https://doi.org/10.1016/j.chemosphere.2019.124580

Alwyn Fernandes, Jerzy Falandysz, Jesus Olivero-Verbel

PII: S0045-6535(19)31804-1

DOI: https://doi.org/10.1016/j.chemosphere.2019.124580

Reference: CHEM 124580

To appear in: ECSN



Please cite this article as: Fernandes, A., Falandysz, J., Olivero-Verbel, J., *Chemosphere*, https://doi.org/10.1016/j.chemosphere.2019.124580.

This is a PDF file of an article that has undergone enhancements after acceptance, such as the addition of a cover page and metadata, and formatting for readability, but it is not yet the definitive version of record. This version will undergo additional copyediting, typesetting and review before it is published in its final form, but we are providing this version to give early visibility of the article. Please note that, during the production process, errors may be discovered which could affect the content, and all legal disclaimers that apply to the journal pertain.

© 2019 Published by Elsevier Ltd.

1 Editorial

2

3

4

A new focus on legacy pollutants: Chlorinated Paraffins (CPs) and

Polychlorinated Naphthalenes (PCNs)

5

6

7

8

9

10

11

12

13

14

15

16

17

18

19

20

21

22

23

24

25

With volume production spanning almost a century, chlorinated paraffins (CPs) are industrial chemicals that are of increasing concern to the environment and to human health. Polychlorinated naphthalenes (PCNs) are legacy contaminants that occur in the environment, food chains and human tissues and contribute to the burden of dioxin-like effects such as carcinogenicity, hepatotoxicity, teratogenicity, embryotoxicity, etc. In recognition of these concerns, and following structured risk assessments, short-chain CPs (SCCPs, C10-C13) and PCNs have recently been listed for elimination, in Annex A of the Stockholm Convention. This followed an earlier listing for SCCPs as a priority hazardous substance for control under the European Union (EU) Water Framework Directive. The importance of these developments has not been lost on scientists who have researched an increased volume of new information in these fields. The Dioxin 2018 symposium in Krakow devoted a full day to the dissemination of the latest findings, with 18 oral presentations and 8 posters covering analytical aspects, occurrence in the environment, materials and food, and the toxicological effects of these pollutants. Chemically, CP products are complex isomeric mixtures of several thousands of individual compounds having carbon chain lengths ranging from C_{10} to C_{28} . Reliable analytical determination is one of the most intractable barriers to the accurate measurement of CP occurrence, a recognised drawback that was illustrated in the inter-laboratory studies by Krätschmer and Schächtele (Chemosphere 234, 252-259). The issue is compounded by the ambiguity in defining CP analytes – a necessary first step in reliable analytical determination.

The Stockholm convention lists SCCPs in its annex, but this relates to commercial products. 26 An innumerable selection of different "SCCP" products of varying compositions were 27 manufactured globally, and the environmental legacy of widespread historical usage that is 28 observed in environmental media is a complex integral of these mixtures. This integral is 29 further modified by transformative processes such as modification during usage, selective 30 rates of evaporation, photochemical and microbial degradation (Heeb et al. Chemosphere 226, 31 744-754), etc. Reported occurrences in food of animal origin (Krätschmer et al., Chemosphere 32 227, 630-637; Labadie et al., Chemosphere 223, 232-239; Jiang et al., Chemosphere 229, 358-33 365) imply additional transformation through metabolic processes which further enhance the 34 complexity of the observed profiles. It is therefore unsurprising that the profiles for SCCPs 35 (and other CPs) observed during analysis do not correspond to individual commercial 36 products, and perhaps more relevantly, to analytical standards. 37 38 In general, early insights into the toxicological effects of CPs were based on the use of standard mixtures that reflected the commercial products. As CP residues in real foods and 39 40 animal tissues have never been completely characterised (but are clearly modified integrals of different mixtures), it would prove difficult for a human exposure based risk assessment to 41 correlate occurrences to the reported effects. Another pressing issue is that of current CP 42 manufacture and use. The most recent literature suggests that shining a regulatory spotlight on 43 SCCPs, has resulted in a shift to the use of medium- (MCCPs, C₁₄-C₁₇) and long-chain CP 44 (LCCPs, > C₁₈) mixtures. Despite some investigations on MCCPs within the EU and North 45 America there is still a lower level of knowledge on the toxicity of these products, particularly 46 the LCCPs. In the long term, regulation of SCCPs on their own is unlikely to address potential 47 risks arising from these other mixtures and their breakdown products. Further, CP mixtures 48 are known to contain other chlorinated contaminants as by-products, such as polychlorinated 49 biphenyls (PCBs), PCNs and chlorinated dioxins and furans (PCDD/Fs), and may also give 50

51	rise to these as by-products during combustion (Matsukami and Kajiwara Chemosphere 230,
52	164-172). Clarification is required as to whether the toxicological effects that have been
53	reported were directly attributable to CPs, and were not influenced, at least in part, by the
54	presence in the test mixtures of such by-products which show more sensitive toxicological
55	endpoints than CPs.
56	As the assessment of exposure is an integral component of human health risk assessment, a
57	useful first step would be to characterise the CP profiles observed in foods as these are
58	expected to constitute an important exposure pathway (as observed with other similar
59	halogenated contaminants). This should include food packaging materials which have also
60	been shown to contain CPs (Wang et al., Chemosphere 225, 557-564). Recent advances in
61	instrumentation that allow qualitative homologue group characterisation of CP occurrence
62	may prove a useful tool in the identification and mapping of groups that predominate in
63	"typical" profiles for different food types. As many household and workplace materials are
64	known to contain CPs, a similar approach to characterising these occurrences would yield
65	information on other possible exposure pathways.
66	This characterisation would have two immediate advantages - it would provide a more
67	focussed approach to toxicological studies by allowing the targeting of relevant (occurring)
68	homologue groups and help identify groups that elicited more potent responses, and also
69	provide direction to the analytical effort by indicating a qualitative definition of the analytes.
70	The characterisation would also allow the formulation of more relevant standard CP mixtures
71	that correlate to a greater extent with observed profiles, thus aiding quantitative
72	determination. In this context specific single chain length mixtures are currently being
73	synthesised and characterised (Sprengel et al., Chemosphere 228, 762-768). The discrepancies
74	observed in the most recent inter-laboratory comparisons (Krätschmer and Schächtele,
75	Chemosphere 234, 252-259) underline the requirement for representative standards, but also

highlight the need for a robust and harmonised approach to the quantitation procedures 76 applied to the identified CP homologue groups. Until there is progress on these issues, the 77 expression of CP concentrations as total CP (either combined, or if analytical advances allow, 78 speciated into short, medium and long chain) would be a sensible interim measure, allowing 79 the generation of much needed occurrence data and laying the groundwork for future control 80 and regulation efforts. 81 For PCNs, information from emerging research continues to define the issues surrounding 82 these contaminants. An increasing amount of recent literature that speciates PCN occurrences 83 in environmental media and foods, by individual congeners, provides further information on 84 the persistence and fate of these chemicals, decades after production ceased. The historical 85 and continuing human exposure arising from PCN occurrence in foods and dietary 86 supplements (Falandysz et al., Chemosphere 231, 240-248; Zhihua et al., Chemosphere 230, 87 88 559-566) underlines the persistence of PCN congeners and the enduring legacy of this contamination in marine regions from where current fish supplies continue to be sourced. 89 90 New insights into the environmental behaviour and chemistry of individual congeners help to 91 explain observed patterns in environmental media and the resulting occurrence, particularly in marine products. 92 Relative to CPs, the analytical determination of PCNs is at an advanced level with reliable 93 measurement of individual congeners allowing behavioural studies of selected compounds. It 94 is particularly encouraging to see new work that adds to the body of toxicological insights into 95 PCN disposition in animal tissues and the effects on reproductive processes (Kilanowicz et 96 al., Chemosphere 226, 75-84; Kilanowicz et al., Chemosphere 228, 577-585). The dioxin-like 97 behaviour of some PCN congeners has been recognised for several years, but the 98 identification of other toxicological effects such as disruption to haemostasis parameters such 99

100	as clot formation and fibrinolysis, adds to the growing evidence of requirement for future
101	regulatory action.
102	Although the majority of the work presented at Dioxin 2018 included targeted studies with
103	specific outcomes directed to ultimately investigating the environmental, human exposure and
104	health effects of PCNs and CPs, it is important that this collated dissemination is viewed
105	within a wider context. Both of these contaminant classes are mass produced anthropogenic
106	products that have seen, often unrestricted, usage for the best part of a century. However, the
107	volume of pertinent literature is relatively small in comparison to other similar contaminants
108	such as PCBs, PCDDs/Fs and flame retardants. In the case of CPs, the combination of a lack
109	of widespread recognition combined with the real difficulty with analytical access is a clear
110	factor. For PCNs, the similarity of chemical behavior and effects to the more widely
111	produced PCBs has overshadowed the potent toxicological response of these chemicals.
112	However, in the light of the current re-evaluation of PCB toxicity, particularly PCB 126, the
113	contribution of PCNs to the cumulative dioxin-like toxicity could potentially become more
114	significant, engendering more interest in the regulation of these contaminants as well.
115	The inclusion of both these classes of contaminants within the Stockholm convention listing
116	has been followed by regional interest, e.g. within the EU, which has set up specific working
117	groups to address human exposure through the occurrence of these chemicals in food. Both of
118	these measures, provide direction to the task for scientists to facilitate and generate
119	information that will ensure that the remaining challenges and risks to human health are
120	characterized and are available for policy making.
121	The editors express their gratitude to all contributors to this special issue, in particular to the
122	contributing authors, and especially to the reviewers who collectively ensured a high standard
123	of scientific dissemination.

125	
126	¹ Alwyn Fernandes*, ^{2,3} Jerzy Falandysz, ³ Jesus Olivero-Verbel
127	
128	
129	¹ School of Environmental Sciences, University of East Anglia, Norwich NR4 7TJ, UK
130	² University of Gdańsk, Environmental Chemistry and Ecotoxicology, 63 Wita Stwosza Str.,
131	80-308 Gdańsk, Poland
132	³ Environmental and Computational Chemistry Group, School of Pharmaceutical Sciences,
133	Zaragocilla Campus, University of Cartagena, 130015 Cartagena, Colombia
134	
135	E-mail addresses:
136	alwyn.fernandes@uea.ac.uk (A. Fernandes),
137	jerzy.falandysz@gmail.com (J. Falandysz),
138	joliverov@unicartagena.edu.co (J. Olivero-Verbel).
139	
140	*Corresponding author: alwyn.fernandes@uea.ac.uk ; arfernand@hotmail.com;
141	
142	
143	
144	
145	
146	
147	
148	
149	

