Title: Contribution of spatially explicit models to climate change adaptation and
 mitigation plans for a priority forest habitat

3

4 Abstract

Climate change will impact forest ecosystems, their biodiversity and the livelihoods they 5 sustain. Several adaptation and mitigation strategies to counteract climate change impacts 6 7 have been proposed for these ecosystems. However, effective implementation of such strategies requires a clear understanding of how climate change will influence the future 8 9 distribution of forest ecosystems. This study uses maximum entropy modelling (MaxEnt) to predict environmentally suitable areas for cork oak (Quercus suber) woodlands, a 10 11 socio-economically important forest ecosystem protected by the European Union Habitats Directive. Specifically, we use two climate change scenarios to predict changes 12 13 in environmental suitability across the entire geographical range of the cork oak and in areas where stands were recently established. Up to 40% of current environmentally 14 15 suitable areas for cork oak may be lost by 2070, mainly in northern Africa and southern Iberian Peninsula. Almost 90% of new cork oak stands are predicted to lose suitability by 16 17 the end of the century, but future plantations can take advantage of increasing suitability in northern Iberian Peninsula and France. The predicted impacts cross country borders, 18 showing that a multinational strategy will be required for cork oak woodland adaptation 19 to climate change. Such a strategy must be regionally adjusted, featuring the protection 20 of refugia sites in southern areas and stimulating sustainable forest management in areas 21 22 that will keep long-term suitability. Afforestation efforts should also be promoted but must consider environmental suitability and land competition issues. 23

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25 Keywords:

Afforestation, climate change impacts, conservation planning, dehesa, environmentalniche modelling, montado

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35 **1 Introduction**

Global climate change is affecting ecosystems worldwide (Parmesan, 2006; 36 Parmesan and Yohe, 2003). Forest ecosystems are particularly susceptible to shifts in 37 natural disturbance regimes induced by climate change (Dale et al., 2001; Trumbore et 38 al., 2015). For example, changes in the intensity, frequency and duration of wildfires or 39 droughts can negatively affect tree growth and recruitment, increase tree defoliation and, 40 ultimately, induce tree mortality (Allen et al., 2010; Caldeira et al., 2015; Lindner et al., 41 2010; Walck et al., 2011). Adaptation and mitigation strategies against climate change 42 43 are crucial to counteract such effects. However, effective implementation of such strategies requires accurate identification of areas that may remain environmentally 44 45 suitable for forest species and areas with potential for afforestation (Aitken et al., 2008; Millar et al., 2007; Settele et al., 2014). This is particularly important in climate change 46 47 hotspots, such as the Mediterranean Basin (Giorgi, 2006), where the development of forest adaptation and mitigation strategies is a pressing challenge (Doblas-Miranda et al., 48 49 2015; Scarascia-Mugnozza et al., 2000).

Cork oak (Quercus suber) woodlands are agro-silvo-pastoral systems of high 50 51 socio-economic and conservation value, typical of the Western Mediterranean Basin. They cover approximately 1.5 million ha across Portugal, Spain, Italy and France, and 1 52 million ha in North Africa between Morocco, Algeria and Tunisia (Bugalho et al., 2011; 53 Diáz et al., 1997). Cork oak woodlands have a relatively open, savannah-like tree 54 structure (about 30 to 60 trees per ha) and a heterogeneous shrub-grassland matrix 55 understory (Bugalho et al., 2009). These woodlands host plant and animal species of high 56 conservation value (Correia et al., 2015a; Diáz et al., 1997), including endemic or 57 threatened species such as Iberian imperial eagle Aquila adalberti, Black stork Ciconia 58 nigra or Iberian lynx Lynx pardinus. Cork oak woodlands are classified under the 59 60 European Union Habitats Directive (92/43/CEE) and are included in the Natura 2000 network of protected areas (Berrahmouni et al., 2009). Cork oak woodlands also have a 61 62 very high socio-economic value, mostly derived from cork and livestock production (Bugalho et al., 2009; Pereira and Tomé, 2004). Cork can be harvested every 9 to 12 years 63 64 without significant damage to the tree or affecting the biodiversity of these woodlands (Leal et al., 2011). Cork is mainly used for wine bottle stoppers (over 70% of production), 65 66 although there has been a recent increase in other applications such as insulation materials and pavements (Bugalho et al., 2009; Bugalho et al., 2011). Approximately 300 thousand 67 68 tons of cork are harvested across the western Mediterranean Basin annually (Berrahmouni

et al., 2007). Cork is the sixth most important non-timber forest product worldwide, with
an estimated export value of US\$329 million, and processed cork products generate an
annual revenue of US\$ 2 billion (Berrahmouni et al., 2007).

72 Current threats to cork oak woodlands include a lack of natural oak regeneration 73 and high adult oak mortality, eventually leading to declines in tree density and area loss (Plieninger et al., 2010; Santos and Thorne, 2010). Climate change will exacerbate these 74 75 threats through an increase in temperatures and frequency of droughts (Acácio et al., 2016; Besson et al., 2014; Caldeira et al., 2015), both of which will contribute to an 76 77 increase in the frequency and severity of wildfires (Acácio et al., 2007; Godinho et al., 78 2016), especially in areas with inadequate management (Bugalho et al., 2011; Godinho 79 et al., 2016).

Here, we use environmental niche models (ENMs) to predict changes in 80 81 environmental suitability across the geographic range of cork oak woodlands in response to climate change. Our objectives were to (i) quantify changes in environmental suitability 82 83 across the cork oak geographic range using two climate change scenarios; (ii) assess whether ongoing afforestation efforts have taken place in those areas most likely to 84 85 remain environmentally suitable for the species, using Portugal as a case study, and; (iii) discuss potential climate change adaptation and mitigation strategies for cork oak 86 woodlands at regional and global scales. 87

88

89 **2 Methods**

90 2.1 Cork oak distribution data

Cork oak distribution was obtained by geo-referencing or collecting geo-91 referenced data in national forestry and biodiversity inventories. Data were collected for 92 all the countries where the species occurs naturally (Fig. 1): Portugal (Autoridade 93 94 Florestal Nacional, 2009), Spain (Dirección General de Medio Natural y Política Forestal, 2009), France (Institut National de l'Information Géographique et Forestiére, 2010), Italy 95 96 (Vessella and Schirone, 2013), Morocco (Haut Commissariat aux Eaux et Forêts et à la Lutte Contre la Désertification, 2005), Algeria (Barry et al., 1974), and Tunisia (Khaldi, 97 98 2004). The spatial resolution of the data differed among countries, so the complete distribution dataset was up-scaled to a 5 arc-minute resolution grid. This process served 99 to homogenise the spatial resolution of the distribution data and match it to the resolution 100 of the climate data used (see below). 101

We also obtained information on the location of cork oak stands recently planted in Portugal (Autoridade Florestal Nacional, 2009), the country with the largest area of cork oak woodland, in order to evaluate the adequacy of the geographic location of new afforestation considering climate change scenarios.

- 106
- 107 2.2 Environmental data

Climate data representing current (1950-2000) and future (2061-2080) conditions 108 were downloaded from the WorldClim database (http://www.worldclim.org/) at 5 arc-109 110 minute resolution. Future climatic conditions were derived from four Global Circulation Models (GCMs - ACCESS1-0, CCSM4, HadGEM2-ES and MPI-ESM-LR) and two 111 112 Representative Concentration Pathways scenarios (RCP 4.5 and 8.5). These scenarios 113 were chosen to represent moderate (RCP 4.5, average increase of 1.8°C by 2100) and 114 extreme (RCP 8.5, average increase of 3.7°C by 2100) warming trends (Stocker et al., 115 2013). We collected data from the standard set of 19 bioclimatic variables (Hutchinson et 116 al., 2009) available in the WorldClim database. Furthermore, we calculated additional variables potentially relevant for cork oak: number of frost days (New et al., 2000) and 117 118 indices of annual and seasonal aridity (Zomer et al., 2008). Two soil related variables, 119 Soil type and Soil pH, were collected from the Harmonized World Soil Database, also at 120 5 arc-minute resolution (FAO/IIASA/ISRIC/ISSCAS/JRC, 2012).

Twelve environmental variables (Table 1) were then selected for model 121 calibration purposes based on the biological knowledge of the species' requirements 122 (Pausas et al., 2009). The choice of proximal variables (i.e. variables that closely relate to 123 the physiological limits of the species) is often recommended for modelling species 124 distributions as it allows for more robust predictions and facilitates model interpretability 125 (Buckley et al., 2010; Kearney and Porter, 2009; Synes and Osborne, 2011). Several of 126 127 the potentially relevant environmental variables were highly correlated (Table S1 in Online Resource 1) which could affect model outcomes (Dormann et al., 2013; Merow et 128 129 al., 2013). To minimize this, we generated environmental variable subsets which only 130 contained variables with a correlation coefficient < |0.7| using the ENMeval library 131 (Muscarella et al., 2014) for R software package (R Core Team, 2016). Model AUC and AIC scores (Table S2 in Online Resource 1) were then used to select the best subset of 132 133 environmental variables for modelling purposes.

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135 **2.3 Modelling framework**

Cork oak suitable areas were identified using a maximum entropy (MaxEnt) 136 137 modelling framework and implemented in library *dismo* (Hijmans et al., 2015) for R 138 software package v3.2 (R Core Team, 2016). MaxEnt modelling is a widely-used method for modelling species distributions (Merow et al., 2013) and has often been recommended 139 140 over other available methods (Elith et al., 2006). Furthermore, it has recently been shown that the MaxEnt approach is analogous to a Poisson regression and thus mathematically 141 equivalent to a Generalized Linear Modelling (GLM) approach (Renner and Warton, 142 143 2013).

144 Models were fitted using only Linear and Quadratic features to make model responses more interpretable (Merow et al., 2014; Merow et al., 2013). This approach has 145 146 also been recommended to obtain more realistic estimates of current and future potential 147 distributions (Jimenez-Valverde et al., 2008; Thuiller et al., 2004), which are important 148 for conservation purposes. Background points were selected from the whole study region, but excluded areas where the cork oak is known to be present. Models were replicated 149 150 100 times by selecting 75% of data records for calibration and 25% for validation, using a sub-sampling approach. Each model run returned a prediction of current suitable areas 151 152 for cork oak on a logistic scale varying from 0 to 1. Response curves for each 153 environmental variable are available in Fig. S1 (Online Resource 1). Future predictions were also obtained on a logistic scale based on model response curves by applying 154 environmental data representing future conditions to each model run. 155

Consensus maps were then calculated for present and future climate scenarios 156 157 using an unweighted average of the logistic predictions obtained from model replicates using the raster library (Hijmans, 2015) for R software package v3.2 (R Core Team, 158 2016). Areas with a suitability score under 0.25 were considered as potentially unsuitable 159 for the cork oak based on the average value of the 'Equal training sensitivity plus 160 161 specificity' threshold, which has been recommended as a good threshold selection approach (Jimenez-Valverde and Lobo, 2007). Suitable areas were then assigned to one 162 163 of two suitability classes based on logistic prediction scores: low (0.25 to < 0.5) and high 164 (≥ 0.5) . These maps were used to estimate range-wide and country-level changes in the extent of suitable areas between present and future scenarios as well as the percentage of 165 166 new stands likely to remain suitable in the future. The estimates of change in the extent 167 of suitable area correspond to the environmental predictions of suitable area loss (measured in total grid cell area) and not observed cork oak range extent lost. Finally, 168 169 Corine Land Cover 2006 raster maps (Version 16) were used to determine the current

land use in areas of high future suitability for the cork oak and explore the viability of
different scenarios for the establishment of cork oak plantations. All map visualization,
analysis and plotting was done using ArcGIS v10.0 (ESRI, 2011).

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174 **3 Results**

175 **3.1 Model assessment**

The final model included six environmental variables: number of frost days, spring aridity index, temperature seasonality, precipitation seasonality, soil type, and soil pH (Table 1). The model with this set of variables had the highest AUC test score (0.944) and the lowest AIC score. Number of frost days was the highest contributing environmental variable (34.9%), followed by precipitation seasonality (23.9%). The remaining environmental variables included in the model had an overall contribution to the model which was inferior to 15% (Table 2).

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184 **3.2 Global range analysis**

185 Current predictions indicate that the large majority of cork oak woodlands (~75%) 186 are in areas of high suitability across their range (Fig. 2). Future predictions indicate an 187 overall decrease in environmental suitability for the cork oak across its current range (average suitability decrease of 0.1 for RCP 4.5 and 0.3 for RCP 8.5, Fig. S2 in Online 188 Resource 1). Our results suggest that suitability will decline in most countries where the 189 species occurs (except in France) due to a northward shift of suitable environmental 190 conditions (Fig. 2). Such changes will lead to a potential loss of suitable areas that varies 191 between ~2 000 (RCP 4.5) and 13 000 km² (RCP 8.5), corresponding to ~5-40% of 192 currently occupied areas (Fig. 3). Moreover, only ~5 000 (RCP 8.5) to 11 000 km² (RCP 193 4.5), which account for ~20 to 35% of currently occupied areas, are likely to remain 194 195 highly suitable. Thus in the future a large proportion of the area currently occupied (between ~65 and 80%) is likely to have a low suitability for the species. 196

Area losses resulting from decreasing environmental suitability may be compensated with afforestation in novel suitable areas (Fig. 2). The models predict that these new suitable areas correspond to \sim 51 000 km² under the RCP 4.5 scenario, and to \sim 58 000 km² under the RCP 8.5 scenario. This represents approximately twice the area currently occupied by cork oak woodlands. However, the conversion of most of these areas to cork oak woodland faces considerable policy and socio-economic challenges including competition with current land uses and alternative management options. Most of the potential new suitable areas are currently occupied by native forests and agricultural
land (~25 000 to 30 000 km², Table 3) while areas of pasture and agro-forestry systems
do not represent more than 2 500 km² (~5% of the total novel suitable area).

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208 **3.3 Country-level analysis**

209 Analyses at country level reveal distinct trends for the northern and southern areas 210 of the cork oak distribution (Fig 2). Most southern areas where the species is currently present will decrease in environmental suitability, particularly in Portugal, Spain and 211 Morocco. In Portugal, large areas currently occupied by the species will change from high 212 to low environmental suitability. These areas total between ~ 5000 and 6000 km² under 213 RCP 4.5 and RCP 8.5 scenarios respectively (Fig. 3), which corresponds to approximately 214 55-66% of high suitability areas in the country. In Spain and Morocco many areas are 215 also likely to become unsuitable for the species. The total extent of these areas range up 216 to ~6 000 km² in Spain and ~3 500 km² in Morocco under RCP 8.5 (corresponding to 217 218 approximately 60 and 70% of currently suitable areas in Spain and Morocco respectively; Fig. 3). In Italy, Algeria and Tunisia, climate change will mostly convert areas of high 219 220 suitability into areas of low suitability. This change may range up to ~2 000, 3 000 and 800 km² for each country respectively under the RCP 8.5 scenario. France is the only 221 222 country where the cork oak environmental suitability is likely to improve in presently occupied areas, with up to 1 000 km² becoming highly suitable. 223

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225 **3.4** Assessment of the climatic suitability of recent afforestations

Many of the areas where cork oak stands were recently established in Portugal are likelly change in suitability due to climate change (Fig. 4). Our results suggest that the most (~99.5%) recent stands were established in areas that presently show high environmental suitability. Under the RCP 4.5 scenario, all stands are likely to remain in suitable areas, although approximately 30% of them will decrease in suitability. However, under the RCP 8.5 scenario, up to 90% of the new stands will have low climatic suitability and approximately 10% will become unsuitable.

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234 **4 Discussion**

235 4.1 Model assessment and predicted scenarios

Our results show that climate change is likely to affect the global distribution ofcork oak woodlands. Up to 40% of the current global distribution of these woodlands is

238 expected to become environmentally unsuitable, under the more extreme climate change scenario (3.7°C by the end of the century). An additional 40% of the current range is 239 240 likely to suffer a decline in suitability. Presently, Portugal and Spain are the countries 241 with the largest area of cork oak woodlands (Pausas et al., 2009) but approximately 60% 242 of this will lose suitability under the more extreme climate change scenarios (RCP 8.5). Southern areas of the current distribution will be the most affected, including Alentejo 243 and Algarve in Portugal, Extremadura and Andalucía in Spain and most of North Africa. 244 Cork production has an important socio-economic role and supports rural livelihoods in 245 246 these regions (Berrahmouni et al., 2007). Additionally, several cork oak woodlands in these regions have high conservation value (Correia et al., 2015a; Dias et al., 2013). To 247 248 counter the negative impact of climate change on these natural and socio-economic 249 values, it is crucial to implement climate change adaptations measures. Such measures 250 can take advantage of the opportunities to compensate predicted losses using new suitable 251 areas (e.g. northern Iberia Peninsula, France and Italy, Fig. 2).

252 This study uses a robust modelling procedure which considers data from the whole 253 cork oak distribution to predict changes in environmental suitability across the natural 254 range of the species. This procedure improves model performance and transferability and 255 is likely to produce more plausible future scenarios (Barbet-Massin et al., 2010; Thuiller 256 et al., 2004). Previous studies, based on other approaches and using limited distribution 257 data, suggested potential losses of up to 96% of environmentally suitable areas for cork 258 oak in the Iberian Peninsula (Benito Garzon et al., 2008) and a potential expansion of up 259 to 522% in Italy (Attorre et al., 2011). While our results partially agree with these 260 forecasts, they give more conservative estimates for area gains and losses due to changes in environmental suitability. We found that the inclusion of the global range of the species 261 262 in the modelling procedure was particularly important; initial models were trained 263 without data from North Africa, the driest part of the species' range, and yielded clearly 264 inadequate predictions.

Our models showed a very good fit, both in terms of AUC scores and in the high correspondence between the model response to environmental variables and the known species environmental limits. The natural distribution of cork oak trees is restricted to areas with an average annual precipitation equal or above 600 millimetres and average annual temperatures above 15°C (Pausas et al., 2009; Pereira, 2007). In Europe, cork oak distribution is partly restricted to southern regions because of its low tolerance to frequent winter frost, an important determinant of the northern limit for the species (Cavender272 Bares et al., 2005; Larcher, 2000; Pausas et al., 2009). In North Africa, however, tolerance 273 to drought is likely the main limitation for cork oak occurrence (Larcher, 2000; Pausas et 274 al., 2009). Our choice of environmental variables for modelling calibration considered 275 these environmental constraints, as recommended to obtain more robust predictions 276 (Buckley et al., 2010; Kearney and Porter, 2009; Synes and Osborne, 2011). The number of frost days (an indicator of the duration of cold spells) and precipitation seasonality 277 278 (indicating the intensity of drought spells) accounted for approximately 60% of the model's explanatory power (Table 1). Soil characteristics are also strong determinants of 279 280 cork oak distribution as the species prefers acidic soils with granite, schist, or sandy substrates (Serrasolses et al., 2009). Our models also reflected this constraint, with soil 281 282 type and pH together accounting for half of the remaining explanatory power. As with 283 any other correlative modelling procedure, our model predictions are based on the 284 characterization of the current conditions supported by the species, and therefore do not account for potential acclimatization and genetic adaptation of the species to future 285 286 conditions or novel management practices (e.g. irrigation).

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4.2 Climate change adaptation recommendations for cork oak woodlands

289 Climate change adaptation targeting cork oak woodlands require development and 290 implementation of global, national and regional-level policies. To be effective, such 291 efforts must incorporate regional differences in predicted changes in environmental 292 suitability. Strategies should be distinct (i) for the regions that are becoming mostly 293 unsuitable for the cork oak, (ii) for those that will in the long term maintain adequate 294 suitability, and (iii) for those regions that will harbour new suitable areas. For example, under the scenario of RCP 8.5, extensive regions of northern Africa and southern Iberian 295 Peninsula will lose their overall suitability; within these regions only small areas with 296 297 different micro-climatic conditions may support cork oak populations. These potential 298 refugia sites will likely be located in northern slopes of hilly areas, where impacts of 299 climate change are buffered by local conditions such as higher moisture and lower 300 temperatures (Correia et al., 2015b). These refugia should be prioritized for protection 301 because they are valuable to preserve existing biodiversity and as potential regeneration 302 islets (Benavas et al., 2008), acting as sources of propagules to colonize neighbouring 303 areas, when conditions become adequate. The protection of these refugia could also 304 complement the current network of protected areas, such as the Natura 2000 network, by 305 increasing connectivity and effectiveness. Such an approach, taking advantage of refugia,

has been suggested as an important adaptation strategy for the Mediterranean region
(Araujo et al., 2011; Klausmeyer and Shaw, 2009) and other parts of the world (Canadell
and Raupach, 2008; Heller and Zavaleta, 2009).

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309 Our models also show that large regions within the current distribution of the 310 species will remain environmentally suitable, in spite of changing local conditions. This includes central Portugal, most of southern and western Italy, and the islands of Sicily, 311 Sardinia and Corsica. In the short to medium term, these regions will remain the 312 stronghold of cork oak woodlands and will require efforts to minimize already existing 313 314 threats, as well as those driven by climate change. Promoting sustainable forest 315 management will be necessary to preserve cork oak woodlands and their associated 316 biodiversity, the delivery of ecosystems services (Bugalho et al., 2011), and can generate 317 synergies between forest adaptation and climate mitigation strategies (Ravindranath, 318 2007). Emergent mechanisms of forest certification and payment for ecosystem services (Bugalho and Silva, 2014; Bugalho et al., 2016; Dias et al., 2013) can be used to 319 320 incentivize sustainable forest management. In areas losing some level of suitability, 321 responsive management practices, such as stand irrigation, may be necessary, but that 322 will require prior ecological and economic evaluation. The cork industry in Portugal is 323 already anticipating the effects of climatic change and supporting research exploring how 324 novel management practices such as irrigation may affect the productivity of cork oak stands (Schmitt, 2016). It remains to be seen, however, whether irrigation is a cost-325 326 effective solution, particularly in drier regions, where limited water availability will be 327 exacerbated by future climate change (Barkhordarian et al., 2013; Cook et al., 2016).

The predicted northward shift of environmentally suitable areas represents an 328 opportunity for promoting the expansion, and thus compensate losses, of cork oak cover. 329 However, such a compensation process implies considerable policy and environmental 330 331 challenges. Concerted regional and national policy efforts would be needed to 332 compensate losses in North Africa and Iberia with expansion of the species in France. 333 Such efforts would probably require a Common European Forest policy framework that 334 currently does not exist (Winkel and Sotirov, 2016). Nevertheless, legal and financial 335 mechanisms for sustainable forest management and afforestation, presently under the 336 European Common Agricultural Policy, could be explored to support this potential 337 northern expansion of cork oak (Bonfiglio et al., 2016). Some natural range expansion into new environmentally suitable areas may take place, but is very limited by the current 338 339 low rates of cork oak regeneration and establishment (Acácio et al., 2007; Caldeira et al.,

2014; Pons and Pausas, 2006). Hence, afforestation will be necessary to support a 340 northward range expansion of the species (de Dios et al., 2007). This study identified 341 342 areas where these measures are more likely to succeed and demonstrates how modelling approaches can inform such decisions (Hidalgo et al., 2008; Vessella and Schirone, 2013). 343 344 For example, our results show that several recently established cork oak stands in Portugal are located in areas that will lose environmental suitability and are thus unlikely to 345 346 balance potential future losses of cork oak cover, particularly under RCP 8.5 (Fig. 4). Finally, afforestation in new areas must consider social, economic and ecologic dynamics 347 348 of present and past land uses in these areas, including local people needs (Linares, 2007; Nyong et al., 2007). Areas that are increasing in suitability for cork oak are currently 349 350 occupied by a matrix of other land uses, such as productive agricultural lands, native 351 vegetation or legally protected areas (Table 3). This competition among land uses 352 highlights the need for further studies addressing the potential socio-economic consequences of land cover changes resulting from climate change (Oliver and Morecroft, 353 354 2014).

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356 **5** Conclusions

357 Forest ecosystems are under increasing pressure from climate change worldwide, 358 and integrative forest adaptation and mitigation frameworks are necessary to address this challenge (Millar et al., 2007). Spatially explicit modelling approaches can contribute 359 360 with useful information for the design of these frameworks (Rowland et al., 2011), and 361 the models we developed illustrate this usefulness in the case of cork oak woodlands. The 362 most satisfactory models we obtained were made robust by the use of predictor variables known to be physiologically important for the cork oak, and by training the models using 363 the full range of climatic conditions presently occupied by the species. Future modelling 364 365 efforts aiming to analyse climate change impacts on forest ecosystems should also 366 incorporate economic and social criteria whenever possible (Aaheim et al., 2011; de 367 Bremond and Engle, 2014).

The results of these models indicate that climate change will cause major shifts in the global distribution of cork oak woodlands. It is urgent to start addressing such shifts with appropriate mitigation and adaptation measures, as this ecosystem is very important for the rural economy and conservation of biodiversity in much of the Western Mediterranean basin (Bugalho et al., 2011). Early implementation of adaptation strategies is particularly important for ecosystems with slow growth and maturation rates, such as 374 cork oak woodlands. Their early deployment can also be carried out in a manner that
375 stimulates synergies with climate change mitigation actions (Millar et al., 2007;
376 Ravindranath, 2007). However, the identified shifts in potential distribution of cork oak
377 occur at a scale requiring a supra national response strategy, spanning several countries
378 in Europe and Africa with very different social and economic realities, which poses
379 particular challenges.

The results of our spatial models reinforce the idea that forest adaptation strategies 380 need to be regionally adjusted (Afreen et al., 2011). The expected changes show a gradual 381 382 pattern across a latitudinal gradient, but it is possible to divide this gradient into three 383 rough regions that require partially distinct actions. In the southernmost region, which 384 includes North Africa and parts of Iberia, the expected decrease in suitability is so great 385 that resources should be concentrated in the preservation of the few refuges where 386 microclimatic conditions will keep long term suitability for the species (Dobrowski, 2011). The intermediate region corresponds to the parts of the current range of cork oak 387 388 that are likely to maintain adequate levels of suitability in the long term. This will for many decades remain the core of the range of this type of woodland, and a major objective 389 390 here should be to improve the economic and ecological sustainability of the system, so 391 that it can resist the greater aridity and the increasing competition for space resulting from 392 the expected northward migration of agriculture. Irrigation may be required in part of this region to counter aridity, but the viability of this measure is questionable due to the 393 394 increasing scarcity of water resources (Barkhordarian et al., 2013; Cook et al., 2016). It has been shown that the ecological and economic viability of deploying irrigation in 395 396 response to increasing aridity can show substantial variation across regions with Mediterranean climate (Salinas and Mendieta, 2013a, b). We strongly recommend a 397 similar assessment be made before the large-scale implementation of irrigation in the case 398 399 of cork oak. Finally, the northernmost region corresponds to the relatively vast areas that 400 are now mostly outside the range of cork oak but will become suitable as a consequence 401 of climate warming. Assisted colonization, mostly in the form of new afforestations, 402 should be a key tool for adaptation in this region, even though greatly limited by the competition of the various land use types that presently dominate the area. Adaptation 403 strategies should address these important land competition issues (Lunda and Iremonger, 404 2000). New afforestations should be planned using robust models predicting future 405 suitability, to increase their chances of long term success. 406

In spite of the growth of forest research, it was evident to us during this study that further multidisciplinary research covering the various dimensions of forest adaptation and mitigation is necessary to provide sound recommendations for the conservation of forests under climate change. In the case of cork oak woodlands, it is particularly crucial to address existing knowledge gaps associated with ecological functioning, bio-cultural heritage and regionally-specific threats in Europe and North Africa, as these may greatly constrain adaptation and mitigation strategies.

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673 Figures

- 674 Figure 1 Present distribution of the Cork oak *Quercus suber*. Areas where the species
- occurs are highlighted in black and country codes identify the countries where the species
- 676 currently occurs: Portugal (PT), Spain (SP), France (FR), Italy (IT), Morocco (MA),
- 677 Algeria (DZ) and Tunisia (TN).

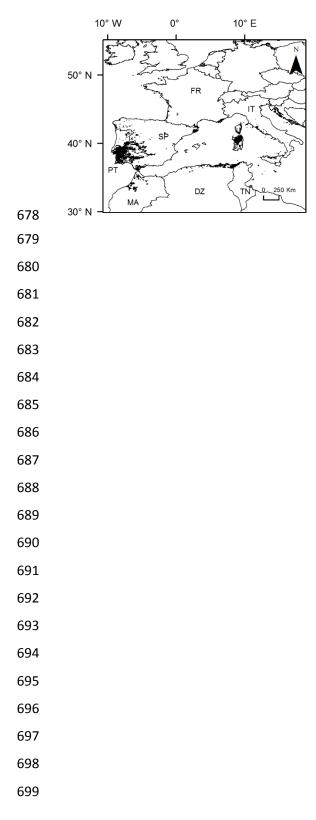
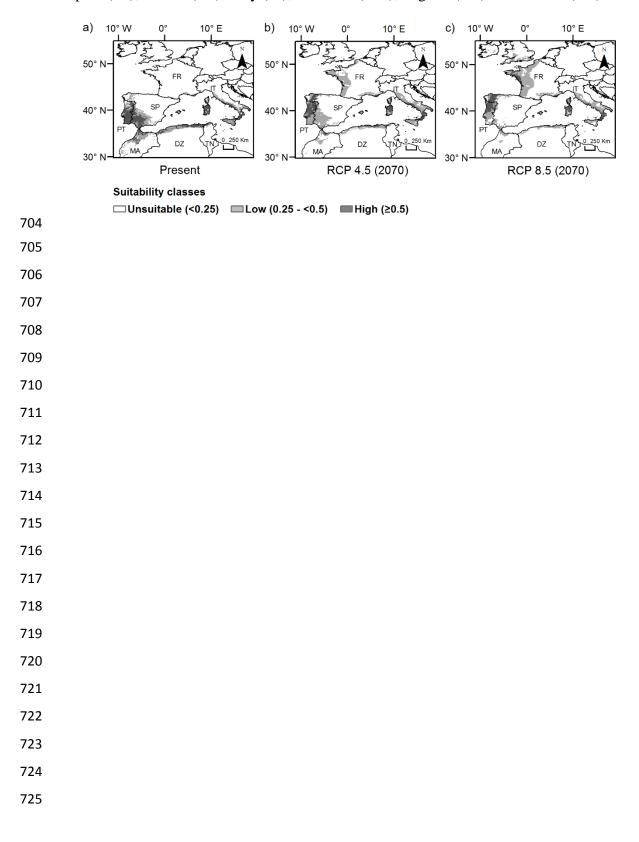


Figure 2 – Cork oak environmentally suitable areas as predicted by MaxEnt models for
present (a) and future scenarios (year 2070), based on RCP 4.5 (b) and RCP 8.5 (c).
Country codes identify the nations where the species currently occurs: Portugal (PT),
Spain (SP), France (FR), Italy (IT), Morocco (MA), Algeria (DZ) and Tunisia (TN).



- Figure 3 Country-level analysis showing changes in environmental suitability for areas
- currently occupied by cork oak. Predictions are shown for the present (left bars) and future
- (year 2070) scenarios based on RCP 4.5 (middle bars) and RCP 8.5 (right bars).

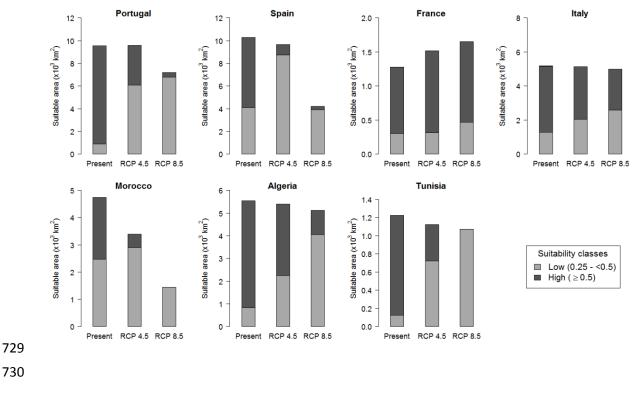


Figure 4 – Location of recently established cork oak stands in Portugal in relation to cork
oak environmental suitability as predicted by MaxEnt models for present (a) and future
(year 2070) environmental conditions based on RCP 4.5 (b) and RCP 8.5 (c) climate
change scenarios. Cork oak stands are represented by black dots.

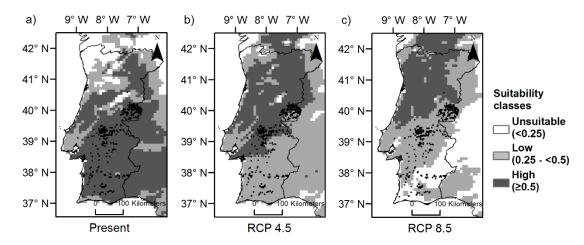




Table 1 – List of environmental variables tested for modelling the cork oak distribution. This set

of environmental variables was selected based on biological knowledge of the species (Pausas et

al. 2009). Variables included in the final model are highlighted in bold.

Environmental variable	Variable code	Description		
Number of frost days Frost		Number of frost days in a year, calculated		
		following New et al. (2000).		
Minimum temperature of	T_Min	Minimum temperature of the coldest month,		
coldest month		obtained from Worldclim database.		
Total annual precipitation	P_total	Total annual precipitation, obtained from		
		Worldclim database.		
Total spring precipitation	P_spr	Total spring precipitation, calculated as the sum of		
		precipitation for the months of March, April and		
		June		
Total winter precipitation	P_win	Total winter precipitation, calculated as the sum of		
		precipitation for the months of December, January		
		and February		
Annual aridity index	Arid	Annual aridity index, calculated following Zomer		
		et al. (2008).		
Spring aridity index	Arid_spr	Aridity index calculated for the months of		
		March, April and May.		
Winter aridity index	Arid_win	Aridity index calculated for the months of		
		December, January and February.		
Temperature	T_seas	Calculated as the standard deviation of mean		
seasonality		daily temperatures * 100, obtained from		
		Worldclim database.		
Precipitation seasonality P_seas		Calculated as the coefficient of variation		
		weekly precipitation estimates. Obtained from		
		Worldclim database.		
Soil type class	Soil_class	Major soil grouping classes, obtained from the		
		Harmonized World Soil Database		
		(FAO/IIASA/ISRIC/ISS-CAS/JRC, 2009).		
Soil pH class	Soil_ph	Soil pH classes, obtained from the Harmonized		
		World Soil Database (FAO/IIASA/ISRIC/ISS		
		CAS/JRC, 2009).		

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778 Table 2 – Environmental variables included in the final MaxEnt model and their contributions to

the model predictions.

Environmental variable	Percent contribution	Permutation importance
Number of frost days	34.9	39.1
Precipitation seasonality	23.9	30.5
Soil type class	12.4	2.4
Soil pH class	10.7	3.6
Temperature seasonality	9.4	3.5
Spring aridity index	8.6	20.9

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Table 3 – Percentage cover of the five most common land uses in areas likely to become
environmentally suitable (medium and high suitability classes) for cork oak in the future
(year 2070) under RCP 4.5 and RCP 8.5 scenarios. Land cover classes according to
Corine Land Cover 2006 Label 3 (Version 16) classification. Area values are shown in
square kilometers and percentage values relate to total novel area of medium and high
suitability where the species is currently absent.

Rank	Representative Concentration Pathways							
	RCP 4.5		RCP 8.5					
	Land cover class	Area (%)	Land cover class	Area (%)				
1	Transitional woodland-shrub	6 633 (13.0%)	Complex cultivation patterns	6 780 (11.9%)				
2	Broad-leaved forest	6 103 (11.9%)	Transitional woodland-shrub	6 582 (11.6%)				
3	Complex cultivation patterns	4 720 (9.2%)	Non-irrigated arable land	6 010 (10.6%)				
4	Agriculture with natural vegetation	4 437 (7.0%)	Broad-leaved forest	5 949 (10.5%)				
5	Non-irrigated arable land	3 982 (4.8%)	Agriculture with natural vegetation	3 619 (5.5%)				
-	Other land uses	25 336 (49.5%)	Other land uses	27 123 (47.7%)				