.5% for both  $\text{CH}_2\text{Cl}_2$  and  $\text{C}_2\text{Cl}_4$  based on the measurement precisions of the working standard used (Prinn et al., 2018), while the NOAA network reports precision for each individual measurement, which is aggregated over each month (typically around 0.7%). In this study, we assume a minimum overall 5% measurement error ( $\sigma_{\text{meas}}$ ) for both compounds. This value is based on the study of Andrews et al. (2016) who performed an intercomparison of  $\text{CH}_2\text{Cl}_2$  mole fractions obtained by four different instruments, operated by four different groups, using the same standards. The results indicated that the mean absolute percentage error between the four instruments was ~5% in the troposphere.

### 3.3.3. Mismatch Error

An additional source of error is the mismatch between the observations and the model. This arises when comparing relatively low spatial resolution model output to point observations. An observational site could be unrepresentative of the model grid cell that it is located in. For example, the Harvard Forest (HFM) surface site is in the same TOMCAT grid cell as New York and other parts of the US Eastern seaboard. However, the site lies in the middle of a forest with presumably lower emissions and concentrations more characteristic of other rural observations. To take this into account, a mismatch error can be defined using the neighboring grid cells (Chen & Prinn, 2006). This is defined in equation 4:

$$\sigma_{\text{mismatch}} = \sqrt{\frac{1}{9} \sum_{i=1}^9 (c_i - \bar{c})^2}, \quad (4)$$

where  $c_i$  is the model concentration output for each of the eight neighboring grid cells, taken as an annual mean, and  $\bar{c}$  is the mean model output over the nine cells. The mismatch error equation is a measure of the spatial variance, and although it is not a perfect metric, it helps to place uncertainty on observations with significant variation in their locality.

The three sources of error are combined in equation 5 to give a total observational error:

$$\sigma_{\text{total}} = \sqrt{\sigma_{sf}^2 + \sigma_{\text{meas}}^2 + \sigma_{\text{mismatch}}^2} \quad (5)$$

Of the three error terms contributing in equation 5, the sampling frequency term is typically small (<0.1% relative to observations) compared to, for example, the measurement error (5%). The size of the mismatch error is on average 2% but can vary strongly across sites. For some sites, particularly ones that neighbor urban locations, it can be as large as 15%, or even up to 150% at one site in particular (GSN). For more remote sites (e.g., in the Arctic), the mismatch error could be as low as 0.5%, 10 times lower than the measurement error.

## 3.4. Prior Emissions: Magnitude and Errors

Our prior emission estimates for  $\text{CH}_2\text{Cl}_2$  and  $\text{C}_2\text{Cl}_4$  are summarized in Tables 3 and 4, respectively. Note that these annual priors are held constant over each year of the inversion period. For  $\text{CH}_2\text{Cl}_2$ , prior estimates of Asian, European, and North American emissions (i.e., the expected three most significant industrialized regions) are 671, 50, and 55  $\text{Gg yr}^{-1}$ , respectively (based on data from Nolan Sherry Associates, NSA). These bottom-up estimates (see also Table S1 in the supporting information) were commissioned for this study and represent expected industrial emissions in 2016, based on a global industry database of chloromethane production and production capacity available to NSA. Production figures are calculated and refined by a combination of this extensive database, industry dialogue, trade data, and back-calculations



**Table 3**

*A Summary of Prior (2016 Best Estimate) and Posterior  $\text{CH}_2\text{Cl}_2$  Emissions ( $\text{Gg yr}^{-1}$ ) and their Uncertainties, from the Synthesis Inversion not Allowing for an Oceanic  $\text{CH}_2\text{Cl}_2$  Source*

Region	2006			2017	
	Prior emissions	Posterior emissions	Error reduction	Posterior emissions	Error reduction
Europe	50.0	$112.0 \pm 9.1$	81.9%	$75.1 \pm 11.4$	77.3%
Africa	9.18	$16.6 \pm 8.4$	8.0%	$19.2 \pm 8.7$	5.2%
Australia	4.85	$3.90 \pm 2.22$	54.2%	$3.41 \pm 2.62$	45.9%
Boreal Asia	6.81	$-19.8 \pm 5.4$	20.5%	$-21.8 \pm 6.2$	9.3%
Boreal NA	1.11	$0.002 \pm 1.08$	3.0%	$0.14 \pm 1.11$	1.1%
Temperate Asia	621.0	$89.9 \pm 22.8$	96.3%	$590.7 \pm 28.4$	95.4%
Temperate LA	8.43	$-2.57 \pm 4.68$	44.5%	$0.96 \pm 5.62$	33.4%
Temperate NA	55.0	$71.1 \pm 4.9$	91.1%	$32.1 \pm 5.9$	89.3%
Tropical Asia	50.0	$341.4 \pm 22.7$	54.5%	$454.2 \pm 28.7$	42.6%
Tropical LA	8.67	$24.1 \pm 7.8$	10.1%	$17.1 \pm 8.1$	7.0%
Combined Asia	671.0	$431.3 \pm 32.2$	-	$1,044.9 \pm 40.4$	-
Global total	815.1	$636.6 \pm 36.5$	-	$1,171.2 \pm 44.9$	-

*Note.* See the main text for a description of the prior emissions. NA = North America; LA = Latin America. Combined Asia = Temperate + Tropical.

based on known feedstock applications and quantities. These are entered into a chloromethanes mass balance scheme which is checked against industry capacity and closely calculated production ratios. Of the  $671 \text{ Gg yr}^{-1}$  industry estimate of total Asian  $\text{CH}_2\text{Cl}_2$  emissions from NSA,  $621 \text{ Gg yr}^{-1}$  (~93%) is set as the inversion prior estimate for our Temperate Asia region (incorporating the NSA data for China, India, Japan, and Korea). The remaining  $50 \text{ Gg yr}^{-1}$  is taken as the prior for our Tropical Asian region (where NSA analysis shows the major markets for  $\text{CH}_2\text{Cl}_2$  are Thailand, Indonesia, Singapore, Malaysia, and Vietnam). For the other six land regions, in the absence of more recent up-to-date data, prior industry  $\text{CH}_2\text{Cl}_2$  emissions are taken from the RCEI, as summarized by Keene et al. (1999).

Recall that for  $\text{CH}_2\text{Cl}_2$ , two inversions are performed, one without ocean emissions and one with. For the without ocean case, our global total  $\text{CH}_2\text{Cl}_2$  prior is  $\sim 815 \text{ Gg yr}^{-1}$  (Table 3), that is, considering industrial emissions only. For the with ocean case, prior estimates of ocean  $\text{CH}_2\text{Cl}_2$  emissions from four different ocean regions (see also sections 3.1 and 3.2) are taken from Khalil et al. (1999), also part of the RCEI framework.

**Table 4**

*A Summary of Prior (2016 Best Estimate) and Posterior  $\text{C}_2\text{Cl}_4$  Emissions ( $\text{Gg yr}^{-1}$ ) and their Uncertainties, from the Synthesis Inversion*

Region	2007			2017	
	Prior emissions	Posterior emissions	Error reduction	Posterior emissions	Error reduction
Europe	48.0	$65.2 \pm 4.4$	90.9%	$36.6 \pm 2.6$	94.6%
Africa	2.30	$3.77 \pm 2.20$	4.5%	$3.65 \pm 2.09$	9.2%
Australia	0.62	$1.44 \pm 0.33$	47.1%	$0.52 \pm 0.25$	59.7%
Boreal Asia	1.80	$-2.34 \pm 1.66$	7.6%	$-2.69 \pm 1.60$	10.9%
Boreal NA	0.50	$-0.06 \pm 0.48$	5.3%	$0.52 \pm 0.47$	6.7%
Temperate Asia	93.3	$1.92 \pm 8.80$	90.6%	$6.47 \pm 7.35$	92.1%
Temperate LA	1.06	$2.00 \pm 1.03$	2.8%	$2.04 \pm 0.97$	9.1%
Temperate NA	24.0	$44.8 \pm 2.7$	88.6%	$33.5 \pm 1.8$	92.3%
Tropical Asia	15.0	$38.1 \pm 8.0$	45.1%	$35.0 \pm 7.9$	46.1%
Tropical LA	1.58	$2.73 \pm 1.55$	2.1%	$2.29 \pm 1.53$	3.3%
Extratropical NO	3.51	$-16.5 \pm 2.2$	37.8%	$-12.6 \pm 1.7$	51.2%
Extratropical SO	5.85	$-0.50 \pm 0.79$	86.4%	$-0.14 \pm 0.65$	89.0%
Tropical NO	3.51	$-0.63 \pm 1.66$	52.7%	$1.06 \pm 1.59$	54.5%
Tropical SO	5.85	$0.93 \pm 1.25$	78.7%	$-0.09 \pm 1.25$	78.7%
Combined Asia	108.3	$40.0 \pm 11.9$	-	$41.4 \pm 10.8$	-
Global total	206.5	$140.8 \pm 13.8$	-	$106.1 \pm 12.0$	-

*Note.* Results are based on inversion that did not include the  $\text{C}_2\text{Cl}_4$  + Cl sink.

The total ocean  $\text{CH}_2\text{Cl}_2$  prior is  $197 \text{ Gg yr}^{-1}$ , increasing the global total prior to  $1,012 \text{ Gg yr}^{-1}$  in the “with ocean” inversion case (Table S2). Note that the original RCEI inventory also included a small biomass burning  $\text{CH}_2\text{Cl}_2$  source of  $59 \text{ Gg yr}^{-1}$  (Lobert et al., 1999). However, this estimate was based on an assumed single global  $\text{CH}_2\text{Cl}_2/\text{CO}$  emission ratio for all fuel types. Subsequent studies have reported a lower (by two orders of magnitude)  $\text{CH}_2\text{Cl}_2/\text{CO}$  ratio (Simmonds et al., 2006) or have found no evidence for significant  $\text{CH}_2\text{Cl}_2$  enhancements in biomass burning plumes (Lawson et al., 2015; Leedham Elvidge et al., 2015; Mühle et al., 2007; Simpson et al., 2011). On this basis, a biomass burning  $\text{CH}_2\text{Cl}_2$  source was not considered in the present work.

For  $\text{C}_2\text{Cl}_4$ , a similar approach was adopted whereby prior industry emission estimates for our Asia, Europe, and North American regions are adapted from 2016 bottom-up estimates obtained from NSA (Tables 4 and S1). Similarly to  $\text{CH}_2\text{Cl}_2$ , the Asia estimate is distributed among our Temperate and Tropical Asian regions as  $93.3$  and  $15.0 \text{ Gg yr}^{-1}$ , respectively. For the other six land regions, prior  $\text{C}_2\text{Cl}_4$  emissions were formulated by reducing industrial emissions from the RCEI inventory by a factor of 2. This reduction was performed because tropospheric  $\text{C}_2\text{Cl}_4$  mixing ratios have been observed to be declining since 2000 or earlier (e.g., Simpson et al., 2004; Simmonds et al., 2006), meaning the older RCEI estimates (formulated in the 1990s) are very likely to overestimate present-day emissions. The magnitude of our resultant global total  $\text{C}_2\text{Cl}_4$  prior emission ( $207 \text{ Gg yr}^{-1}$ ), of which 9% is from the ocean, is therefore in closer agreement to more recent independent global estimates (e.g., Engel et al., 2018).

In addition to the observational errors necessary to the inversion procedure (section 3.3), there are also errors in the prior emission estimates discussed above. As these are generally poorly quantified in inversion studies, they are set to  $\pm 100\%$  for all regions as default. The sensitivity of our results to assumptions about prior errors is discussed in section 4.3.

### 3.5. Prior Emissions: Distribution

Within the continental-scale regions considered in this study (Figure 1(a)),  $\text{CH}_2\text{Cl}_2$  emissions are distributed according to a recent  $1^\circ \times 1^\circ$  global HCFC-22 emissions inventory reported by Xiang et al. (2014). The rationale behind this choice is that  $\text{CH}_2\text{Cl}_2$  is coproduced by industry with  $\text{CHCl}_3$  (Oram et al., 2017), and the latter is used almost exclusively as a feedstock in the production of HCFC-22 and fluoropolymers (Fang et al., 2019; Mühle et al., 2019; Tsai, 2017), despite  $\text{CH}_2\text{Cl}_2$  emissions likely being primarily associated with use, not production. On this basis, the use of the HCFC-22 emission distribution can be used as a reasonable proxy for  $\text{CH}_2\text{Cl}_2$  and is a desirable alternative to the far older RCEI distribution. We understand that HCFC-22 is also likely to be emitted where it is used, not where it is produced, which makes this a rough approximation. In the similar absence of more recent data, the HCFC-22 distribution was used as a proxy for  $\text{C}_2\text{Cl}_4$ . It is important to stress that (a) these distributions only affect fluxes within regions (Figure 1(a)) and (b) that the inversion procedure adjusts the integrated regional total emissions, on a region-by-region basis. The distribution of our prior  $\text{CH}_2\text{Cl}_2$  emissions is presented in Figure S1. It is assumed that the within-region distribution does not change over our study period (2006–2017).

## 4. Results and Discussion

### 4.1. Posterior $\text{CH}_2\text{Cl}_2$ Emissions and Trends

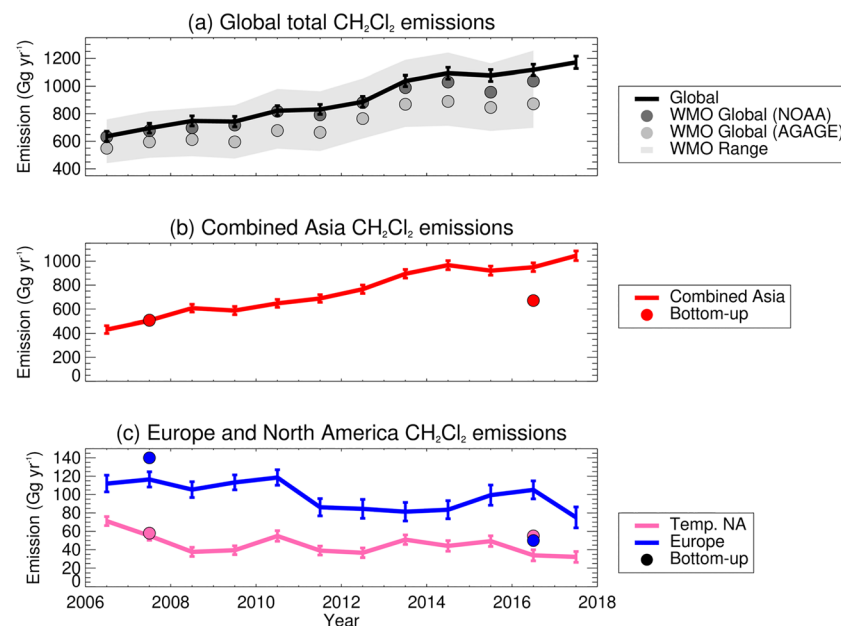
The synthesis inversion produces regional emission estimates, on an annual basis, for each of the 12 years studied. We investigated the degree to which our inversion was able to differentiate between emissions arising from one region over another. A strong negative covariance was found for the closely located regions, Temperate Asia and Tropical Asia, which implies a difficulty in differentiating between these two regions. On this basis, in the ensuing discussion results from these regions are combined and referred to as “combined Asia.” We first consider results from the “no ocean”  $\text{CH}_2\text{Cl}_2$  inversion. Table 3 compares prior and posterior  $\text{CH}_2\text{Cl}_2$  emissions for 2006 and 2017, the first and last years of our study, highlighting an increase in posterior global total  $\text{CH}_2\text{Cl}_2$  emissions from  $637 \pm 37 \text{ Gg yr}^{-1}$  (2006) to  $1,171 \pm 45 \text{ Gg yr}^{-1}$  (2017). This 84% increase is largely due to increasing emissions from combined Asia, estimated to rise from  $431 \pm 32 \text{ Gg yr}^{-1}$  in 2006 to  $1,045 \pm 40 \text{ Gg yr}^{-1}$  in 2017. Our results thus imply that combined Asian emissions more than doubled during the study period and account for  $\sim 70\%$  of global total  $\text{CH}_2\text{Cl}_2$  emissions in

2006 and ~90% in 2017. The latter is a similar relative proportion to that derived from the bottom-up information from NSA presented in Table S1.

While there are no other estimates of *total* Asian  $\text{CH}_2\text{Cl}_2$  emissions in the literature, to our knowledge, some country-specific estimates have been reported. Oram et al. (2017) roughly estimated Chinese  $\text{CH}_2\text{Cl}_2$  emissions of 455 (410–500)  $\text{Gg yr}^{-1}$  in 2015 from bottom-up information from NSA. However, significantly smaller Chinese  $\text{CH}_2\text{Cl}_2$  emissions in 2016 of 318 (254–384)  $\text{Gg yr}^{-1}$  have also been reported, apparently also based on bottom-up information (Feng et al., 2018), thus highlighting the uncertainty in the regional budget. Our estimate of total Asian emissions (1,045  $\text{Gg yr}^{-1}$  in 2017) includes emissions from other major economies, such as India, expected to be significant emitters of  $\text{CH}_2\text{Cl}_2$  (e.g., Leedham Elvidge et al., 2015). The sensitivity of the above findings to inclusion of ocean emissions in our inversion is discussed in section 4.4.

For other major industrialized regions, North America and Europe, our posterior emissions show a decrease over the 12-year study period (2006–2017).  $\text{CH}_2\text{Cl}_2$  emissions from North America decreased from  $71 \pm 5$  to  $32 \pm 6$   $\text{Gg yr}^{-1}$  (–55%) and from Europe decreased from  $112 \pm 9$  to  $75 \pm 11$   $\text{Gg yr}^{-1}$  (–33%). Again, there is limited information in the literature to compare these findings to. Combining surface observations, model calculations, and CO ratio methods, Simmonds et al. (2006) derived European top-down  $\text{CH}_2\text{Cl}_2$  emissions of 51–61  $\text{Gg yr}^{-1}$  over the 2002–2004 period. Our estimate of European  $\text{CH}_2\text{Cl}_2$  emissions, for the closest year to their study (2006), is larger at 112  $\text{Gg yr}^{-1}$ . Simmonds et al. (2006) also reported a bottom-up estimate of industrial  $\text{CH}_2\text{Cl}_2$  emissions of 139  $\text{Gg yr}^{-1}$  from Europe, based on industry sales, in 2002/2003. We note that this is a very similar figure to our bottom-up estimate from NSA (albeit for 2007, Table S1).

Figure 2 presents a time series of annual posterior  $\text{CH}_2\text{Cl}_2$  emissions for a selection of the most important regions. The top panel shows the global total  $\text{CH}_2\text{Cl}_2$  emission over the 12-year study period, the middle panel the contribution from our combined Asian region, and the bottom panel European and North American emissions. Also shown in the top panel are independent estimates of global  $\text{CH}_2\text{Cl}_2$  emissions



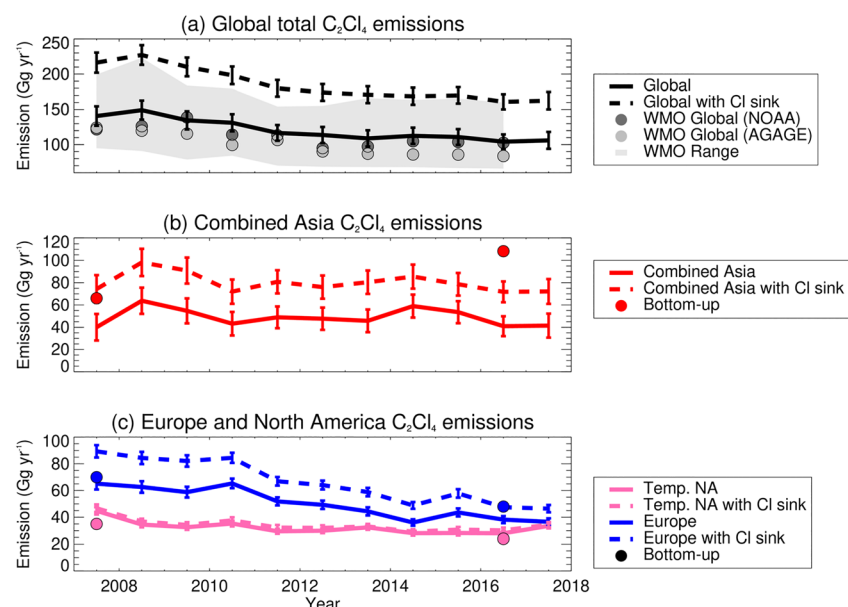
**Figure 2.** Timeseries of posterior  $\text{CH}_2\text{Cl}_2$  emissions ( $\text{Gg yr}^{-1}$ ) over the 12-year (2006–2017) study period. (a) Global total emissions from the inversion (black line, this work) alongside estimates from a 12-box model (circles) forced by NOAA (dark gray) and AGAGE (light gray) observations, as reported in Engel et al. (2018). The full 12-box model uncertainty range is represented by pale gray shading. (b) Asian emissions from the inversion showing Combined Asia (Temperate + Tropical), alongside bottom-up estimates from NSA (circles). (c) European and North American emissions, alongside bottom-up estimates from NSA (circles). See section 3.4 for a description of the bottom-up data. Note that the  $\text{CH}_2\text{Cl}_2$  results shown here are for the no oceanic emission scenario. Error bars represent uncertainty ranges included in Table 3.

(e.g., Cunnold et al., 1983; Rigby et al., 2013). These annual emission data were prepared for the 2018 WMO/UNEP Scientific Assessment of Ozone Depletion and show  $\text{CH}_2\text{Cl}_2$  emissions increasing from  $442\text{--}759\text{ Gg yr}^{-1}$  in 2006 to  $698\text{--}1,256\text{ Gg yr}^{-1}$  in 2016, with the ranges in each period reflecting results obtained considering the two different observational networks analyzed by the 12-box model (Engel et al., 2018). Good agreement between results from this study and those of the 12-box model is found, particularly when the latter assimilates NOAA data, which is plausible as our model also incorporates  $\text{CH}_2\text{Cl}_2$  data on the NOAA calibration scale. For example, our global total  $\text{CH}_2\text{Cl}_2$  emission in 2006 ( $637 \pm 37\text{ Gg yr}^{-1}$ ) and 2016 ( $1,117 \pm 41\text{ Gg yr}^{-1}$ ) fall within the 12-box model ranges noted above. In the most recent years, our posterior emissions fall towards the upper bound of the full uncertainty range of the 12-box model calculations (Figure 2). The relative increase in global  $\text{CH}_2\text{Cl}_2$  emissions between 2006 and 2016 is 61% (12-box model average) and 75% (this work), and the mean annual differences ( $\pm 1\text{SD}$ ) between our total emissions and the 12-box model AGAGE and NOAA estimates are  $159 \pm 51$  and  $42 \pm 36\text{ Gg yr}^{-1}$ , respectively.

Our inversion approach allows us to examine the regional drivers of the increase in global  $\text{CH}_2\text{Cl}_2$  emissions, which—as apparent from Figure 2(b)—are strongly driven by increasing emissions from Asia. In contrast, the relative changes in emissions from Europe and North America over the study period are relatively small. As previously noted, both these latter regions experienced an overall decrease in emissions, though the time series is also characterized by significant interannual variability (Figure 2(c)). Figure 2 also includes, on a regional basis, the bottom-up estimates of  $\text{CH}_2\text{Cl}_2$  emissions from NSA for Asia, Europe, and North America in the years 2007 and 2016 (Table S1). Recall that these 2016 inventory-based estimates were used as the prior emissions for these three respective regions in our inversion (see section 3.4). Like top-down estimates, any bottom-up inventory-based emission data is subject to uncertainty. Therefore, we do not overinterpret these data, though note (a) that they imply a striking decrease in European  $\text{CH}_2\text{Cl}_2$  emissions between 2007 and 2016 that is larger than predicted by our posterior emissions and (b) that discrepancies between top-down  $\text{CH}_2\text{Cl}_2$  emissions (from Europe) and bottom-up estimates have been previously reported (Simmonds et al., 2006). Our North American posterior emissions in 2007 more closely relate to the bottom-up estimate; however, our posterior emissions in 2016 are slightly lower than the bottom-up estimate but agree within the uncertainty range of the inversion (Figure 2(c)).

#### 4.2. Posterior $\text{C}_2\text{Cl}_4$ Emissions and Trends

The posterior  $\text{C}_2\text{Cl}_4$  emissions are summarized in Table 4 and Figure 3. The tabulated results are based on an inversion that only included loss of  $\text{C}_2\text{Cl}_4$  by OH and photolysis, ignoring the  $\text{C}_2\text{Cl}_4 + \text{Cl}$  sink. Correlations



**Figure 3.** As Figure 2 but for  $\text{C}_2\text{Cl}_4$ . Results are shown for simulations with (dashed line) and without (solid line) the  $\text{C}_2\text{Cl}_4 + \text{Cl}$  sink reaction.



between regions were analyzed in the same manner as for  $\text{CH}_2\text{Cl}_2$ , and a strong negative covariance was found between the Temperate and Tropical Asia regions. Therefore, we also employ the same “Combined Asia” for the ensuing discussion. Based on this inversion setup, global total  $\text{C}_2\text{Cl}_4$  emissions decreased from  $141 \pm 14 \text{ Gg yr}^{-1}$  in 2007 to  $106 \pm 12 \text{ Gg yr}^{-1}$  in 2017. Both are significant reductions compared to our prior estimate of  $207 \text{ Gg yr}^{-1}$ . Also shown in Figure 3 are (a) emission estimates prepared using a 12-box model as reported by Engel et al. (2018) and (b) regional bottom-up estimates of  $\text{C}_2\text{Cl}_4$  emissions commissioned for this work (Table S1). The 12-box model results show global  $\text{C}_2\text{Cl}_4$  emissions decreasing from  $95\text{--}199 \text{ Gg yr}^{-1}$  in 2007 to  $66\text{--}160 \text{ Gg yr}^{-1}$  in 2016. The ranges in these values reflect the two different observational datasets used to force the model. Our posterior emissions show similar absolute decreases, from  $141 \pm 14$  to  $104 \pm 10 \text{ Gg yr}^{-1}$  over the same period. Similarly, the relative decrease in global  $\text{C}_2\text{Cl}_4$  emissions (2007–2016) is 24% (12-box model average) and 26% (this work), and the mean annual differences ( $\pm 1\text{SD}$ ) between our total emissions and the 12-box model AGAGE and NOAA estimates are  $26 \pm 6$  and  $13 \pm 7 \text{ Gg yr}^{-1}$ , respectively.

As described above, there is very good agreement between global  $\text{C}_2\text{Cl}_4$  emissions derived in this work and those from Engel et al. (2018), which used primarily the same observations, though analyzed with a simpler (12-box) model.

However, there are clear discrepancies between our regional posterior emissions and the regional bottom-up estimates from NSA shown in Figure 3. First, our posterior results for Europe show declining emissions over the study period. While a decline is consistent with the bottom-up data, the magnitude of emissions is not in agreement, with the inversion showing lower  $\text{C}_2\text{Cl}_4$  emissions in both 2007 and 2016 (in 2016 by a factor of  $\sim 2.2$  lower). Second, our inversion produces far lower Asian emissions than implied from the bottom-up data. The latter show an increase in Asian  $\text{C}_2\text{Cl}_4$  emissions from  $66 \text{ Gg yr}^{-1}$  in 2007 to  $108 \text{ Gg yr}^{-1}$  in 2016. For comparison, our posterior combined Asian emissions in the same years are  $40 \pm 12$  and  $41 \pm 9 \text{ Gg yr}^{-1}$ , respectively. The bottom-up Asian 2016 estimate ( $108 \text{ Gg yr}^{-1}$ ) is larger than the 2016 global total emissions calculated from both our inversion and from the average of the 12-box model estimates (Engel et al., 2018). Better agreement is found for North American emissions (Figure 3).

Unlike for  $\text{CH}_2\text{Cl}_2$ , tropospheric loss of  $\text{C}_2\text{Cl}_4$  via Cl radicals (in addition to OH oxidation) can be a significant sink, although its magnitude is not well constrained as the concentration of tropospheric Cl radicals is uncertain. Its inclusion in global models has been shown to lead to better agreement with  $\text{C}_2\text{Cl}_4$  observations, particularly in the upper troposphere (e.g., Hossaini et al., 2019). The main inversion results discussed above did not consider this sink, nor did the 12-box model estimates (Engel et al., 2018). A second inversion was performed that did include this additional  $\text{C}_2\text{Cl}_4$  sink. The posterior results from that inversion are presented in Table S3 and shown with dashed lines in Figure 3. It is evident that inclusion of the Cl atom sink for  $\text{C}_2\text{Cl}_4$  significantly changes the predicted global total  $\text{C}_2\text{Cl}_4$  emissions. As would be expected, emissions are larger in the presence of an additional atmospheric loss process. For example, we estimate global total  $\text{C}_2\text{Cl}_4$  emissions of  $106 \pm 12 \text{ Gg yr}^{-1}$  in 2017 without the Cl sink (Table 4) and  $162 \pm 12 \text{ Gg yr}^{-1}$  with it (Table S3, i.e., 53% larger). On a regional basis, Asian emissions provide the bulk of this increase, with 90% larger combined Asian  $\text{C}_2\text{Cl}_4$  in 2017 when the  $\text{C}_2\text{Cl}_4 + \text{Cl}$  sink is included compared to without it. Inclusion of the sink reduces the discrepancy between our posterior Asian emissions and the NSA bottom-up estimates (Figure 3(b)), though our emissions are still lower in the present day. Better agreement is also obtained for Europe, while North American emissions are broadly unchanged. It is expected that agreement between our posterior  $\text{C}_2\text{Cl}_4$  emissions and the 12-box model are poorer in absolute magnitude when the Cl sink is included, as this sink is absent in the 12-box model study. However, we note that the trends remain similar.

### 4.3. Posterior Errors

Our inversion procedure calculates the error in the posterior emissions from the terms in equation 1, using the relationship in equation 6:

$$\text{Posterior error matrix, } \mathbf{A} = [\mathbf{G}^T \cdot \mathbf{R}^{-1} \cdot \mathbf{G} + \mathbf{B}^{-1}]^{-1} \quad (6)$$

To find the regional emission error, the square root of the leading diagonal elements of  $\mathbf{A}$  is taken.

Tables 3 and 4 show the regional posterior errors for  $\text{CH}_2\text{Cl}_2$  and  $\text{C}_2\text{Cl}_4$ , respectively. The percentage reductions between the prior error and the posterior errors are also given in these tables. A large error reduction implies more confidence in the posterior solution, and for  $\text{CH}_2\text{Cl}_2$ , the largest reductions occur for the main emitting regions where observations are available, for example, Temperate North America (89% error reduction for  $\text{CH}_2\text{Cl}_2$  in 2017), Temperate Asia (95%), and Europe (77%). Note that these values should be considered in tandem with the posterior errors themselves. For example, although Europe has a significant error reduction (e.g., in 2017 a reduction of 77%), this results in a posterior error of  $\pm 11 \text{ Gg yr}^{-1}$ , which is a 15% of the actual posterior  $\text{CH}_2\text{Cl}_2$  emission from this region. The inverse is true for Temperate Asia, where relatively large posterior emissions ( $591 \text{ Gg yr}^{-1}$  in 2017) and a large error reduction (95%) lead to a very small (5%) error in the posterior emission. The  $\text{C}_2\text{Cl}_4$  errors generally show a similar behavior, with the largest prior versus posterior error reduction achieved for the main industrial regions where large emissions are derived.

A small error reduction corollary is a sign of less confidence in the posterior emissions. For both compounds, these generally apply to regions that are minimally constrained by local observations, such as Africa and tropical Latin America. Fortunately, as it is assumed that these regions do not contribute much to the total global emissions, relatively large uncertainty in their regional posterior emissions have minimal impact on our findings. That said, we highlight Boreal Asia, a region that is a small net source in our prior emissions ( $6.8 \text{ Gg yr}^{-1}$  for  $\text{CH}_2\text{Cl}_2$ ,  $1.8 \text{ Gg yr}^{-1}$  for  $\text{C}_2\text{Cl}_4$ ) but becomes a net sink for both compounds in our posterior solution. In 2017, our posterior emissions for Boreal Asia are  $-22 (\pm 6) \text{ Gg yr}^{-1}$  for  $\text{CH}_2\text{Cl}_2$  and  $-2.7 (\pm 1.6) \text{ Gg yr}^{-1}$  for  $\text{C}_2\text{Cl}_4$ . For this region, the percentage reductions between the prior error and the posterior error are small (Tables 3 and 4). While some chlorocarbons are taken up by terrestrial ecosystems (e.g., Khalil & Rasmussen, 1999), no terrestrial sinks of  $\text{CH}_2\text{Cl}_2$  have been reported, and the lack of observational constraints in this region could point towards a small inversion artifact. An analysis of covariances did not reveal a strong coupling between Boreal Asia and another region.

#### 4.4. Ocean Emissions

Thus far, we have largely focused on our posterior emissions from land (in the “no ocean” inversion for  $\text{CH}_2\text{Cl}_2$ ). In this section, we examine ocean emissions. A summary of posterior  $\text{CH}_2\text{Cl}_2$  emissions with the ocean source included is given in Table S2 and can be compared to the equivalent no ocean case (Table 3). In the inversion in which net emissions from the ocean are allowed, oceanic  $\text{CH}_2\text{Cl}_2$  emissions (sum from all four ocean bands) account for  $197 \text{ Gg yr}^{-1}$  (19%) of our prior global total emission, decreasing to  $162 \text{ Gg yr}^{-1}$  (14%) in our posterior solution in 2017. The global total  $\text{CH}_2\text{Cl}_2$  emission is relatively insensitive to the inclusion of the ocean source,  $1,171 \pm 45 \text{ Gg yr}^{-1}$  (no ocean) versus  $1,166 \pm 64 \text{ Gg yr}^{-1}$  (with ocean) for 2017. However, inclusion of the ocean decreases the combined Asia posterior emissions by  $\sim 18\%$ , from  $1,045 \pm 40 \text{ Gg yr}^{-1}$  (no ocean) in 2017 to  $886 \pm 53 \text{ Gg yr}^{-1}$  (with ocean). This effect is largely explained by the inversion placing emissions of  $\text{CH}_2\text{Cl}_2$  in the tropical Northern Ocean ( $0\text{--}30^\circ\text{N}$  latitude). As was the case in the “no ocean” inversion, Figure S2 highlights increasing Asian emissions over our study period, while European and North American emissions decrease. With the ocean included, the Combined Asian emissions provide a closer match to the bottom-up NSA estimates of industrial Asian emissions (i.e., our prior) in 2016 (Figure S2). However, as mentioned later in section 4.6, there is no discernible difference in performance between the inversions with and without the ocean source, when compared with observations (even independent data), despite these changes in Asian emissions.

The geographical distribution of our posterior ocean  $\text{CH}_2\text{Cl}_2$  emissions differs significantly from the prior in the inversion that allows nonzero ocean fluxes. For example, when averaged over the entire 12-year study period, the Extratropical Northern Ocean represents a net sink of  $\text{CH}_2\text{Cl}_2$  (Table 5). In the extratropical Southern Ocean, the derived net flux is significantly lower than the prior but remains positive. Also shown in Table 5 are results from the inversion study of AGAGE and NOAA observations by Xiao (2008), who allowed nonzero fluxes from the ocean and tabulated ocean  $\text{CH}_2\text{Cl}_2$  emissions from these same latitude bands. Posterior emissions for the Tropical Northern and Extratropical Southern Oceans from our study fall within the Xiao (2008) uncertainty ranges. However, notably the

**Table 5**

Optimised  $\text{CH}_2\text{Cl}_2$  Oceanic Emissions ( $\text{Gg yr}^{-1}$ ) Derived from Positive Prior Fluxes for this Inversion, a Previous Inversion (Xiao, 2008), and Measurements from Four Observational Studies

Ocean band	This work <sup>a</sup>	Xiao (2008) <sup>b</sup>	AMT-22 Campaign <sup>c</sup>	ACCACIA-2 Campaign <sup>c</sup>	Kolusu et al. (2016) <sup>c</sup>	Oooki and Yokouchi (2011) <sup>d</sup>
Extratropical NO	$-64.4 \pm 10.9$	$3.5 \pm 3.0$	$14.7 \pm 26.9$	$-47.9 \pm 32.7$		
Tropical NO	$111.8 \pm 43.1$	$88 \pm 29$	$70.9 \pm 63.6$		$236 \pm 237$	
Tropical SO	$9.9 \pm 12.7$	$31 \pm 24$	$64.0 \pm 43.3$			$12.5 \pm 2.5$
Extratropical SO	$4.3 \pm 3.9$	$2 \pm 5$				

Note. All measurements converted into  $\text{Gg yr}^{-1}$ . Reported uncertainties for inversion calculations and campaign ocean tracks of 1SD. NO = Northern Ocean; SO = Southern Ocean.

<sup>a</sup>Twelve-year average posterior emission. <sup>b</sup>Five-year average posterior emission from 2000–2005. <sup>c</sup>Atlantic Ocean sea-to-air flux measurements, originally reported as  $\text{nmol m}^{-2} \text{ day}^{-1}$ , and converted into  $\text{Gg band}^{-1} \text{ yr}^{-1}$  for the relevant ocean latitude bands. AMT-22 campaign measurements took place in October–November 2012, ACCACIA-2 campaign measurements in July–August 2013, and Kolusu et al. (2016) measurements in April–May 2009. <sup>d</sup>Indian Ocean biogenic production from phytoplankton, reported between  $10^\circ\text{S}$  and  $40^\circ\text{S}$  in  $\mu\text{g m}^{-2} \text{ day}^{-1}$ . These measurements were taken from November 2009 to January 2010 and account for physical effects that are the likely principle source in the sea-to-air flux measurements.

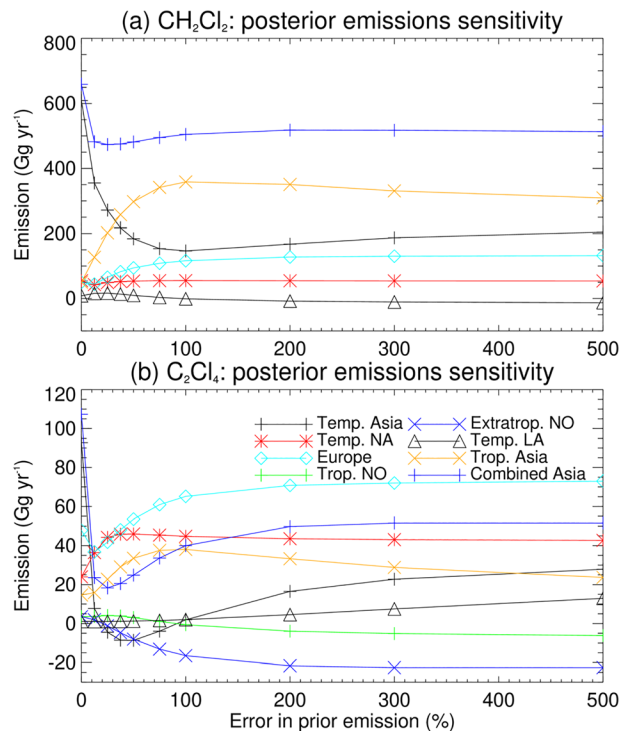
significant mean  $\text{CH}_2\text{Cl}_2$  sink we derive to the Extratropical Northern Ocean is not apparent in the Xiao (2008) study for the observations during 2000–2005, a period before the large atmospheric increase in  $\text{CH}_2\text{Cl}_2$  occurred.

It is also possible to perform a basic comparison of our posterior ocean  $\text{CH}_2\text{Cl}_2$  emissions to observed estimates based on (limited) cruise data. As noted in section 3.2.3, ocean  $\text{CH}_2\text{Cl}_2$  sea-to-air fluxes from the AMT-22 and ACCACIA-2 ship cruises (see tracks in Figure 1(b)) were derived based on concentration measurements without consideration of the potential influence of physical effects on the derived fluxes.

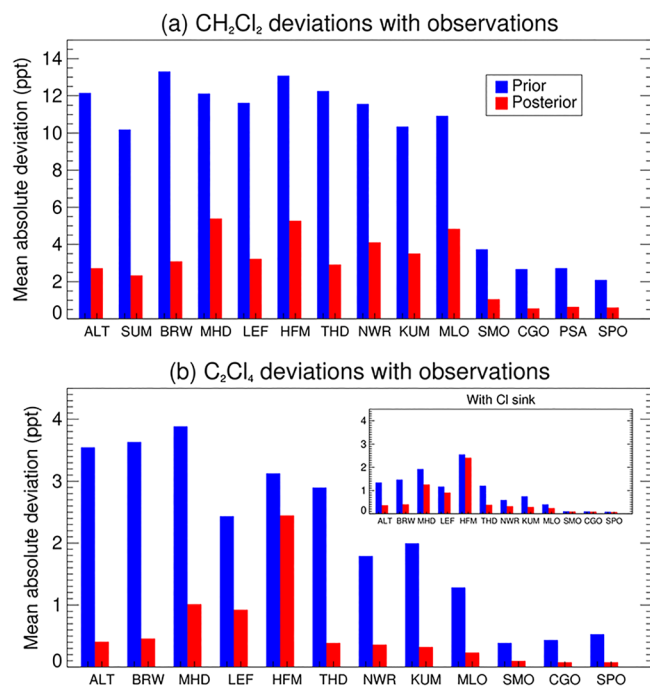
These cruises sampled in three out of the four ocean bands used in our inversion and the integrated flux from each ocean band is presented in Table 5. Broadly, the sign of the emissions agrees in all three regions observed; however, it is important to note that this comparison is potentially confounded by considering annual average inversion results to cruise data that we expect could be influenced by seasonally varying sea-to-air fluxes based on seasonal changes in solubility and atmospheric concentrations (Moore, 2004).

Xiao (2008) reports a large seasonal cycle in ocean emissions, and for the  $30^\circ\text{--}90^\circ\text{N}$  region, there is a maximum of approximately  $20 \text{ Gg yr}^{-1}$  in the summer and a minimum of  $-10 \text{ Gg yr}^{-1}$  in the winter. We note that ACCACIA-2 took place during July–August, and reported an average flux of  $-48 \text{ Gg yr}^{-1}$ , which is contrary to what the seasonal cycle states. However, neither our inversion nor that of Xiao (2008) can resolve between  $60^\circ$  and  $80^\circ\text{N}$ , where ACCACIA-2 took place. AMT-22, measuring from  $30^\circ$  to  $50^\circ\text{N}$  in autumn, calculates an average flux of  $15 \text{ Gg yr}^{-1}$ , which is close to the average autumnal values from Xiao (2008). In the original RCEI estimates, none of the ocean tracks used to infer  $\text{CH}_2\text{Cl}_2$  fluxes took place above  $60^\circ\text{N}$  (Khalil et al., 1999); therefore, the large summer sink observed by ACCACIA-2 at these very high latitudes could be evidence for the significant interregional variation we see in our inversion results.

Also presented in Table 5 is an estimate of an ocean  $\text{CH}_2\text{Cl}_2$  source based on measurements in the tropical Atlantic from April to May in 2009 (Kolusu et al., 2016). This source is very uncertain and is generally much higher than any single ocean band emission derived from the inversion. As with the cruises previously discussed, these data do not necessarily support marine  $\text{CH}_2\text{Cl}_2$  production because of the strong potential for changes in sea-to-air flux related to seasonality in solubility and



**Figure 4.** Summary of results testing the sensitivity of posterior emissions to the assumed error in the prior emissions for (a)  $\text{CH}_2\text{Cl}_2$  and (b)  $\text{C}_2\text{Cl}_4$ . Results are shown indicatively for the year 2007 and for seven different regions (five for  $\text{CH}_2\text{Cl}_2$ ), including a Combined Asian result. Note that a 0% prior emission error equates to the prior emissions. For  $\text{CH}_2\text{Cl}_2$ , results are shown for the no ocean inversion scenario.



**Figure 5.** Mean absolute deviation (ppt) between modeled and observed (a)  $\text{CH}_2\text{Cl}_2$  and (b)  $\text{C}_2\text{Cl}_4$ , at NOAA sites. The deviations are averages calculated from monthly mean data over the study periods (2006–2017 for  $\text{CH}_2\text{Cl}_2$  and 2007–2017 for  $\text{C}_2\text{Cl}_4$ ) and are shown for model output generated using the prior emissions and the posterior emissions. The  $\text{C}_2\text{Cl}_4$  + Cl sink comparisons are inset in (b). For  $\text{CH}_2\text{Cl}_2$ , results are generated using the posterior emissions from the no ocean inversion scenario.

are progressively increased, each of the inversion regions is given a greater degree of freedom to reach a target. There are several features apparent in Figure 4 that warrant attention. For  $\text{CH}_2\text{Cl}_2$ , Temperate North America and Europe are examples of regions whose emission magnitudes are insensitive to the uncertainty assumed for the prior when it is above  $\sim 50\%$ . The Temperate and Tropical Asian regions were found to vary more with the assumed prior uncertainty, and the derived emissions are slightly anticorrelated. However, our Combined Asia region is insensitive to prior error when assumed to be equal or larger than  $\pm 100\%$  (Figure 4(a)). Temperate Latin America is an example of a region that reaches its optimum emission value at a larger emissions error than  $\pm 100\%$ , and at further increased error drifts negatively. As previously noted, our posterior  $\text{CH}_2\text{Cl}_2$  emissions from this region are negative, possibly reflecting a small inversion artifact due to the lack of data, but are also very small and thus of limited global importance:  $-0.96$  ( $\pm 5.62$ )  $\text{Gg yr}^{-1}$  in 2017. A similar sensitivity analysis to prior errors for the with ocean  $\text{CH}_2\text{Cl}_2$  case was also performed (Figure S3).

For  $\text{C}_2\text{Cl}_4$ , the posterior emissions are generally stable beyond  $\pm 100\%$  prior error. Again, Temperate and Tropical Asia have a small tendency to drift towards each other, though likewise, the Combined Asia region is insensitive to prior error, justifying our initial assumption, but only at errors beyond  $\pm 200\%$  (Figure 4(b)).

#### 4.6. Posterior Versus Prior Emissions Performance

With prior and posterior  $\text{CH}_2\text{Cl}_2$  and  $\text{C}_2\text{Cl}_4$  emissions calculated, their performance can be evaluated by comparing modeled mixing ratios obtained with each to observations. We first focus on a single year (2016) and consider how well the posterior emissions reproduce background NOAA and AGAGE surface observations. Both the NOAA and AGAGE data used in these comparisons were assimilated by the inversion (i.e., to construct the posterior emissions). Comparisons to independent observational

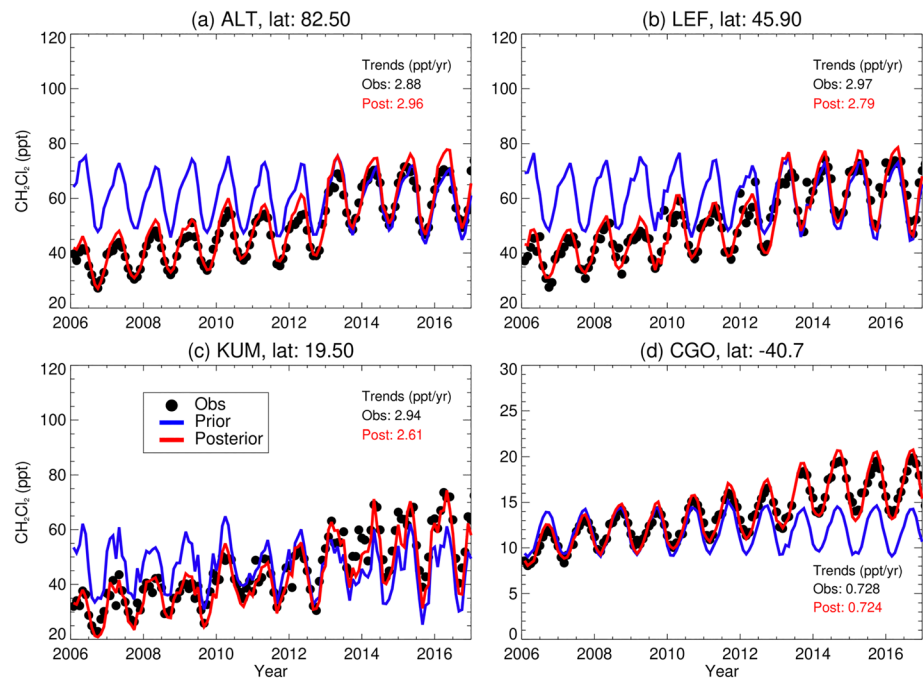
atmospheric concentrations. Rather, this data simply highlights an ocean region where a significant quantity of  $\text{CH}_2\text{Cl}_2$  may enter the atmosphere (at least during the period of the campaign of April–May). Due to the large uncertainty, it is difficult to say how representative the observations are, as the study showed there was a strong latitudinal gradient, especially when crossing the Equator (Kolusu et al., 2016). Lastly, in Table 5, are estimated biogenic  $\text{CH}_2\text{Cl}_2$  emissions reported by Ooki and Yokouchi (2011) based on data collected in the tropical Southern Indian Ocean ( $10$ – $40^\circ\text{S}$ ). Factoring in the uncertainty ranges, our inversion emissions from the tropical Southern Ocean are comparable.

In summary, our posterior  $\text{CH}_2\text{Cl}_2$  net ocean source (11% of the global total from all sources in 2017 or  $125 \text{ Gg yr}^{-1}$ ) is comparable to previous inversion estimates and to a small set of available oceanic observations. However, while the total source magnitude is comparable, the distribution shifts the majority of the emissions into the Tropical NO region and very few emissions into the Tropical SO region. This distribution is likely a consequence of the large driving force of Combined Asian land emissions; practically, it is plausible that the tropical distribution is more even, as observations suggest. For  $\text{C}_2\text{Cl}_4$ , the posterior ocean source is negligible and often negative (Table 4). For both compounds, our inversion does not distinguish between ocean re-emission and, if it exists, “true” marine production.

#### 4.5. Sensitivity to Prior Uncertainty

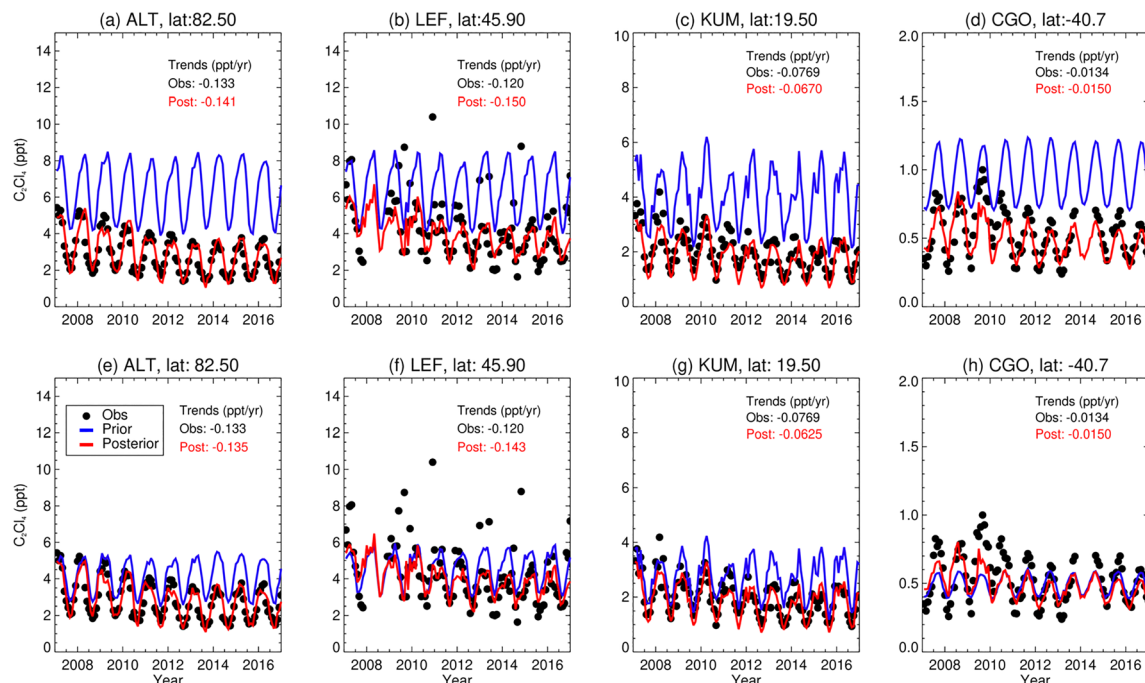
Our prior emission errors were set to  $\pm 100\%$  for all regions (section 3.4), and we tested the sensitivity of our posterior emissions to this value. Figure 4 illustrates this by presenting posterior  $\text{CH}_2\text{Cl}_2$  and  $\text{C}_2\text{Cl}_4$  emissions for seven of the most important inversion regions under different prior error assumptions applied to each region simultaneously. Note that for this analysis, a 0% error simply represents the prior emission. As errors



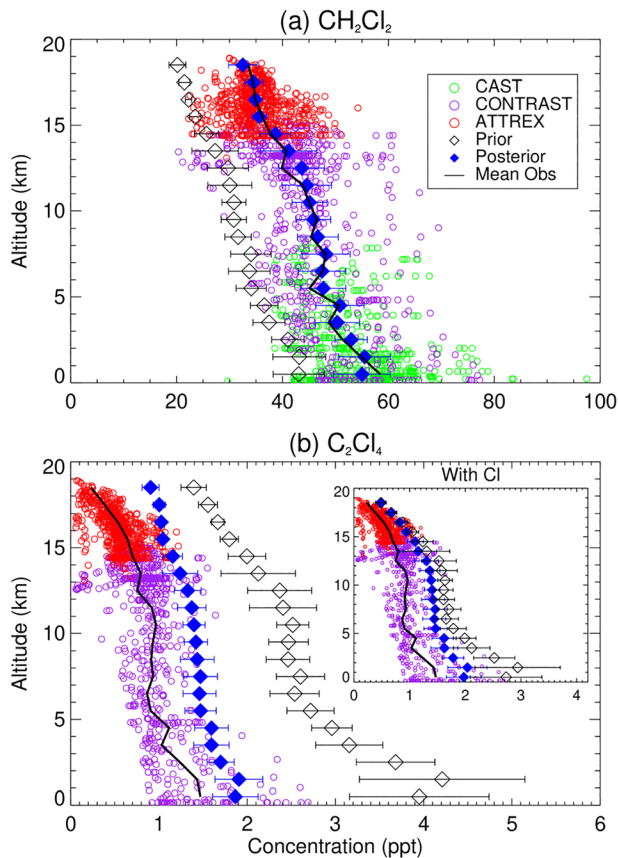


**Figure 6.** Comparison of modeled monthly mean  $\text{CH}_2\text{Cl}_2$  mixing ratio (ppt) versus NOAA observations (2006–2017) at stations (a) ALT, (b) LEF, (c) KUM, and (d) CGO. Each panel contains model output based on the prior (blue) and posterior (red) emissions, with annual trends ( $\text{ppt yr}^{-1}$ ) in the model (posterior) and observations annotated. For  $\text{CH}_2\text{Cl}_2$ , results are generated using the posterior emissions from the no ocean inversion scenario.

data are considered in section 4.7. Modeled monthly mean  $\text{CH}_2\text{Cl}_2$  (“no ocean” inversion) and  $\text{C}_2\text{Cl}_4$  are compared to NOAA measurements in Figures S4 and S5, respectively. For  $\text{CH}_2\text{Cl}_2$ , Figure S4 reveals generally very good agreement between the model (posterior emissions) and the observations. At several sites, particularly those at midlatitude and high latitude in the Northern Hemisphere (NH),



**Figure 7.** Comparison of modeled monthly mean  $\text{C}_2\text{Cl}_4$  mixing ratio (ppt) versus NOAA observations (2007–2017) at the same stations as in Figure 6. Panels (a)–(d) are comparisons without the  $\text{C}_2\text{Cl}_4 + \text{Cl}$  reaction, and (e)–(h) are with the reaction.



**Figure 8.** Modeled versus observed vertical profiles of (a)  $\text{CH}_2\text{Cl}_2$  and (b)  $\text{C}_2\text{Cl}_4$  volume mixing ratio (ppt) during the 2014 CAST/CONTRAST/ATTREX field missions in the West Pacific (see section 3.2). Model output has been averaged in 1 km vertical bins and is shown for both the prior and posterior emissions. Note that these aircraft data are “independent” in that they were not used in the inversion to produce the posterior emissions. The  $\text{C}_2\text{Cl}_4$  + Cl sink data are inset in (b). For  $\text{CH}_2\text{Cl}_2$ , results are generated using the posterior emissions from the no ocean inversion scenario.

to below ~5 ppt at most NH sites and below ~0.9 ppt at all SH sites (Figure 5(a)). This is not entirely unexpected given that these observations were included in the inversion itself. However, overall, the reduction of the prior deviations by roughly 60% in the posterior indicates that the inversion procedure has been successful.  $\text{C}_2\text{Cl}_4$  is equally successful, also with an average deviation reduction of roughly 80% and 40% when the  $\text{C}_2\text{Cl}_4$  + Cl sink is included (Figure 5(b)). Evident from the same figure, the magnitude of these improvements is not overly sensitive to inclusion of the +Cl sink, but the decreased reduction is due to the reduction in the prior deviations.

In addition to the above deviations, it is important that the time-dependent posterior emissions adequately capture trends. The four NOAA sites shown in Figure 6 for  $\text{CH}_2\text{Cl}_2$  are a selection from various geographical locations (Table 1): a high latitude NH site (ALT), a midlatitude NH site (LEF), a tropical site (KUM), and an SH site (CGO). The posterior  $\text{CH}_2\text{Cl}_2$  model output is far better at matching with the observations over the 12-year period compared to the prior model output. This is especially true of the earlier parts of our study period given that our prior emissions (for the main industrialized regions) were based on bottom-up data from 2016. Annotated in Figure 6 are the modeled (posterior) and observed  $\text{CH}_2\text{Cl}_2$  trends over the 2006–2017 period, calculated using a simple least squares regression. The modeled and observed trends are 3.0 and 2.9 ppt yr<sup>−1</sup> at ALT, 2.8 and 3.0 ppt yr<sup>−1</sup> at LEF, 2.6 and 2.9 ppt yr<sup>−1</sup> at KUM, and 0.7 and 0.7 ppt yr<sup>−1</sup> at CGO and thus are in excellent agreement. Despite the geographical range of the three NH sites, similar trends, roughly 3 ppt yr<sup>−1</sup>, are found.

the prior and posterior emissions perform similarly. However, a clear improvement is obtained when using the posterior emissions in the tropics and in the Southern Hemisphere (SH). For  $\text{C}_2\text{Cl}_4$ , posterior improvements at NOAA sites are more striking. Figure S5 shows that our prior  $\text{C}_2\text{Cl}_4$  emissions were too large, leading to a significant overestimation of observed mole fractions.

Similar comparisons but for AGAGE observations (converted to the NOAA calibration scale) are shown in Figures S6 and S7. These comparisons are also for the year 2016, with the exception of the Gosan (GSN) site in South Korea (2015). For  $\text{CH}_2\text{Cl}_2$ , the measurements are better reproduced using our posterior emissions at most sites, except for ZEP, though as noted above, the differences between prior and posterior results are relatively small in the NH. For  $\text{C}_2\text{Cl}_4$ , again, the posterior emissions lead to much better agreement between the model and observations. A notable feature for  $\text{CH}_2\text{Cl}_2$  in Figure S6, apparent when using both the prior and posterior emissions, is the model overestimation of baseline  $\text{CH}_2\text{Cl}_2$  observations at Gosan. This site is heavily influenced by several large nearby sources, and the mismatch errors in the inversion are particularly large. Significantly improved agreement to the Gosan data is, however, obtained when the “raw” measurements are used to construct monthly means (i.e., without filtering out pollution events). Such events are inherently included in the model monthly means.

The comparisons discussed above focused on a single recent year. A more informative approach is to consider the performance of the posterior emissions over the entire study period. To quantify this performance, we calculate the mean absolute deviation (equation (7)) over the full study periods at each available NOAA site (14 for  $\text{CH}_2\text{Cl}_2$  and 12 for  $\text{C}_2\text{Cl}_4$ ) based on monthly means:

$$\text{Mean absolute deviation} = \frac{\sum_{\text{months}} |\text{model} - \text{observation}|}{n_{\text{months}}}. \quad (7)$$

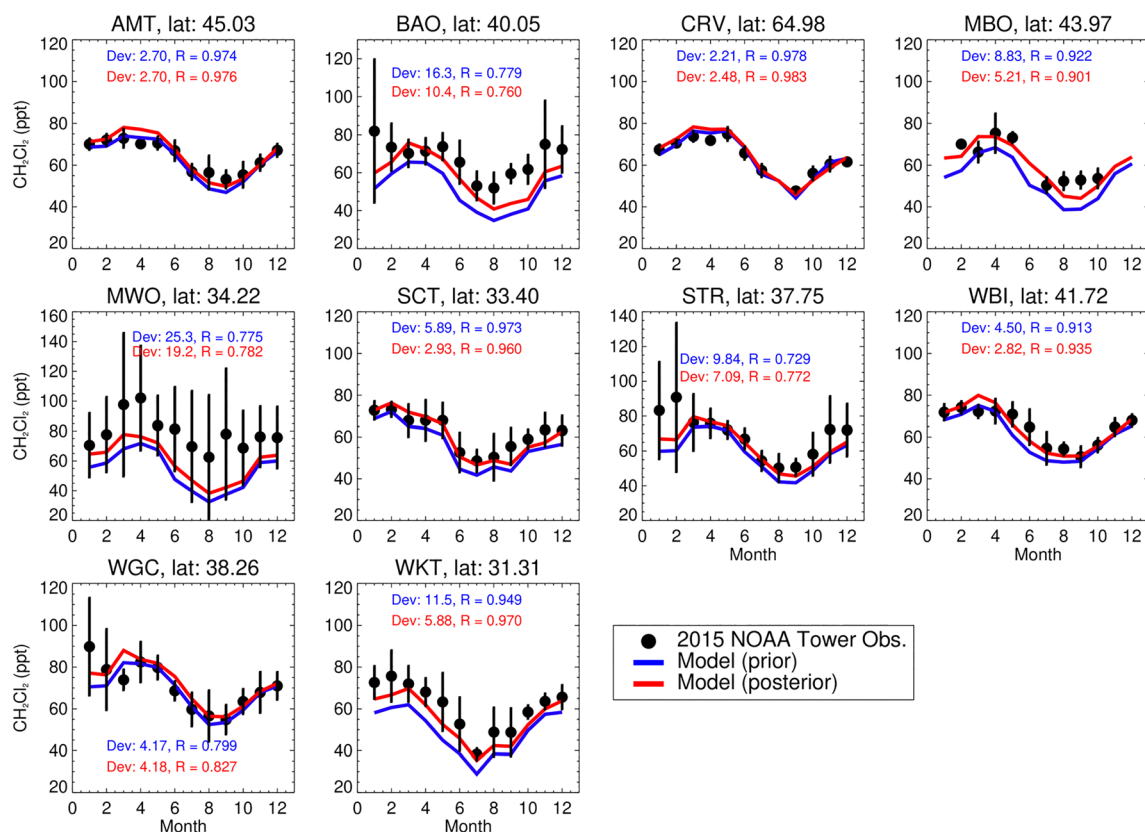
For  $\text{CH}_2\text{Cl}_2$ , the posterior emissions provide much improved agreement to the observations at all sites, reducing the model/observation deviation

A similar analysis for  $C_2Cl_4$  was performed at the same four sites, with again good agreement between the model (with posterior emissions) and the observations (Figure 7). The modeled and observed trends without the  $C_2Cl_4 + Cl$  sink reaction included are  $-0.14$  and  $-0.13$  ppt  $yr^{-1}$  at ALT,  $-0.15$  and  $-0.12$  ppt  $yr^{-1}$  at LEF,  $-0.07$  and  $-0.08$  ppt  $yr^{-1}$  at KUM, and  $-0.02$  and  $-0.01$  ppt  $yr^{-1}$  at CGO. In addition, Figure 7 includes the  $Cl$  sink of  $C_2Cl_4$ . The trends are slightly improved in the posterior for three of the sites (except KUM), and the figure shows the stark contrast between the two prior model outputs. The addition of the  $Cl$  sink leads to a decreased lifetime of  $C_2Cl_4$  (Hossaini et al., 2019), and therefore, prior concentrations are decreased. The inversion compensates for this by increasing posterior emissions, as shown in Figure 3. Despite two very different prior positions, the two almost identical  $C_2Cl_4$  posterior outputs (in Figures 5(b) and 7) indicate how effective the inversion process can be.

Note that for  $CH_2Cl_2$ , the discussion above has focused on the “no ocean” inversion. The posterior modeled  $CH_2Cl_2$  mixing ratios from the with ocean inversion (not shown) are found to be almost identical; thus, the performance of the two inversions is very similar. This implies that it cannot be concluded (or excluded) that a significant ocean  $CH_2Cl_2$  source exists from this analysis.

#### 4.7. Independent Observations

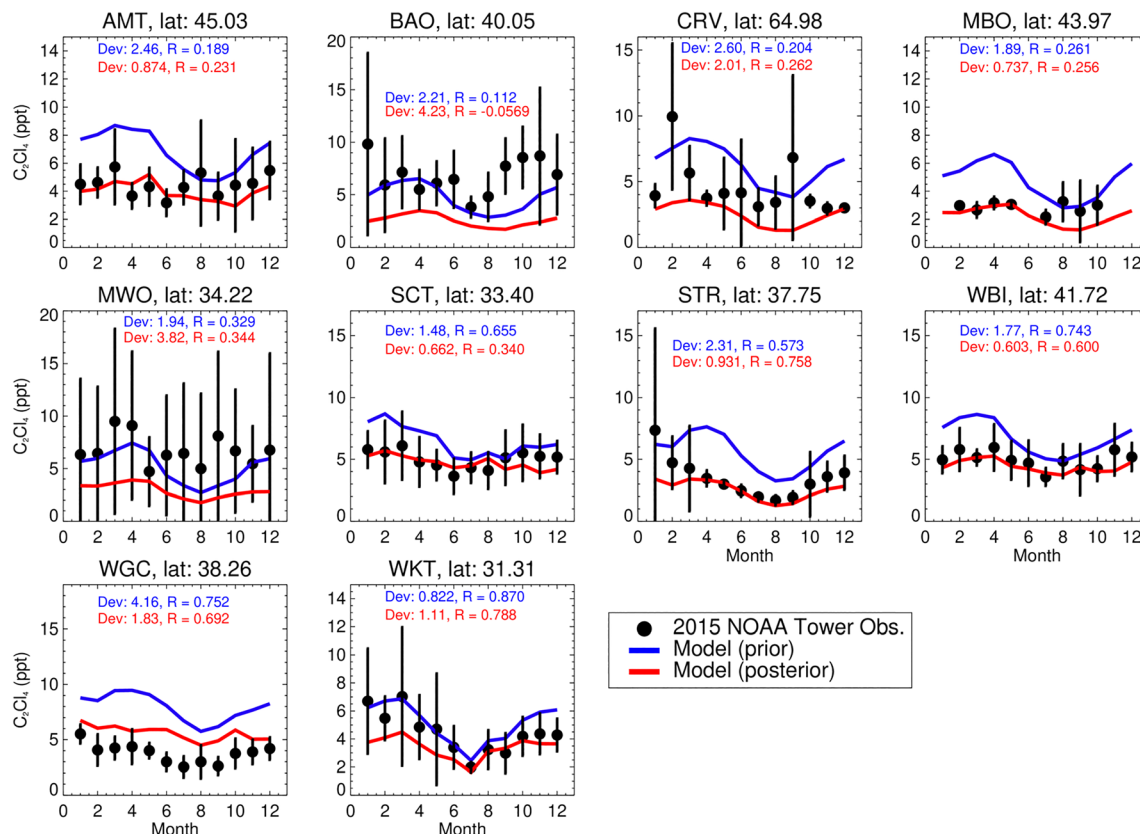
In the previous section, we compared modeled  $CH_2Cl_2$  and  $C_2Cl_4$  mixing ratios, generated using our posterior emissions, to observations used in the inversion itself. Here, we examine independent observations, first considering aircraft measurements during the 2014 CAST, CONTRAST, and ATTREX missions over the tropical West Pacific (section 3.2). Modeled and observed vertical profiles (surface to  $\sim 20$  km) are displayed in Figure 8. Throughout the vertical extent of these profiles, there is (a) near-perfect agreement between modeled  $CH_2Cl_2$  using the posterior emissions (“no ocean” inversion) compared to the observations and (b) a



**Figure 9.** Comparison between monthly mean NOAA tall tower observations of  $CH_2Cl_2$  (ppt) in 2015 (independent observations) and modeled values obtained using the prior and posterior emissions. The vertical bars on the observations indicate  $\pm 1SD$  of all measurements acquired at that site during that month. Annotated for each site are the average annual deviations (Dev) between the two model outputs and the observations, and the correlation coefficient,  $R$ . Results are generated using the posterior emissions from the no ocean inversion scenario.

significant improvement over the prior. The latter presumably reflects the larger Asian emissions in our posterior model occurring in the vicinity to the measurement campaigns. For  $C_2Cl_4$  (CONTRAST and ATTREX only), Figure 8(b) shows results from the model both with and without the Cl sink. In both cases, the posterior emissions outperform the prior. The Cl sink case matches with the observations more effectively at higher altitudes, whereas both cases are similar towards the surface. On average, the model output overestimates  $C_2Cl_4$  by 0.53 ppt for the no Cl sink and by 0.47 ppt for the Cl sink case. This overestimation could be caused by the large Combined Asia emissions, which heavily influence these observations.

The second test for our inversion is from the independent network of NOAA tall tower sites (USA-based, e.g., see Figure 1). At each of the 10 sites where  $CH_2Cl_2$  observations are available, the posterior model provides a reasonable representation of the measurements in 2015 (Figure 9). Annotated on this figure are the mean absolute deviations at each site between the model (with prior and with posterior emissions) versus the observations, in that year. At most sites, the posterior model outperforms the prior, but generally, only small improvements in the  $CH_2Cl_2$  average deviation are achieved. The correlations between model and observations (also annotated) also show a consistent improvement for the posterior. At certain sites, large standard deviations on the monthly mean observations coincide with poor model-measurement agreement, in some months. For example, the proximity of the MWO site to large urban areas may partly explain why the monthly mean observations are consistently larger than the model outputs, and for the other instances where large standard deviations occur, the model outputs lie at the lower range of the observations. For  $C_2Cl_4$ , a similar analysis was performed and reveals a more varied picture (Figure 10). As prior emissions generally overestimate  $C_2Cl_4$ , this leads to large improvements in the posterior output at most sites. However, at some sites the modeled-observed  $C_2Cl_4$  deviations are larger for the posterior compared to the prior (e.g., BAO and MWO). As for  $CH_2Cl_2$ , this is largely due to the close proximity of substantial sources influencing observations that are likely not well captured by the model. However, at most sites, we note that our posterior model output lies within the observed variability. We further note that as with



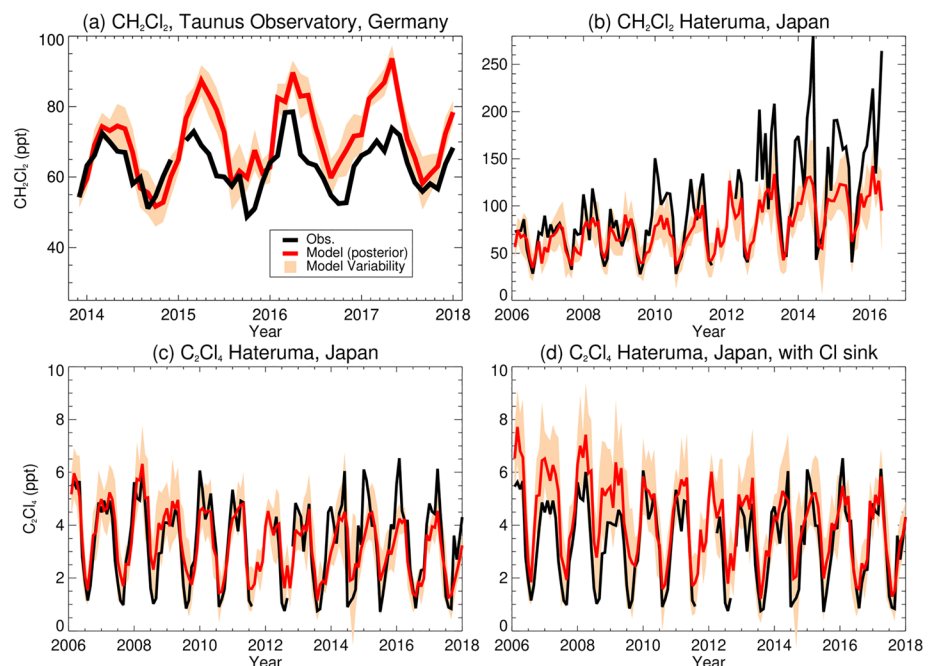
**Figure 10.** As Figure 9 but for  $C_2Cl_4$ . Note that the model results here did not include the  $C_2Cl_4$  + Cl sink.



all comparisons of relatively coarse global-scale models with point-based observations, sampling errors in the model can affect such comparisons.

For Europe, beyond the NOAA and AGAGE observational data used in the inversion, there are few long-term surface measurements of  $\text{CH}_2\text{Cl}_2$  and  $\text{C}_2\text{Cl}_4$ . However, a recent network addition is the establishment of a  $\text{CH}_2\text{Cl}_2$  record at the Taunus Observatory ( $50.22^\circ\text{N}$ ,  $8.44^\circ\text{E}$ , at 825 m) in central Germany (Schuck et al., 2018). Taunus reports  $\text{CH}_2\text{Cl}_2$  measurements using the AGAGE SIO-14 scale (here converted to the NOAA scale). Figure 11(a) compares modeled and observed monthly mean  $\text{CH}_2\text{Cl}_2$  at the Taunus site between 2014 and 2017. The agreement between the model (with posterior emissions) and observations is reasonable, with the shape of the seasonal cycle generally well captured. The model does overestimate  $\text{CH}_2\text{Cl}_2$  at this particular site during some periods. Since sampling errors in the model could cause this overestimation, Figure 11 also investigates model variability by including the standard deviation between the eight neighboring model grid cells, using equation 4. At this particular site, variability introduced from neighboring model grid cells is reasonably small and does not fully rationalize the small discrepancies between the model and observations, including apparent slight offsets in the seasonal cycle. However, we note that at other NH sites—including from the USA-based tall tower network—the  $\text{CH}_2\text{Cl}_2$  seasonal cycle is very well captured (e.g., Figures S4, S6, and 9).

Also included in Figure 11 are comparisons of modeled  $\text{CH}_2\text{Cl}_2$  and  $\text{C}_2\text{Cl}_4$  using our posterior emissions to baseline measurements obtained from the AGAGE-affiliated site in Hateruma, Japan (HAT,  $24.1^\circ\text{N}$ ,  $123.8^\circ\text{E}$ , at 46.5 m). Hateruma is calibrated with the NIES-08 scale and for  $\text{CH}_2\text{Cl}_2$  can be converted to the AGAGE SIO-14 scale by a factor of  $1.066 \pm 0.008$ . The conversion factor between the NIES-08 scale and AGAGE's NOAA-2003B scale for  $\text{C}_2\text{Cl}_4$  is  $0.994 \pm 0.010$ . Given that the number of monitoring stations in Asia is limited and that this is where the largest Cl-VSLs emissions are predicted to occur, these independent comparisons are particularly useful. For  $\text{CH}_2\text{Cl}_2$ , model-observation agreement is generally good, though in the most recent years of our study period, the model underestimates observed  $\text{CH}_2\text{Cl}_2$  mixing



**Figure 11.** A comparison between modeled and observed monthly mean mixing ratios (ppt) of  $\text{CH}_2\text{Cl}_2$  and  $\text{C}_2\text{Cl}_4$  at independent measurement sites. (a)  $\text{CH}_2\text{Cl}_2$  at the Taunus Observatory in central Germany (Schuck et al., 2018), (b)  $\text{CH}_2\text{Cl}_2$  at Hateruma, Japan, (c)  $\text{C}_2\text{Cl}_4$  at Hateruma (no  $\text{C}_2\text{Cl}_4$  + Cl sink), and (d)  $\text{C}_2\text{Cl}_4$  at Hateruma (including +Cl sink). The model output was generated using the posterior emissions from the inversion. In the case of  $\text{CH}_2\text{Cl}_2$ , results from the no ocean inversion scenario are shown (see main text). Model variability (light orange) was calculated from the standard deviation of the surrounding eight model grid cells (equation 4). All observations here are calibrated to NOAA scales.

ratios, particularly in winter. This wintertime disparity could represent a combination of uncaptured seasonality in  $\text{CH}_2\text{Cl}_2$  emissions and underestimated model  $\text{CH}_2\text{Cl}_2$  lifetime and is apparent even when considering the model sampling/mismatch issues noted above. For  $\text{C}_2\text{Cl}_4$ , comparisons to the Hateruma data are shown for the model with and without the Cl sink used to construct the posterior emissions. Both cases lead to adequate model-measurement agreement, though including the Cl sink provides far better agreement in the most recent years.

## 5. Summary and Conclusions

Combining long-term surface observations and a chemical transport model, we have performed a global-scale synthesis inversion to (a) constrain regional/global emissions of  $\text{CH}_2\text{Cl}_2$  and  $\text{C}_2\text{Cl}_4$ , (b) investigate emission trends over the 2006 to 2017 period, and (c) produce a set of evaluated emission inventories for future global modeling studies. Our main findings are the following:

- For an inversion in which only industrial  $\text{CH}_2\text{Cl}_2$  emissions are considered, we estimate that global  $\text{CH}_2\text{Cl}_2$  emissions increased from  $637 \pm 37 \text{ Gg yr}^{-1}$  in 2006 to  $1,171 \pm 45 \text{ Gg yr}^{-1}$  in 2017, with reasonably good agreement between our results and those reported in the recent WMO Ozone Assessment with a simplified model and similar data as input (Engel et al., 2018). This increase is largely attributed to an increase in Asian emissions, while relatively small European and North American emissions decrease over the same period. This geographical shift in the emission distribution is broadly consistent with studies that have highlighted the growing importance of major Asian economies as a  $\text{CH}_2\text{Cl}_2$  source (e.g., Leedham Elvidge et al., 2015; Feng et al., 2018). In 2017, we estimate Asian emissions accounted for 89% ( $1,045 \pm 40 \text{ Gg yr}^{-1}$ ) of total  $\text{CH}_2\text{Cl}_2$  emissions, up from 68% ( $431 \pm 32 \text{ Gg yr}^{-1}$ ) in 2006.  $\text{CH}_2\text{Cl}_2$  emissions from Europe and North America combined represented 9% of the global total in 2017, down from 29% in 2006. Decreases in these regions may, in part, reflect recent concerns over the compound's toxicity in consumer products.
- For an inversion in which both oceanic and industrial  $\text{CH}_2\text{Cl}_2$  sources are considered, we estimate global  $\text{CH}_2\text{Cl}_2$  emissions of  $1,166 \pm 64 \text{ Gg yr}^{-1}$  in 2017, that is, very similar to the no ocean case. However, including the ocean, source reduces the estimate of 2017 Asian emissions from  $1,045 \pm 40$  to  $886 \pm 53 \text{ Gg yr}^{-1}$  (a reduction of 15%). A large portion of this difference is explained by the inversion placing a significant emission of  $\text{CH}_2\text{Cl}_2$  in the tropical Northern Ocean ( $0\text{--}30^\circ\text{N}$  latitude). Averaged over our study period, oceanic  $\text{CH}_2\text{Cl}_2$  emissions from this latitude band are approximately  $123 (\pm 45) \text{ Gg yr}^{-1}$ , which is comparable to  $88 (\pm 29) \text{ Gg yr}^{-1}$  for the same band estimated from a previous inversion study using primarily the same observational data as input but for an earlier time period (Xiao, 2008). The inclusion of an ocean source does not affect our overarching conclusions on a shift in global  $\text{CH}_2\text{Cl}_2$  emissions, with an increasing contribution from Asia, and a declining contribution from Europe and North America since the mid-2000s. Additionally, comparisons of atmospheric measurements between this and the “no ocean” inversion lead to no evidence for (or against) an ocean  $\text{CH}_2\text{Cl}_2$  source.
- Unlike  $\text{CH}_2\text{Cl}_2$ , which has increased in the atmosphere since the early/mid-2000s,  $\text{C}_2\text{Cl}_4$  has been in long-term decline. Our results indicate a decrease in global emissions from  $141 \pm 14 \text{ Gg yr}^{-1}$  in 2007 to  $106 \pm 12 \text{ Gg yr}^{-1}$  in 2017. These values were obtained from an inversion setup in which the  $\text{C}_2\text{Cl}_4 + \text{Cl}$  sink reaction was not included and agree well with estimates produced for the recent WMO Ozone Assessment using a simplified model and similar input data (Engel et al., 2018). Inclusion of the  $\text{C}_2\text{Cl}_4 + \text{Cl}$  reaction, shown to be an important, albeit uncertain, sink of  $\text{C}_2\text{Cl}_4$  in recent modeling studies (Hossaini et al., 2019), increases the estimated global  $\text{C}_2\text{Cl}_4$  emissions to  $216 \pm 14 \text{ Gg yr}^{-1}$  in 2007 and  $162 \pm 12 \text{ Gg yr}^{-1}$  in 2017, that is, around 50% larger. Further work to constrain tropospheric Cl atom concentrations may help to constrain top-down emission estimates for  $\text{C}_2\text{Cl}_4$ . Inclusion of the Cl sink generally leads to slight improvements compared to the default inversion when comparing against atmospheric measurements.
- Using observational data not included in the inversion, the performance of the posterior  $\text{CH}_2\text{Cl}_2$  and  $\text{C}_2\text{Cl}_4$  emissions was evaluated. For both compounds, observed surface trends between the mid-2000s and 2017 are well reproduced:  $\sim 3 \text{ ppt CH}_2\text{Cl}_2 \text{ yr}^{-1}$  and  $\sim -0.1 \text{ ppt C}_2\text{Cl}_4 \text{ yr}^{-1}$  at NH sites. Independent measurements from the 2014 CAST/CONTRAST/ATTREX aircraft missions over the tropical West Pacific are also reproduced very well throughout the vertical extent of the troposphere for  $\text{CH}_2\text{Cl}_2$ , but

do not allow any constraint on the magnitude of an oceanic source. These independent data are reproduced relatively less successfully for  $C_2Cl_4$ . Similarly, the posterior emissions show improvement over the prior at numerous USA-based NOAA tall tower sites in 2015. Comparisons for other surface sites were performed, including Taunus Observatory (Germany) and Hateruma (Japan), and reveal generally good agreement. Our emissions are thus suitable for inclusion in global atmospheric modeling studies.

In conclusion, emissions of  $CH_2Cl_2$ —a substance known to cause ozone depletion—have increased significantly since the mid-2000s. Given that Asian emissions lead to a relatively large  $CH_2Cl_2$  ODP compared to emissions from other regions (Claxton et al., 2019), its regional and global abundance should continue to be monitored. As emissions from the Asian continent are by far the largest, a denser set of measurements, from the surface to the tropopause, would be beneficial to distinguish emissions from different subregions and ultimately constrain the troposphere-to-stratosphere input of chlorine from VSLS. A consideration of how emissions are likely to change in coming decades would also help constrain the influence of  $CH_2Cl_2$  on the timescale of stratospheric ozone recovery. Future work should also focus on elucidating the mechanism by which  $CH_2Cl_2$  is recycled through the ocean and quantifying the magnitude and distribution of biogenic sources.

## Acknowledgments

This work was supported by RH's NERC Independent Research Fellowship (NE/N014375/1), the NERC ISHOC project (NE/R004927/1), and the NERC SISLAC project (NE/R001782/1). MPC is supported by a Wolfson Merit Award. We thank Dr. Wuhu Feng for his work developing the TOMCAT/SLIMCAT model. We thank Dr. Prabir Patra and Dr. Bin Xiang for providing the HCFC-22 emission distribution. LJC thanks NERC for funding the CAST and ship observations (NE/J00619X/1, NE/K004980/1, and NE/I028769/1). Data from the NOAA monitoring network are available from <https://www.esrl.noaa.gov/gmd/>. We thank C. Siso for technical assistance in making the NOAA measurements and Dr B. Miller and Dr A. Andrews for assistance in providing results from NOAA's tall tower network. Gosan measurements were supported by Basic Science Research Program through the National Research Foundation of Korea (NRF) funded by the Ministry of Education (No. 2017R1D1A1B03034034). The operation of the AGAGE station at Ragged Point, Barbados, is supported by the National Aeronautics and Space Administration (NASA, USA; Grant NNX16AC98G to MIT and subaward 5710002970 to the University of Bristol) with additional support from the National Oceanic and Atmospheric Administration (NOAA, USA; contract RA133R15CN0008 to the University of Bristol). Observations at Hateruma were partly supported by the Ministry of the Environment of Japan. Data from the AGAGE monitoring network are available from <https://agage.mit.edu/data/> and are supported by the NASA Upper Atmospheric Research Program in the United States with grants NNX16AC96G and NNX16AC97G to SIO and grant NNX16AC98G to MIT. ATTREX and CONTRAST measurements were supported by National Science Foundation Grant AGS1261689 and NASA Grant NNX13AH20G. The model data can be accessed via the Lancaster University data repository at DOI: 10.17635/lancaster/researchdata/353. The Supporting Information to this article consists of seven figures and three tables.

## References

- Andrews, S. J., Carpenter, L., Apel, E., Atlas, E., Donets, V., Hopkins, J. R., et al. (2016). A comparison of very short lived halocarbon (VSLS) and DMS aircraft measurements in the tropical west Pacific from CAST, ATTREX and CONTRAST. *Atmospheric Measurement Techniques*, 9(10), 5213–5225. <https://doi.org/10.5194/amt-9-5213-2016>
- Andrews, S. J., Hackenberg, S. C., & Carpenter, L. J. (2015). Technical note: A fully automated purge and trap GC-MS system for quantification of volatile organic compound (VOC) fluxes between the ocean and atmosphere. *Ocean Science*, 11(2), 313–321. <https://doi.org/10.5194/os-11-313-2015>
- Ashfold, M. J., Pyle, J. A., Robinson, A. D., Meneguz, E., Nadzir, M. S. M., Phang, S. M., et al. (2015). Rapid transport of East Asian pollution to the deep tropics. *Atmospheric Chemistry and Physics*, 15(6), 3565–3573. <https://doi.org/10.5194/acp-15-3565-2015>
- Baker, D. F., Law, R. M., Gurney, K., Rayner, P., Peylin, P., Denning, A. S., et al. (2006). TransCom 3 inversion intercomparison: Impact of transport model errors on the interannual variability of regional  $CO_2$  fluxes, 1988–2003. *Global Biogeochemical Cycles*, 20(1), GB1002. <https://doi.org/10.1029/2004GB002439>
- Bousquet, P., Yver, C., Pison, I., Li, Y. S., Fortems, A., Hauglustaine, D., et al. (2011). A three-dimensional synthesis inversion of the molecular hydrogen cycle: Sources and sinks budget and implications for the soil uptake. *Journal of Geophysical Research*, 116(D1), D01302. <https://doi.org/10.1029/2010JD014599>
- Burkholder, J. B., Sander, S. P., Abbatt, J., Barker, J. R., Huie, R. E., Kolb, C. E., et al. (2015). *Chemical kinetics and photochemical data for use in atmospheric studies: Evaluation number 18*, JPL Publication 15-10, (). Pasadena: Jet Propulsion Laboratory. <http://jpldataeval.jpl.nasa.gov/>
- Carpenter, L. J., Reiman, S., Burkholder, J. B., Clerbaux, C., Hall, B. D., Hossaini, R., Laube, J. C., & Yvon-Lewis, S. A. (2014). Ozone-depleting substances (ODSs) and other gases of interest to the Montreal Protocol. In *Scientific assessment of ozone depletion: 2014, Global Ozone Research and Monitoring Project—Report No. 55, (Chap. 1)*. Geneva, Switzerland: World Meteorological Organization.
- Chen, Y.-H., & Prinn, R. G. (2006). Estimation of atmospheric methane emissions between 1996 and 2001 using a three-dimensional global chemical transport model. *Journal of Geophysical Research*, 111, D10307. <https://doi.org/10.1029/2005JD006058>
- Chipperfield, M. P. (2006). New version of the TOMCAT/SLIMCAT off-line chemical transport model: Intercomparison of stratospheric tracer experiments. *Quarterly Journal of the Royal Meteorological Society*, 132(617), 1179–1203. <http://doi.org/10.1256/qj.05.51>
- Claxton, T., Hossaini, R., Wild, O., Chipperfield, M. P., & Wilson, C. (2019). On the regional and seasonal ozone depletion potential of chlorinated, very short-lived substances. *Geophysical Research Letters*, 46(10), 5489–5498. <https://doi.org/10.1029/2018GL081455>
- Cunnold, D. M., Prinn, R. G., Rasmussen, R. A., Simmonds, P. G., Alyea, F. N., Cardelino, C. A., et al. (1983). The atmospheric lifetime experiment: 3. Lifetime methodology and application to three years of  $CFCl_3$  data. *Journal of Geophysical Research, Oceans*, 88(C13), 8379–8400. <https://doi.org/10.1029/JC088iC13p08379>
- Dee, D., Uppala, S., Simmons, A., Berrisford, P., Poli, P., Kobayashi, S., et al. (2011). The ERA-Interim reanalysis: Configuration and performance of the data assimilation system. *Quarterly Journal of the Royal Meteorological Society*, 137(656), 553–597. <https://doi.org/10.1002/qj.828>
- Engel, A., Rigby, M., Burkholder, J. B., Fernandez, R. P., Froidevaux, L., Hall, B. D., et al. (2018). Update on ozone-depleting substances (ODSs) and other gases of interest to the Montreal Protocol. In *Scientific assessment of ozone depletion: 2018, Global Ozone Research and Monitoring Project—Report No. 58, (Chap. 1)*. Geneva, Switzerland: World Meteorological Organization.
- Fang, X., Park, S., Saito, T., Tunnicliffe, R., Ganesan, A. L., Rigby, M., et al. (2019). Rapid increase in ozone-depleting chloroform emissions from China. *Nature Geoscience*, 12(2), 89–93. <https://doi.org/10.1038/s41561-018-0278-2>
- Feng, Y., Bie, P., Wang, Z., Wang, L., & Zhang, J. (2018). Bottom-up anthropogenic dichloromethane emission estimates from China for the period 2005–2016 and predictions of future emissions. *Atmospheric Environment*, 186, 241–247. <https://doi.org/10.1016/j.atmosenv.2018.05.039>
- Fernandez, R. P., Salawitch, R. J., Kinnison, D. E., Lamarque, J.-F., & Saiz-Lopez, A. (2014). Bromine partitioning in the tropical tropopause layer: Implications for stratospheric injection. *Atmospheric Chemistry and Physics*, 14(24), 13,391–13,410. <https://doi.org/10.5194/acp-14-13391-2014>
- Hackenberg, S. C., Andrews, S. J., Airs, R., Arnold, S. R., Bouman, H. A., Brewin, R. J. W., et al. (2017). Potential controls of isoprene in the surface ocean. *Global Biogeochemical Cycles*, 31(4), 644–662. <https://doi.org/10.1002/2016GB005531>
- Harris, N. R. P., Carpenter, L. J., Lee, J. D., Vaughan, G., Filus, M. T., Jones, R. L., et al. (2017). Co-ordinated Airborne Studies in the Tropics (CAST). *B. American Meteorological Society*, 98(1), 145–162. <https://doi.org/10.1175/BAMS-D-1400290.1>

- Hossaini, R., Atlas, E., Dhomse, S. S., Chipperfield, M. P., Bernath, P. F., Fernando, A. M., et al. (2019). Recent trends in stratospheric chlorine from very short-lived substances. *Journal of Geophysical Research: Atmospheres*, 124(4), 2318–2335. <https://doi.org/10.1029/2018JD029400>
- Hossaini, R., Chipperfield, M. P., Monge-Sanz, B. M., Richards, N. A. D., Atlas, E., & Blake, D. R. (2010). Bromoform and dibromomethane in the tropics: A 3-D model study of chemistry and transport. *Atmospheric Chemistry and Physics*, 10(2), 719–735. <https://doi.org/10.5194/acp-10-719-2010>
- Hossaini, R., Chipperfield, M. P., Montzka, S. A., Leeson, A. A., Dhomse, S. S., & Pyle, J. A. (2017). The increasing threat to stratospheric ozone from dichloromethane. *Nature Communications*, 8(1), 1–9. <https://doi.org/10.1038/ncomms15962>
- Hossaini, R., Chipperfield, M. P., Saiz-Lopez, A., Fernandez, R., Monks, S., Feng, W., et al. (2016). A global model of tropospheric chlorine chemistry: Organic versus inorganic sources and impact on methane oxidation. *Journal of Geophysical Research*, 121, 14,271–14,297. <https://doi.org/10.1002/2016JD025756>
- Huijnen, V., William, J., van Weele, M., van Noije, T., Krol, M., Dentener, F., et al. (2010). The global chemistry transport model TM5: Description and evaluation of the tropospheric chemistry version 3.0. *Geoscientific Model Development*, 3(2), 445–473. <https://doi.org/10.5194/gmd-3-445-2010>
- Johnson, M. T. (2010). A numerical scheme to calculate temperature and salinity dependent air-water transfer velocities for any gas. *Ocean Science*, 6(4), 913–932. <https://doi.org/10.5194/os-6-913-2010>
- Jones, C. E., & Carpenter, L. J. (2005). Solar photolysis of CH<sub>2</sub>I<sub>2</sub>, CH<sub>2</sub>ICl, and CH<sub>2</sub>IBr in water, saltwater, and seawater. *Environmental Science and Technology*, 39(16), 6130–6137. <https://doi.org/10.1021/es050563g>
- Keene, W. C., Khalil, M. A. K., Erickson, D. J. III, McCulloch, A., Graedel, T. E., Lobert, J. M., et al. (1999). Composite global emissions of reactive chlorine from anthropogenic and natural sources: Reactive Chlorine Emissions Inventory. *Journal of Geophysical Research*, 104(D7), 8429–8440. <https://doi.org/10.1029/1998JD100084>
- Khalil, M. A. K., Moore, R. M., Harper, D. B., Lobert, J. M., Erickson, D. J., Koropalov, V., et al. (1999). Natural emissions of chlorine-containing gases: Reactive Chlorine Emissions Inventory. *Journal of Geophysical Research*, 104(D7), 8333–8346. <https://doi.org/10.1029/1998JD100079>
- Khalil, M. A. K., & Rasmussen, R. A. (1999). Atmospheric chloroform. *Atmospheric Environment*, 33(7), 1151–1158. [https://doi.org/10.1016/S1352-2310\(98\)00233-7](https://doi.org/10.1016/S1352-2310(98)00233-7)
- Kolusu, S., Schlünzen, H., Grawe, D., & Seifert, R. (2016). Chloromethane and dichloromethane in the tropical Atlantic Ocean. *Atmospheric Environment*, 150, 417–424. <https://doi.org/10.1016/j.atmosenv.2016.11.037>
- Kolusu, S. R., Schlünzen, K. H., Grawe, D., & Seifert, R. (2018). Determination of chloromethane and dichloromethane in a tropical terrestrial mangrove forest in Brazil by measurements and modelling. *Atmospheric Environment*, 173, 185–197. <https://doi.org/10.1016/j.atmosenv.2017.10.057>
- Laube, J. C., Engel, A., Bönisch, H., Möbius, T., Worton, D. R., Sturges, W. T., et al. (2008). Contribution of very short-lived organic substances to stratospheric chlorine and bromine in the tropics—A case study. *Atmospheric Chemistry and Physics*, 8(23), 7325–7334. <https://doi.org/10.5194/acp-8-7325-2008>
- Law, R. M., Peters, W., Rodenbeck, C., Aulagnier, C., Baker, I., Bergmann, D. J., et al. (2008). TransCom model simulations of hourly atmospheric CO<sub>2</sub>: Experimental overview and diurnal cycle results for 2002. *Global Biogeochemical Cycles*, 22(3), GB3009. <https://doi.org/10.1029/2007gb003050>
- Lawson, S. J., Keywood, M. D., Galbally, I. E., Gras, J. L., Caine, J. M., Cope, M. E., et al. (2015). Biomass burning emissions of trace gases and particles in marine air at Cape Grim. *Tasmania, Atmospheric Chemistry and Physics*, 15(23), 13,393–13,411. <https://doi.org/10.5194/acp-15-13393-2015>
- Leedham Elvidge, E. C., Oram, D. E., Laube, J. C., Baker, A. K., Montzka, S. A., Humphrey, S., et al. (2015). Increasing concentrations of dichloromethane, CH<sub>2</sub>Cl<sub>2</sub>, inferred from CARIBIC air samples collected 1998–2012. *Atmospheric Chemistry and Physics*, 15, 1939–1958. <https://doi.org/10.5194/acp-15-1939-2015>
- Lobert, J., Keene, W. C., Logan, J., & Yevich, R. (1999). Global chlorine emissions from biomass burning—Reactive Chlorine Emissions Inventory. *Journal of Geophysical Research*, 104, 8373–8390. <https://doi.org/10.1029/1998JD100077>
- McCulloch, A., Aucott, M. L., Graedel, T. E., Kleiman, G., Midgley, M., & Li, Y.-F. (1999). Industrial emissions of trichloroethene, tetrachloroethene, and dichloromethane: Reactive Chlorine Emissions Inventory. *Journal of Geophysical Research*, 104, 8417–8427. <https://doi.org/10.1029/1999JD900011>
- McCulloch, A., & Midgley, P. M. (1996). The production and global distribution of emissions of trichloroethene, tetrachloroethene and dichloromethane over the period 1988–1992. *Atmospheric Environment*, 30, 601–608. [https://doi.org/10.1016/1352-2310\(96\)00032-5](https://doi.org/10.1016/1352-2310(96)00032-5)
- McNorton, J., Wilson, C., Gloor, M., Parker, R. J., Boesch, H., Feng, W., et al. (2018). Attribution of recent increases in atmospheric methane through 3-D inverse modelling. *Atmospheric Chemistry and Physics*, 18, 18,149–18,168. <https://doi.org/10.5194/acp-18-18149-2018>
- Monks, S., Arnold, S. R., Hollaway, M., Pope, R., Wilson, C., Feng, W., et al. (2016). The TOMCAT global chemistry transport model: Description of chemical mechanism and model evaluation. *Geoscientific Model Development Discussion*, 1–51. <https://doi.org/10.5194/gmd-2016-212>
- Montzka, S. A., Dutton, G. S., Yu, P., Ray, E., Portmann, R. W., Daniel, J. S., et al. (2018). An unexpected and persistent increase in global emissions of ozone-depleting CFC-11. *Nature*, 557(7705), 413–417. <https://doi.org/10.1038/s41586-018-0106-2>
- Montzka, S. A., Krol, M., Dlugokencky, E., Hall, B., Jöckel, P., & Lelieveld, J. (2011). Small inter annual variability of global atmospheric hydroxyl. *Science*, 331(67-69). <https://doi.org/10.1126/science.1197640>
- Montzka, S. A., Reimann, S., Engel, A., Kruger, K., O'Doherty, S., Sturges, W. T., et al. (2011). Ozone-depleting substances (ODSs) and related chemicals. In *Scientific assessment of ozone depletion: 2010, Global Ozone Research and Monitoring Project—Report No. 52, (Chap. 1)*. Geneva, Switzerland: World Meteorological Organization.
- Moore, R. M. (2004). Dichloromethane in North Atlantic waters. *Journal of Geophysical Research*, 109, C09004. <https://doi.org/10.1029/2004JC002397>
- Mühle, J., Lueker, T. J., Su, Y., Miller, B. R., Prather, K. A., & Weiss, R. F. (2007). Trace gas and particulate emissions from the 2003 southern California wildfires. *Journal of Geophysical Research*, 112, D03307. <https://doi.org/10.1029/2006JD007350>
- Mühle, J., Trudinger, C. M., Western, L. M., Rigby, M., Vollmer, M. K., Park, S., Manning, A. J., et al. (2019). Perfluorocyclobutane (PFC-318, c-C<sub>4</sub>F<sub>8</sub>) in the global atmosphere. *Atmospheric Chemistry and Physics*, 19(15), 10,335–10,359. <https://doi.org/10.5194/acp-19-10335-2019>



- Navarro, M. A., Atlas, E. L., Saiz-Lopez, A., Rodriguez-Lloveras, X., Kinnison, D. E., Lamarque, J. F., et al. (2015). Airborne measurements of organic bromine compounds in the Pacific tropical tropopause layer. *Proceedings of the National Academy of Sciences of the United States of America*, 112(45), 13,789–13,793. <https://doi.org/10.1073/pnas.1511463112>
- Ooki, A., & Yokouchi, Y. (2011). Dichloromethane in the Indian Ocean: Evidence for in-situ production in seawater. *Marine Chemistry*, 124(1–4), 119–124. <https://doi.org/10.1016/j.marchem.2011.01.001>
- Oram, D. E., Ashfold, M. J., Laube, J. C., Gooch, L. J., Humphrey, S., Sturges, W. T., et al. (2017). A growing threat to the ozone layer from short-lived anthropogenic chlorocarbons. *Atmospheric Chemistry and Physics*, 17(19), 11,929–11,941. <https://doi.org/10.5194/acp-17-11929-2017>
- Pan, L. L., Altas, E. L., Salawitch, R. J., Honomichl, S. B., Bresch, J. F., Randel, W. J., et al. (2017). The Convective Transport of Active Species in the Tropics (CONTRAST) experiment. *Bulletin of the American Meteorological Society*, 98(1), 106–128. <https://doi.org/10.1175/BAMS-D-14-00272.1>
- Patra, P. K., Houweling, S., Krol, M., Bousquet, P., Belikov, D., Bergmann, D., et al. (2011). TransCom model simulations of CH<sub>4</sub> and related species: Linking transport, surface flux and chemical loss with CH<sub>4</sub> variability in the troposphere and lower stratosphere. *Atmospheric Chemistry and Physics*, 11(24), 12,813–12,837. <https://doi.org/10.5194/acp-11-12813-2011>
- Pétron, G., Granier, C., Khattatov, B., Lamarque, J. F., Yudin, V., Müller, J. F., & Gille, J. (2002). Inverse modelling of carbon monoxide surface emissions using Climate Monitoring and Diagnostics Laboratory network observations. *Journal of Geophysical Research*, 107(D24), 4761. <https://doi.org/10.1029/2001JD001305>
- Prinn, R. G., Huang, J., Weiss, R. F., Cunnold, D. M., Fraser, P. J., Simmonds, P. G., et al. (2005). Evidence for variability of atmospheric hydroxyl radicals over the past quarter century. *Geophysical Research Letters*, 32, L07809. <https://doi.org/10.1029/2004GL022228>
- Prinn, R. G., Weiss, R. F., Arduini, J., Arnold, T., DeWitt, H. L., Fraser, P. J., et al. (2018). History of chemically and radiatively important atmospheric gases from the Advanced Global Atmospheric Gases Experiment (AGAGE). *Earth System Science Data*, 10(985–1018), 2018. <https://doi.org/10.5194/essd-10-985-2018>
- Quack, B., & Wallace, D. W. R. (2003). Air-sea flux of bromoform: Controls, rates, and implications. *Global Biogeochemical Cycles*, 17(1), 1023. <https://doi.org/10.1029/2002GB001890>
- Rigby, M., Prinn, R. G., O'Doherty, S., Montzka, S. A., McCulloch, A., Harth, C. M., et al. (2013). Re-evaluation of the lifetimes of the major CFCs and CH<sub>3</sub>CCl<sub>3</sub> using atmospheric trends. *Atmospheric Chemistry and Physics*, 13, 2691–2702. <https://doi.org/10.5194/acp-13-2691-2013>
- Rudolph, J., Koppmann, R., & Plass-Dülmer, C. (1996). The budgets of ethane and tetrachloroethene: Is there evidence for an impact of reactions with chlorine atoms in the troposphere? *Atmospheric Environment*, 30(11), 1887–1894. [https://doi.org/10.1016/1352-2310\(95\)00385-1](https://doi.org/10.1016/1352-2310(95)00385-1)
- Schuck, T. J., Lefrançois, F., Gallmann, F., Wang, D., Jesswein, M., Hoker, J., et al. (2018). Halocarbons at Taunus Observatory. *Atmospheric Chemistry and Physics*, 18, 16,553–16,569. <https://doi.org/10.5194/acp-18-16553-2018>
- Simmonds, P. G., Manning, A. J., Cunnold, D. M., McCulloch, A., O'Doherty, S., Derwent, R. G., et al. (2006). Global trends, seasonal cycles, and European emissions of dichloromethane, trichloroethene, and tetrachloroethene from the AGAGE observations at Mace Head, Ireland, and Cape Grim, Tasmania. *Journal of Geophysical Research*, 111, D18304. <https://doi.org/10.1029/2006JD007082>
- Simpson, I. J., Akagi, S. K., Barletta, B., Blake, N. J., Choi, Y., Diskin, G. S., et al. (2011). Boreal forest fire emissions in fresh Canadian smoke plumes: C<sub>1</sub>–C<sub>10</sub> volatile organic compounds (VOCs), CO<sub>2</sub>, CO, NO<sub>2</sub>, NO, HCN and CH<sub>3</sub>CN. *Atmospheric Chemistry and Physics*, 11, 6445–6463. <https://doi.org/10.5194/acp-11-6445-2011>
- Simpson, I. J., Meinardi, S., Blake, N. J., Rowland, F. S., & Blake, D. R. (2004). Long-term decrease in the global atmospheric burden of tetrachloroethene (C<sub>2</sub>Cl<sub>4</sub>). *Geophysical Research Letters*, 31(8), L08108. <https://doi.org/10.1029/2003GL019351>
- Spivakovsky, C. M., Logan, J. A., Montzka, S. A., Balkanski, Y. J., Foreman-Fowler, M., Jones, D. B. A., et al. (2000). Three-dimensional climatological distribution of tropospheric OH: Update and evaluation. *Journal of Geophysical Research*, 105(D7), 8931–8980. <https://doi.org/10.1029/1999JD901006>
- Sturges, W., Oram, D., Carpenter, L., Penkett, S., & Engel, A. (2000). Bromoform as a source of stratospheric bromine. *Geophysical Research Letters*, 27(14), 2081–2084. <https://doi.org/10.1029/2000GL011444>
- Tarantola, A., & Valette, B. (1982). Generalized nonlinear inverse problems solved using the least squares criterion. *Reviews of Geophysics*, 20(2), 219–232. <https://doi.org/10.1029/RG020i002p00219>
- Trudinger, C. M., Etheridge, D. M., Sturrock, G. A., Fraser, P. J., Krummel, P. B., & McCulloch, A. (2004). Atmospheric histories of halocarbons from analysis of Antarctic firn air: Methyl bromide, methyl chloride, chloroform and dichloromethane. *Journal of Geophysical Research*, 109, D22310. <https://doi.org/10.1029/2004JD004932>
- Tsai, W. T. (2017). Fate of chloromethanes in the atmospheric environment: Implications for human health, ozone formation and depletion, and global warming impacts. *Toxics*, 5(4), 23. <https://doi.org/10.3390/toxics5040023>
- Turner, A. J., Fung, I., Naik, V., Horowitz, L. W., & Cohen, R. C. (2018). Modulation of hydroxyl variability by ENSO in the absence of external forcing. *Proceedings of the National Academy of Sciences*, 115(36), 8931–8936. <https://doi.org/10.1073/pnas.1807532115>
- Wales, P. A., Salawitch, R. J., Nicely, J. M., Anderson, D. C., Canty, T. P., Baidar, S., et al. (2018). Stratospheric injection of brominated very short-lived substances: Aircraft observations in the Western Pacific and representation in global models. *Journal of Geophysical Research: Atmospheres*, 123, 5690–5719. <https://doi.org/10.1029/2017JD027978>
- Wang, J. S., Kawa, S. R., Collatz, G. J., Sasakawa, M., Gatti, L. V., Machida, T., et al. (2018). A global synthesis inversion analysis of recent variable in CO<sub>2</sub> fluxes using GOSAT and in situ observations. *Atmospheric Chemistry and Physics*, 18(15), 11,097–11,124. <https://doi.org/10.5194/acp-18-11097-2018>
- Xiang, B., Patra, P. K., Montzka, S. A., Miller, S. M., Elkins, J. W., Moore, F. L., et al. (2014). Global emissions of refrigerants HCFC-22 and HFC-134a: Unforeseen seasonal contributions. *Proceedings of the National Academy of Sciences of the United States of America*, 111(49), 17,379–17,384. <https://doi.org/10.1073/pnas.1417372111>
- Xiao, X. (2008). Optimal estimation of the surface fluxes of chloromethanes using a 3-D global atmospheric chemical transport model (Doctoral dissertation). ([https://globalchange.mit.edu/sites/default/files/Xiao\\_PhD\\_08.pdf](https://globalchange.mit.edu/sites/default/files/Xiao_PhD_08.pdf)). Cambridge, Massachusetts, USA: Massachusetts Institute of Technology.
- Ziska, F., Quack, B., Abrahamsson, K., Archer, S. D., Atlas, E., Bell, T., et al. (2013). Global sea-to-air flux climatology for bromoform, dibromomethane and methyl iodide. *Atmospheric Chemistry and Physics*, 13(17), 8915–8934. <https://doi.org/10.5194/acp-13-8915-2013>