



Impact of a mouth parasite in a marine fish differs between geographical areas

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Considerable variation exists in parasite virulence and host tolerance which may have a genetic and/or environmental basis. In this article, we study the effects of a striking, mouth-dwelling, blood-feeding isopod parasite (*Ceratothoa italica*) on the life history and physiological condition of two Mediterranean populations of the coastal fish, *Lithognathus mormyrus*. The growth and hepatosomatic index (HSI) of fish in a heavily human-exploited population were severely impacted by this parasite, whereas *C. italica* showed negligible virulence in fish close to a marine protected area. In particular, for HSI, the parasite load explained 34.4% of the variation in HSI in the exploited population, whereas there was no significant relationship (0.3%) between parasite load and HSI for fish in the marine protected area. Both host and parasite populations were not differentiated for neutral genetic variation and were likely to exchange migrants. We discuss the role of local genetic adaptation and phenotypic plasticity, and how deteriorated environmental conditions with significant fishing pressure can exacerbate the effects of parasitism. © 2012 The Linnean Society of London, *Biological Journal of the Linnean Society*, 2012, ••, ••–••.

ADDITIONAL KEYWORDS: *Ceratothoa* – connectivity – life history – plasticity – Sparidae – virulence.

INTRODUCTION

Parasites occur in all living taxa and encompass at least one-third of all eukaryotic life on earth (De Meeus & Renaud, 2002), giving rise to an astounding range of adaptations fitting to the vast diversity of host environments (Cornell, Desdevises & Rigby, 1999). Parasites are credited with a vital role in the evolution of their hosts (Boots *et al.*, 2009), as the loss in fitness caused by parasite exploitation results in counter-adaptations in the host, ultimately resulting in an ‘arms race’ that drives evolutionary changes in both interacting organisms (Hochberg, Michalakis & de Meeus, 1992; Møller, Martin-Vivaldi & Soler, 2004; Paterson *et al.*, 2010).

Genetic and phenotypic variations in both host and parasite are responsible for the occurrence

of coadaptive processes (Gandon *et al.*, 1996). The exploitation of the host by a parasite can trigger immunological responses (Simkova *et al.*, 2008), which, in turn, can have an impact on the overall metabolic balance and compromise the investment in life history traits, such as growth, survival and reproduction (Møller *et al.*, 2001). Genetic variation and immunological responses in the host will, in turn, influence the fitness and the life history traits of the parasite, causing changes in the levels of virulence (defined as the amount of damage a parasite causes to its host). However, importantly, ecological circumstances can also govern virulence evolution (Day & Proulx, 2004), and genomic interactions between host and parasite can occur in environmental contexts that may dramatically change both spatially (Thompson & Cunningham, 2002; Gandon & Nuismer, 2009) and temporally (Brooks & Hoberg, 2007). Therefore, the analysis of the impact of different environmental scenarios on interacting species is key to our understanding of host–parasite coevolution.

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Theoretical studies have shown that host–parasite coevolution can be highly dynamic, depending, for example, on the migration rates of both hosts and parasites (Gandon & Michalakis, 2002). Given that gene flow and selective factors can vary considerably across geographical ranges, host–parasite associations are typically heterogeneous across their spatial distribution (Thompson, 1999; Gomulkiewicz *et al.*, 2000; Gandon & Nuismer, 2009), generating a dynamic adaptive landscape that depends on the interaction of selection, gene flow and drift, which can result in a geographical mosaic of coevolutionary hotspots and coldspots (Thompson & Cunningham, 2002). The majority of these studies have focused their attention on the parasite’s capability to harm its host and evolve rapidly in response to the defence mechanisms generated by the host, increasing with this the parasite’s virulence and the ability of the parasite to harm its host (Nuismer, Thompson & Gomulkiewicz, 2000; Gandon & Nuismer, 2009).

Several empirical investigations have focused on the effects of parasites on the life history and fitness of hosts, especially in fish (van Oosterhout *et al.*, 2007; van Oosterhout, 2008; Fogelman, Kuris & Grutter, 2009; Blanchet, Rey & Loot, 2010), but these have seldom been contrasted with explicit environmental scenarios (but see Wolinska & King, 2009). The aim of this study was to examine the life history effects of a mouth-dwelling, blood-feeding isopod parasite, the cymothoid *Ceratothoa italica* (Schioedte & Meinert, 1883), on two Mediterranean stocks of a coastal benthopelagic teleost, the striped sea bream (*Lithognathus mormyrus* L.). The stocks show indistinguishable gene pools and parasite communities (Sala-Bozano, Ketmaier & Mariani, 2009), but are found under very different environmental pressures. We examined differences in life history descriptors (weight and length at age) and a fitness-related trait that is associated with metabolic activity (hepatosomatic index, HSI) in the host populations, in relation to the parasitic infection and other interacting factors, such as age and sex. We demonstrate temporally stable differences in parasite impact in a scenario of high gene flow, underscoring the importance of environmental factors in host–parasite coevolution.

MATERIAL AND METHODS

STUDY AREA AND SAMPLING

A total of 626 *Lithognathus mormyrus* were collected during the spawning season, between May and September, in 2006, 2007 and 2008. Sampling areas were representative of the main marine basins in which the species is found (East Atlantic Ocean, ATL; Alborán Sea, ALB; Balearic Sea, BA; Tyrrhenian Sea,

TYR; Adriatic Sea, ADR; Aegean Sea, AEG). Four localities were sampled in both 2006 and 2007, a fifth site was only sampled in 2007 (the Adriatic Sea) and a sixth in 2008 (Aegean Sea) (Fig. 1). Our study focuses, in particular, on the BA and TYR fish, which inhabit areas that are climatically comparable. However, the BA fish were caught close to a marine protected area, whereas the TYR fish occur in a habitat with negligible fishery regulation and greater harvesting pressure. Adult fish (> 15 cm) were collected by fishermen using trammel and gill nets, and juveniles (< 15 cm) were collected by anglers using hook and line. Fishermen and anglers donated dead adult fish and juveniles to this study; hence, no special licences were required.

INFECTION AND LIFE HISTORY PARAMETERS

Fish were measured to the nearest millimetre (standard length, L_S) and weighed (wet weight, W_W) to the nearest gram. They were dissected to identify their sex, and their livers were removed and weighed to the nearest 0.01 g. To assess the physiological condition of the fish, HSI (Lloret & Planes, 2003) was calculated as $HSI = W_L/W_W$, where W_L is the wet liver weight.

Scale reading was used to age the fish (Suau, 1970; Kraljevik *et al.*, 1996). For each fish, a minimum of five scales was removed from above the lateral line and used for reading under a stereomicroscope with $\times 400$ magnification. Two independent readers examined five scales per fish, showing high repeatability in age estimates between observers (correlation coefficient $r > 0.99$, $P < 0.001$).

The buccal cavity and gill arches of each fish were examined to detect the presence of *C. italica*. No *Ceratothoa* species possess a planktonic larval stage, and hence newly released juveniles can only swim short distances, ‘hopping’ among individual hosts and briefly between the substrate and hosts (Horton, 2000). Juveniles then usually enter from the gill cavity and crawl to settle on the tongue of the host. The first individual to settle grows and matures into a large female, which is fertilized by one or more small males, infecting the host after the female. In the case of death of the female, the largest male undergoes a sex change to replace it (Horton & Okamura, 2003). The number of adult *C. italica* was recorded for each infected individual, and the prevalence (infected fish/total number of fish) of parasite infection was calculated for each infected stock.

GENETIC DIFFERENTIATION OF *L. MORMYRUS*

Fin tissue was used to isolate DNA using a modified salt/chloroform extraction protocol (Miller, Dykes & Polesky, 1988; Petit, Excoffier & Mayer, 1999). All

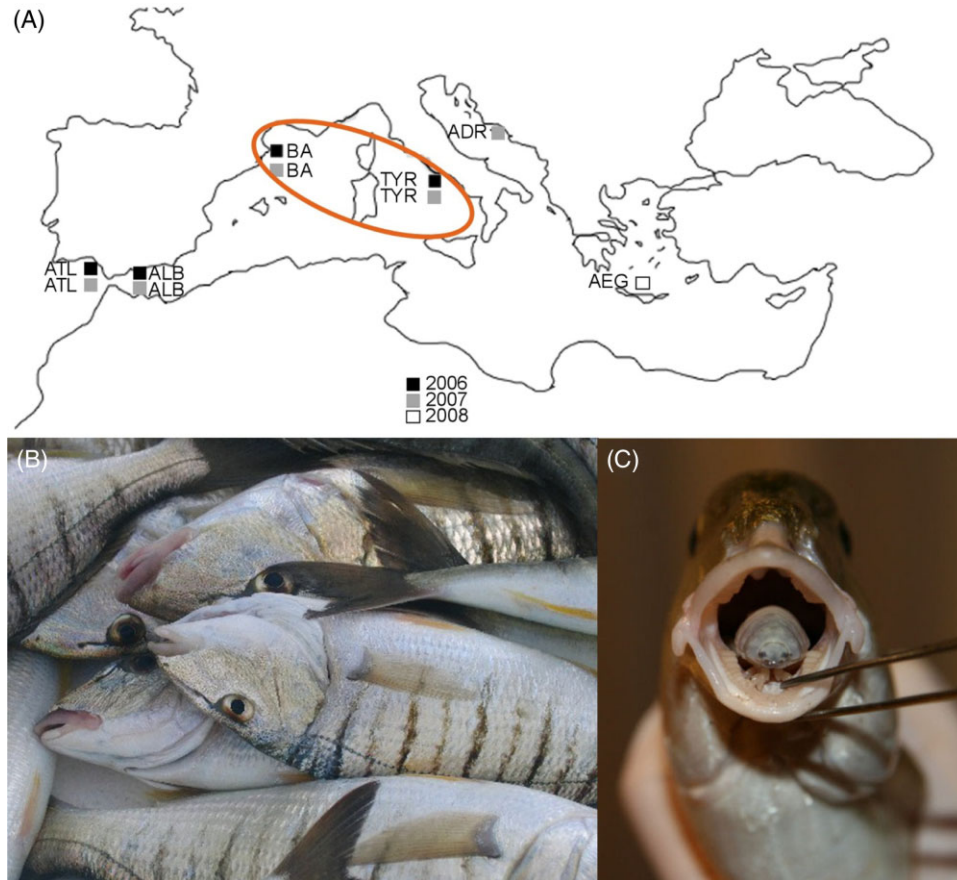


Figure 1. Sampling locations and species of interest: A, map indicating the six sampling locations; the ellipse circumscribes the populations infected by *Ceratothoa italica*; B, the striped sea bream, *Lithognathus mormyrus*; C, a live *C. italica* inside the mouth of its host. ADR, Adriatic Sea; AEG, Aegean Sea; ALB, Alborán Sea; ATL, East Atlantic Ocean; BA, Balearic Sea; TYR, Tyrrhenian Sea.

individuals were polymerase chain reaction (PCR) amplified and genotyped at nine polymorphic microsatellite loci: Lm68, Lm72, Lm19, Lm86, Lm12, Ad05, Ad66, SaL15 and SaL19 (Brown *et al.*, 2005; Franch *et al.*, 2006). Samples were processed in two multiplex reactions (see Sala-Bozano, Tsalavouta & Mariani, 2008 for details). Genotyping of individuals was performed by allele sizing on an ABI 3130xl Genetic Analyser (Applied Biosystems) using forward primers labelled with NED, PET, FAM and VIC dyes and an internal size standard labelled with LIZ 600 (Applied Biosystems). The software GENEMAPPER version4x (Applied Biosystems) was used to score alleles. The data were checked for the presence of null alleles, large allele drop out, errors caused by stutter peaks and possible scoring errors using MICROCHECKER (van Oosterhout *et al.*, 2004).

Expected and observed heterozygosity (H_E and H_O), allelic richness (A_R) and linkage disequilibrium were calculated using FSTAT version 2.9.3 (Goudet, 2001). FSTAT was also used for F_{ST} analysis, and pairwise

F_{ST} values were tested for their significance with 10 000 permutations. A multidimensional scaling (MDS) ordination, inferred from the pairwise F_{ST} matrix, was employed to visually represent the genetic relationships among population samples.

Historical effective population sizes (N_e) and rates of gene flow were estimated only for infected stocks using a Bayesian approach implemented in the software MIGRATE-n version 2.4 (Beerli & Felsenstein, 1999, 2001). This software estimates theta (θ), which is equal to four times the effective population size multiplied by the mutation rate μ ($\theta = 4N_e\mu$), and a migration rate parameter M , which is equal to the immigration rate m divided by the mutation rate. M quantifies the number of new alleles introduced into the population by immigration relative to mutation (Beerli & Felsenstein, 1999, 2001).

The θ values were translated into N_e estimates using a microsatellite mutation rate of 5×10^{-4} , which is the most frequently employed rate in fish (Barson, Cable & van Oosterhout, 2009). Ten short chains,

each with 500 generations and a sampling increment of 200 generations, and three long chains, each with a total of 5000 generations and a sampling increment of 20 generations, were run. The chains visited a total of 100 000 and 1 000 000 genealogies, respectively (recorded steps multiplied by the sampling increment). The first 10 000 genealogies were discarded (burn-in). MIGRATE was run three times, until N_e and N_{em} estimates were consistent between runs. The first run used F_{ST} -based estimates as the starting point. Subsequent runs used the results of the previous run as starting values.

GENETIC VARIATION IN *C. ITALICA*

Genetic differentiation between *C. italica* parasite populations was examined by analysing a portion of the mtDNA cytochrome oxidase subunit I (COI) gene. Sixteen individuals per population were sequenced. Genomic DNA was extracted using the same protocol as for *L. mormyrus* from single legs of ethanol-preserved specimens. PCR amplifications of a 592-bp COI fragment were carried out using universal invertebrate primers (Folmer *et al.*, 1994). The PCR recipe and conditions were similar to those of Ketmaier *et al.* (2008). In our study, however, more stringent conditions were used, by increasing the annealing temperature by up to 14 °C and by using high-fidelity *taq*-polymerase (Invitrogen, Platinum).

Products were sequenced in both directions by MacroGen Inc. (Seoul, South Korea). Chromatogram contigs were edited and assembled in Sequencher 4.2 (Gene Codes Corporation, Ann Arbor, MI, USA); sequences were aligned in MEGA-4 (Tamura *et al.*, 2007). Nucleotide (π) and haplotype (H) diversity were calculated using DnaSP 4.10.9 (Rozas *et al.*, 2003). The software ARLEQUIN version 3.5 (Excoffier, Laval & Schneider, 2005) was employed to calculate haplotype frequencies and to estimate variance in haplotype frequencies between samples, using 10 000 permutations to test for statistical significance.

COI sequences were employed to construct a median-joining network in Network 5.5.1 (Bandelt, Forster & Röhl, 1999) using the default settings in order to illustrate graphically the relationship between the two populations.

PARASITE LOAD

We used ordinal logistic regression to examine the factors explaining the variation in parasite load between individual *L. mormyrus*. In this model, the 'number of parasites' was used as response variable, and the population of origin ('Pop'), sex and age were employed as predictor variables. The logit link function was used to calculate the mean odds ratios (the

odds of the infection occurring in one group to the odds of it occurring in another group) and their 95% confidence intervals (CIs). The log-likelihood from the maximum likelihood iterations and the G statistic were used to examine whether all the slopes were significantly different from zero. The logit variable z , which expresses the total contribution of the independent variables used in the model, was also noted.

TESTING PARASITE IMPACT

Size (L_S) and weight (W_W) at age, and HIS, were used as estimates of the biometric and metabolic condition of the fish. The primary objective was to examine whether there was a difference between the effects of *C. italica* infection on the two host populations. Thus, we first quantified differences in age structure and sex ratios between populations, in order to take into account the variation in size and HSI explained by these factors, and to assess the effect of the parasite on the residual variance.

Differences in size (L_S) and weight (W_W) between the two infected stocks were analysed with a general linear model (GLM). In the GLM, log-transformed L_S and log-transformed W_W were used as response variables, sex, infection and 'Pop' were used as fixed crossed factors and age was employed as a covariate. We discarded data of fish older than 4 years as there were too few data points in the data from TYR. The same approach was also used to analyse whether size differences existed between groups of fish without *C. italica* infection. For this analysis, we used the standard length and weight of the uninfected samples only ($N = 138$ and $N = 36$ for BA and TYR, respectively) as response variables.

Differences in the metabolic condition of parasitized fish were tested by examining the variation in HSI as response variable, and sex, age and 'Pop' as explanatory variables. The explanatory variables were crossed in two- and three-way interactions. In this model, age was used as a covariate. In all GLMs, we used a backwards elimination of nonsignificant two-way and three-way interactions, and the minimum adequate model is presented in Table S1 (see Supporting Information). Standardized residuals were calculated and used in regression analyses to examine whether residual variation in HSI could be explained by parasite load, and to test whether the impact of parasites on HSI differed between stocks. All statistical analyses were performed using MINITAB 15 (Software Minitab Inc., State College, PA, USA).

RESULTS

GENERAL INFECTION CHARACTERISTICS

Two hundred and twenty-eight adult *C. italica* were recovered from 110 fish. The only stocks infected were

Table 1. Sample information for the infected localities Balearic Sea (BA) and Tyrrhenian Sea (TYR)

Locality	Code	Basin	GPS coordinates		N	H_E	A_R
			Latitude	Longitude			
L'Estartit 06	BA06	Balearic	42°02'10.18"N	3°12'22.03"W	95	0.837	12.095
L'Estartit 07	BA07	Balearic	42°02'10.18"N	3°12'22.03"W	99	0.835	11.775
Foce Verde 06	TYR06	Tyrrhenian	41°24'0.0"N	41°24'0.0"W	40	0.829	11.269
Foce Verde 07	TYR07	Tyrrhenian	41°24'0.0"N	41°24'0.0"W	50	0.830	12.453

GPS coordinates, sample size (N) and genetic diversity indices inferred from microsatellites (H_E , expected heterozygosity; A_R , allelic richness) are provided.

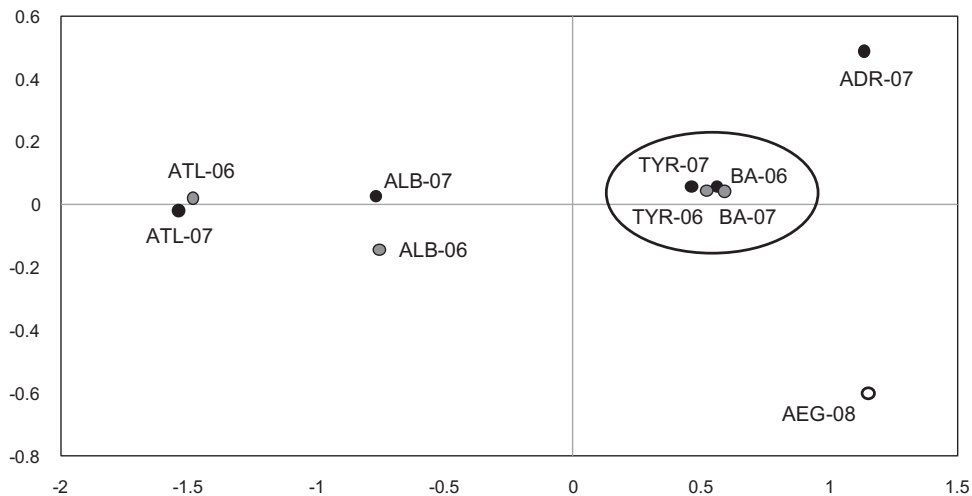


Figure 2. A multidimensional scaling (MDS) plot of F_{ST} pair-wise values among population samples screened for microsatellite variation. Location codes are as in Figure 1. The ellipse highlights the temporally stable genetic identity of TYR and BA.

within the western Mediterranean, in the BA and TYR basins (Fig. 1, Table 1). Parasite prevalence was significantly higher in the TYR (47%; 52 infected fish of a total of 90) than in the BA (30%; 58 of 194 fish) population sample ($\chi^2 = 20.139$, d.f. = 1, $P < 0.0001$).

GENETIC DIFFERENTIATION

Global F_{ST} over all populations examined was strong and highly significant (0.036; CI, 0.009–0.045) (Fig. 2). However, the overall genetic differentiation of the two infected *L. mormyrus* stocks, TYR and BA, over 2 years was low and not significant ($F_{ST} = 0.002$; $P = 0.62$), with none of the pairwise comparisons significantly different from zero. The allelic richness (A_R) and expected heterozygosity (H_E) were also similar (Table 1), with no significant departures from Hardy–Weinberg equilibrium or deviations from linkage equilibrium. This pattern was congruent with the results from MIGRATE: the number of migrants per

generation ($N_e m$) going from the TYR population into BA was 1.2 (95% CI, 1.04–2.17), and the number of migrants going from BA into TYR was 13.1 (95% CI, 12.28–17.04). The historical effective population size (N_e) for BA was 295 (95% CI, 290–325) and for TYR was 280 (95% CI, 275–330).

Only four nucleotide sites of the 597 bp sequenced in 32 *C. italica* individuals were polymorphic, corresponding to only three unique haplotypes (Fig. S1, see Supporting Information). Despite the fact that more specific DNA isolation and more stringent PCR conditions were used, a stop codon in position 481 was invariably found in all sequences. This particular stop codon was also found, in the same position, in another species of the same genus (*Ceratothoa oestroides*) by Mladineo, Šegvić & Grubišić (2009), who also showed that this did not affect the usefulness of the marker. Overall haplotype frequency did not differ between TYR and BA ($F_{ST} = 0.024$, not significant). Haplotype frequencies for haplotypes I, II and III did not differ

Table 2. Ordinal logistic regression

Predictor	Coefficient	SD	<i>z</i>	<i>P</i> value	Odds ratio (and 95% CI)
'Pop'	-1.0914	0.3111	-3.51	< 0.001	0.34 (0.18–0.62)
Sex	0.9609	0.3268	2.94	0.003	2.61 (1.38–4.96)
Age	1.6793	0.2280	7.36	< 0.001	5.36 (3.43–8.38)

The number of parasites was used as the response variable and the population of origin ('Pop'), sex and age were used as independent factors. Significant variation in individual parasite load is explained by the independent variables in the model: log-likelihood = -197.911; test that all slopes are zero: $G = 151.97$, d.f. = 3, $P < 0.001$. The negative z value and the odds ratio below unity for 'Pop' indicate that the risk of parasite infection is higher in the Tyrrhenian (TYR) than Balearic (BA) population. Similarly, males and younger fish have a higher infection risk than females and older fish, respectively. CI, confidence interval; SD, standard deviation.

between TYR and BA ($\chi^2 = 0.051$, d.f. = 1, $P = 0.821$, $\chi^2 = 0.031$, d.f. = 1, $P = 0.858$ and $\chi^2 = 0.025$, d.f. = 1, $P = 0.873$, respectively).

FACTORS EXPLAINING PARASITE LOAD

Ordinal logistic regression showed that significant variation in parasite load was explained by 'Pop', sex and age (Table 2). The odds ratio revealed that the parasite load was reduced significantly with age, that male fish had between two to three times higher parasite load than females, and that fish in TYR had an almost three times higher load than those in BA (Table 2).

EFFECTS OF INFECTION ON HOST TRAITS

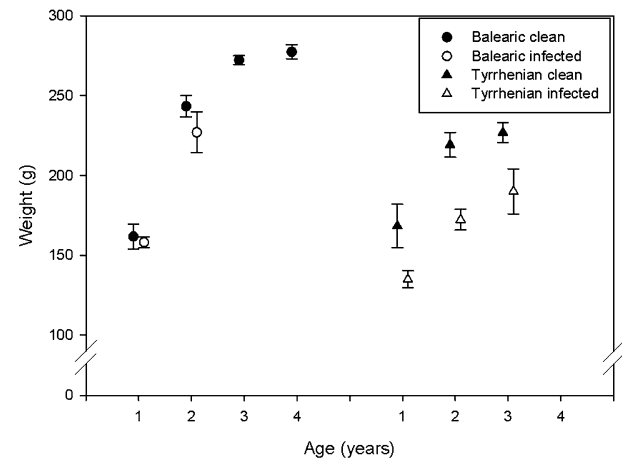
There was a significant male-biased sex ratio in both samples (BA, male : female = 106 : 81; TYR, male : female = 57 : 27). Although this skew did not differ significantly between groups ($\chi^2 = 2.570$, d.f. = 1, $P = 0.109$), the male-biased sex ratio was more prominent in TYR (binomial probability: $P = 0.0007$) than in BA (binomial probability: $P = 0.039$).

Size (GLM $F_{1,266} = 126.03$, $P < 0.001$) and weight (GLM $F_{1,265} = 117.69$, $P < 0.001$) were significantly lower in infected fish (see Fig. 3). The model showed that, although age and 'Pop' explained significant variation in fish size and weight, there was no difference in size or weight between the sexes when accounting for age differences (Table 3). Furthermore, the impact on weight caused by *C. italica* infection appeared to be more severe in TYR than in BA (Fig. 3), which explains the significant interaction term 'Pop' \times infection (GLM $F_{1,265} = 5.63$, $P = 0.018$) (Table 3).

Next, we analysed and compared the size of the uninfected fish between samples. This showed that, irrespective of infection, TYR fish were significantly smaller than their BA counterparts of the same age

Table 3. General linear model (GLM) with natural log-transformed wet weight as response variable, infection, population of origin ('Pop') and sex as fixed factors, and age as covariate

Factor	d.f.	MS	<i>F</i>	<i>P</i>
Age	1	31.840	117.69	< 0.001
Sex	1	0.103	0.38	0.539
'Pop'	1	5.774	21.34	< 0.001
Infection	1	34.523	127.60	< 0.001
'Pop' \times infection	1	1.522	5.63	0.018
Error	265	0.271		
Total	270			

**Figure 3.** Mean (\pm SE) wet weight across year classes of clean (filled symbols) and infected (open symbols) Balearic (circles) and Tyrrhenian (triangles) fish. Fish infected by *Ceratomyxa italica* parasites are significantly smaller than clean fish, and this effect is significantly more pronounced in the Tyrrhenian (see main text and Table 3 for details on statistical analysis).

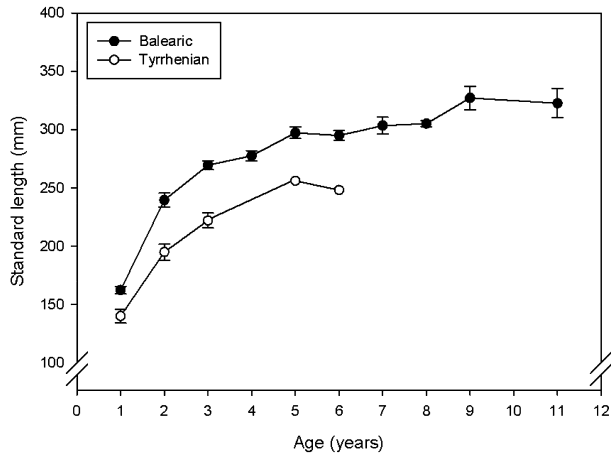


Figure 4. Mean (\pm SE) standard length across year classes of uninfected Tyrrhenian (open symbols) and Balearic (filled symbols) fish. The Tyrrhenian fish are significantly smaller than fish from the Balearic (see main text for details).

Table 4. General linear model (GLM) with hepatosomatic index (HSI) as response variable, age and parasite load as covariates, and sex and population of origin ('Pop') as fixed factors

Factor	d.f.	MS	<i>F</i>	<i>P</i>
Age	1	0.0001792	10.12	0.002
Sex	1	0.0007773	43.87	< 0.001
'Pop'	1	0.0006453	36.42	< 0.001
Parasite load	1	0.0006317	35.65	< 0.001
'Pop' \times parasite load	1	0.0003364	18.99	< 0.001
Error	265	0.0000177		
Total	270			

The significant interaction between 'Pop' and parasite load shows that the effect of parasite infection on HSI differs significantly between populations.

(GLM $F_{1,164} = 9.17$, $P = 0.003$) (Fig. 4), confirming that BA is a more favourable environment than TYR for fish growth.

Finally, we analysed the impact of parasitism on HSI. GLM showed that age, sex, 'Pop' and parasite load all explained significant variation in HSI, and that there was a highly significant 'Pop' \times parasite load interaction (Table 4). Similar to the analysis of parasite infection on weight, the effects of *C. italyca* on HSI appeared to differ significantly between stocks. To examine this further, we omitted parasite load and the interaction term from the model and calculated the standardized residuals, which were subsequently regressed against the parasite load. In TYR, the parasite load explained 34% of the residual

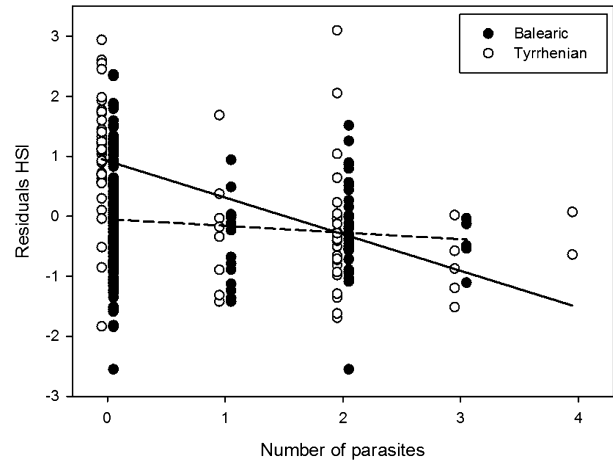


Figure 5. Standardized residuals of hepatosomatic index (HSI) regressed against the number of parasites per individual for the Tyrrhenian (open symbols, full trend line) and Balearic (filled symbols, broken line) fish. Only for the Tyrrhenian fish is there a significant decline in HSI with increasing parasitic load (see main text for details).

variation in HSI after removing the effects of age, sex and population (regression: $r^2 = 34.4\%$, $F_{1,82} = 43.01$, $P < 0.001$), but, in BA, this was reduced to an insignificant 0.3% (regression: $r^2 = 0.3\%$, $F_{1,185} = 0.64$, $P = 0.424$) (Fig. 5).

In other words, *C. italyca* parasite infection appears to have a significantly worse physiological impact on TYR fish.

DISCUSSION

This study examined the infection prevalence and impact of the parasite *C. italyca* in the striped sea bream, *L. mormyrus*. The two host populations found to be infected by the mouth-dwelling isopod were genetically indistinguishable on the basis of neutral markers and overall parasitic fauna (Sala-Bozano *et al.*, 2009), but the prevalence of *C. italyca* infection, parasite load and impact on host weight were significantly greater in TYR than in BA. In addition, the parasite load differed as a function of the host sex and age. Most importantly, the parasite's impact was greater in TYR, and 34.4% of the variation in HSI of fish from this area was explained by the *C. italyca* parasite load, whereas parasite infection had no discernible effect on fish in the BA stock. The striking spatial variation in parasite effects between two apparently similar populations could be explained by both local genetic adaptation and phenotypic differences caused by environmental factors, and we evaluate the evidence for both explanations here.

The lack of genetic differentiation at neutral markers between BA and TYR fish is in stark contrast

with the levels of genetic differentiation observed previously among all other *L. mormyrus* populations that were analysed with the same set of genetic markers (Sala-Bozano *et al.*, 2009). This suggests that the lack of differentiation is unlikely to be an artefact of the microsatellite markers used (e.g. size homoplasy; Estoup, Jarne & Cornuet, 2002), but rather reflects actual connectivity between these fish stocks (Fig. 2). Analysis with the software MIGRATE showed that the numbers of migrants per generation (N_m) being exchanged between TYR and BA is considerable ($N_m = 1.2$ for TYR to BA, and $N_m = 13.07$ from BA to TYR). Evidence for high genetic exchange was further corroborated by the mtDNA data of the parasite, which showed that the *C. italica* BA and TYR populations have virtually identical haplotype frequencies. In addition, *C. italica* is an obligate blood-feeder with poor swimming ability (Horton & Okamura, 2003), which depends on its hosts to disperse. Hence, the observation that – at least on *L. mormyrus* – *C. italica* was exclusively present in both focal stocks and was not observed in any of the other four populations surveyed, further supports that both stocks are likely to be demographically connected by migration.

A high rate of gene flow does not necessarily preclude the existence of local genetic adaptation, as reviewed recently in marine fish (Nielsen *et al.*, 2009). However, local genetic differentiation and adaptations are more likely to be eroded under medium/high levels of migration (Hendry, Day & Taylor, 2001; Räsänen & Hendry, 2008). Assuming a common host–parasite coevolutionary model, this is particularly the case at immune and virulence genes, given that the effective migration rate is considerably higher for genes under balancing selection (Schierup, Mikkelsen & Hein, 2001; Muirhead, Glass & Slatkin, 2002). The divergence of such genes requires genetic isolation and strong genetic drift (Miller, Allendorf & Daugherty, 2010), and this does not seem to be supported by our genetic data of both the host and parasite.

Environmentally induced differences in fitness between host populations can also affect parasite tolerance (Thompson & Cunningham, 2002; Blanchet *et al.*, 2010; Kohler *et al.*, 2010). For instance, different environmental circumstances can induce plasticity in the host's condition and growth (Wild, Costain & Day, 2007), which consequently can result in a more severe impact of the parasite on the host. Parasite virulence can also be phenotypically plastic in response to variability in the life history of the host (Nagasawa, 2004; Frank & Schmid-Hempel, 2008). The parasite's virulence is expected to increase with reduced life expectancy of the host, as this will shorten the period of time that parasites can exploit

their host (Day, Gaham & Read, 2007) and reproduce before its death (Frank, 1996). This may be particularly relevant to *Ceratothoa* species, which exhibit a relatively long life cycle (Garrey & Maxwell, 1982).

Evidence for environmentally induced differences between BA and TYR stocks was detected when comparing the size and growth rate of noninfected fish. TYR fish were smaller and lighter than those from BA, even when they were not infected, indicating a population under the influence of environmental stresses, beyond the parasitic infection. Moreover, fish from TYR have been shown to mature at significantly smaller sizes than those from BA (Sala-Bozano & Mariani, 2011). Smaller size and earlier maturation in TYR fish are consistent with the existence of compensatory trade-offs between growth and development and reproductive output in a scenario of increased extrinsic mortality (Kuparinen & Merila, 2007). The studied BA and TYR areas are climatically comparable, but are under different local conditions. BA fish were caught close to a marine protected area (Medes Islands), the establishment of which has been shown to greatly benefit the local fish populations (Goñi *et al.*, 2008). TYR fish, however, are native to a more heavily impacted area, with negligible fishery regulation and greater harvesting pressure. This is also supported by FAO catch data (FAO, 2009), which show that the total catches of *L. mormyrus* in Italy have exceeded those in Spain in recent years. Greater fishing pressure may have had an impact on the condition, growth, maturation and sex change (Sala-Bozano & Mariani, 2011) of TYR fish, which could have exacerbated the pathogenic effects of *C. italica* parasite infection. This explanation is in agreement with previous studies which have found that the fitness costs of tolerance and/or resistance are dependent on the environmental conditions (Sandland & Minchella, 2003; Zibiden, Haag & Ebert, 2008), and that these costs can be significantly higher under stressful conditions (Raymond, Sayyed & Wright, 2005; Gassman *et al.*, 2006; Schwarzenbach & Ward, 2006).

Overall, our data appear to be most consistent with the hypothesis that human-induced environmental differences affect the fitness and life history of host populations with cascading effects on parasite impact. Thus, with some caution, by pinpointing the environmental differences that drive life history variation, it may become possible to predict the direction and rate of evolution (Mennerat *et al.*, 2010). As well as providing additional empirical evidence of geographical variation in host–parasite coevolutionary dynamics (Gandon *et al.*, 1996; Thompson, 1999; Gandon, 2002; Thompson & Cunningham, 2002; Gandon & Nuismer, 2009), this study also warrants a more judicious assessment of the environmental context of biological interactions, especially in the case of

exploited resources facing increasingly perturbed natural habitats.

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REFERENCES

- Bandelt HJ, Forster P, Röhl A. 1999.** Median-joining networks for inferring intraspecific phylogenies. *Molecular Biology and Evolution* **16**: 37–48.
- Barson NJ, Cable J, van Oosterhout C. 2009.** Population genetic analysis of microsatellite variation of guppies (*Poecilia reticulata*) in Trinidad and Tobago: evidence for a dynamic source–sink metapopulation structure, founder events and population bottlenecks. *Journal of Evolutionary Biology* **22**: 485–497.
- Beerli P, Felsenstein J. 1999.** Maximum likelihood estimation of migration rates and population numbers of two populations using a coalescent approach. *Genetics* **152**: 763–773.
- Beerli P, Felsenstein J. 2001.** Maximum likelihood estimation of a migration matrix and effective population sizes in subpopulations by using a coalescent approach. *Proceedings of the National Academy of Sciences* **98**: 4563–4568.
- Blanchet S, Rey O, Loot G. 2010.** Evidence for host variation in parasite tolerance in a wild fish population. *Evolutionary Ecology* **24**: 1129–1139.
- Boots M, Best A, Miller MR, White A. 2009.** The role of ecological feedbacks in the evolution of host defence: what does theory tell us? *Philosophical Transactions of the Royal Society B: Biological Sciences* **364**: 27–36.
- Brooks DR, Hoberg EP. 2007.** How will global climate change affect parasite–host assemblages? *Trends in Parasitology* **23**: 571–574.
- Brown RC, Tsalavouta M, Terzoglou V, Magoulas A, McAndrew BJ. 2005.** Additional microsatellites for *Sparus aurata* and cross species amplification within the Sparidae family. *Molecular Ecology Notes* **5**: 605–607.
- Cornell SJ, Desdevises Y, Rigby M. 1999.** Evolutionary biology of host–parasite relationships: reality meets models. *Trends in Ecology and Evolution* **14**: 423–425.
- Day T, Gaham AL, Read AF. 2007.** Evolution of parasite virulence when host responses cause disease. *Proceedings of the Royal Society B: Biological Sciences* **274**: 2685–2695.
- Day T, Proulx R. 2004.** A general theory for the evolutionary dynamics of virulence. *The American Naturalist* **163**: 40–63.
- De Meeus T, Renaud F. 2002.** Parasites within the new phylogeny of eukaryotes. *Trends in Parasitology* **18**: 247–251.
- Estoup A, Jarne P, Cornuet JM. 2002.** Homoplasmy and mutation model at microsatellite loci and their consequences for population genetics analysis. *Molecular Ecology* **11**: 1591–1604.
- Excoffier L, Laval LG, Schneider S. 2005.** Arlequin ver. 3.0: an integrated software package for population genetics data analysis. *Evolutionary Bioinformatics* **1**: 47–50.
- FAO. 2009.** FishSTAT fishery statistical collections global capture production. Available at: <http://www.fao.org/fishery/statistics/global-capture-production/en> (Accessed September 2009).
- Fogelman RM, Kuris AM, Grutter AS. 2009.** Parasitic castration of a vertebrate: effect of the cymothoid isopod, *Anilocra apogonae*, on the five-lined cardinalfish, *Cheilodipterus quinquelineatus*. *International Journal for Parasitology* **39**: 577–583.
- Folmer O, Black M, Hoeh W, Lutz R, Vrijenhoek R. 1994.** DNA primers for amplification of mitochondrial cytochrome c oxidase subunit I from diverse metazoan invertebrates. *Molecular Marine Biology and Biotechnology* **3**: 294–299.
- Franch R, Louro B, Tsalavouta M, Chatziplis D, Tsigonopoulos CS, Sarropoulou E, Antonello J, Magoulas A, Mylonas CC, Babbucci M, Patarnello T, Power DM, Kotoulas G, Bargelloni L. 2006.** A genetic linkage map of the hermaphrodite teleost fish *Sparus aurata*. *Genetics* **174**: 851–861.
- Frank SA. 1996.** Models of parasite virulence. *The Quaternary Review of Biology* **71**: 37–78.
- Frank SA, Schmid-Hempel P. 2008.** Mechanisms of pathogenesis and the evolution of parasites virulence. *Journal of Evolutionary Biology* **21**: 396–404.
- Gandon S. 2002.** Local adaptation and the geometry of host–parasite coevolution. *Ecology Letters* **5**: 246–256.
- Gandon S, Capowiez Y, Dubois Y, Michalakis Y, Olivieri I. 1996.** Local adaptation and gene-for-gene coevolution in a metapopulation model. *Proceedings of the Royal Society B: Biological Sciences* **263**: 1003–1009.
- Gandon S, Michalakis Y. 2002.** Local adaptation, evolutionary potential and host–parasite coevolution: interactions between migration, mutation, population size and generation time. *Journal of Evolutionary Biology* **15**: 451–462.
- Gandon S, Nuismer S. 2009.** Interactions between genetic drift, gene flow, and selection mosaics drive parasite local adaptation. *The American Naturalist* **173**: 212–224.
- Garrey J, Maxwell H. 1982.** Infestation of the jack mackerel, *Trachurus declivis* Jenyns, with the cymothoid isopod

- Ceratothoa imbricata* (Fabricus), in south eastern Australian waters. *Journal of Fish Biology* **20**: 341–349.
- Gassman AJ, Stock SP, Carriere Y, Tabashnik BE. 2006.** Effect on entomopathogenic nematodes of the fitness cost of resistance to Bt toxin Cry1 Ac in pink bollworm (Lepidoptera: Gelechiidae). *Journal of Economic Entomology* **99**: 920–926.
- Gomulkiewicz R, Thompson JN, Holt RD, Nuismer SL, Hochberg ME. 2000.** Hot spots, cold spots, and the geographic mosaic theory of coevolution. *The American Naturalist* **156**: 156–174.
- Goñi R, Adlerstein S, Alvarez-Berastegui D, Forcada A, Renones O, Criquet G, Polti S, Cadiou G, Valle C, Lenfant P, Bonhomme P, Pérez-Ruzafa A, Sánchez-Lizaso JL, García-Charton JA, Bernard G, Stelzenmüller V, Planes S. 2008.** Spillover from six western Mediterranean marine protected areas: evidence from artisanal fisheries. *Marine Ecology Progress Series* **366**: 159–174.
- Goudet J. 2001.** FSTAT, a program to estimate and test gene diversities and fixation indices, version 2.9.3. Available at: <http://www2.unil.ch/popgen/softwares/fstat.htm>
- Hendry AP, Day T, Taylor EB. 2001.** Population mixing and the adaptive divergence of quantitative traits in discrete populations: a theoretical framework for empirical tests. *Evolution* **55**: 459–466.
- Hochberg ME, Michalakis Y, de Meeus T. 1992.** Parasitism as a constant on the rate of life-history evolution. *Journal of Evolutionary Biology* **5**: 491–504.
- Horton T. 2000.** *Ceratothoa steindachneri* (Isopoda: Cymothoidae) new to British waters with a key to north-east Atlantic and Mediterranean *Ceratothoa*. *Journal of the Marine Biological Association of the United Kingdom* **80**: 1041–1052.
- Horton T, Okamura B. 2003.** Post-haemorrhagic anaemia in sea bass *Dicentrarchus labrax* caused by blood feeding of *Ceratothoa oestroides* (Isopoda: Cymothoidae). *Journal of Fish Diseases* **26**: 401–406.
- Ketmaier V, Joyce DA, Horton T, Mariani S. 2008.** A molecular phylogenetic framework for the evolution of parasitic strategies in cymothoid isopods. *Journal of Zoological Systematics and Evolutionary Research* **46**: 19–23.
- Kohler T, Perron GG, Buckling A, van Delden C. 2010.** Quorum sensing inhibition selects for virulence and cooperation in *Pseudomonas aeruginosa*. *PLoS Pathogens* **6**: 5.
- Kraljević M, Dulčić M, Cetinić P, Pallaoro A. 1996.** Age, growth and mortality of the striped sea bream, *Lithognathus mormyrus* L., in the Northern Adriatic. *Fisheries Research* **28**: 361–370.
- Kuparinen A, Merila J. 2007.** Detecting and managing fisheries-induced evolution. *Trends in Ecology and Evolution* **22**: 652–659.
- Lloret J, Planes S. 2003.** Condition, feeding and reproductive potential of white seabream *Diplodus sargus* as indicators of habitat quality and the effect of reserve protection in the northwestern Mediterranean. *Marine Ecology Progress Series* **248**: 197–208.
- Mennerat A, Nilsen F, Ebert D, Skorping A. 2010.** Intensive farming: evolutionary implications for parasites and pathogens. *Evolutionary Biology* **37**: 59–67.
- Miller HC, Allendorf F, Daugherty CH. 2010.** Genetic diversity and differentiation at MHC genes in island populations of tuatara (*Sphenodon* spp.). *Molecular Ecology* **19**: 3894–3908.
- Miller SA, Dykes DD, Polesky HF. 1988.** A simple salting out procedure for extracting DNA from human nucleated cells. *Nucleic Acids Research* **16**: 1215–1215.
- Mladineo I, Šegvić I, Grubišić L. 2009.** Molecular evidence for the lack of transmission of the monogenean *Sparicotyle chrysophrii* (Monogenea, Polyopisthocotylea) and isopod *Ceratothoa oestroides* (Crustacea, Cymothoidae) between wild bogue (*Boops boops*) and cage-reared seabream (*Sparus aurata*) and sea bass (*Dicentrarchus labrax*). *Aquaculture* **295**: 160–167.
- Møller AP, Martin-Vivaldi M, Soler JJ. 2004.** Parasitism, host immune defence and dispersal. *Journal of Evolutionary Biology* **17**: 603–612.
- Møller AP, Merino S, Brown CR, Robertson RJ. 2001.** Immune defense and host sociality: a comparative study of swallows and martins. *The American Naturalist* **158**: 136–145.
- Muirhead CA, Glass NL, Slatkin M. 2002.** Multilocus self-recognition systems in fungi as a cause of trans-species polymorphism. *Genetics* **161**: 633–641.
- Nagasawa K. 2004.** Sea lice, *Lepeophtheirus salmonis* and *Caligus orientalis* (Copepoda: Caligidae), of wild and farmed fish in sea and brackish waters of Japan and adjacent regions: a review. *Zoological Studies* **43**: 173–178.
- Nielsen EE, Hemmer-Hansen J, Larsen PF, Bekkevold D. 2009.** Population genomics of marine fishes: identifying adaptive variation in space and time. *Molecular Ecology* **18**: 3128–3150.
- Nuismer SL, Thompson JN, Gomulkiewicz R. 2000.** Coevolutionary clines across selection mosaic. *Evolution* **54**: 1102–1115.
- van Oosterhout C. 2008.** The guppy as a conservative model: implications of parasitism and inbreeding for reintroduction success. *Conservation Biology* **22**: 228.
- van Oosterhout C, Hutchinson WF, Wills DPM, Shipley P. 2004.** MICRO-CHECKER: software for identifying and correcting genotyping errors in microsatellite data. *Molecular Ecology Notes* **4**: 535–538.
- van Oosterhout C, Mohammed RS, Hansen H, Archard GA, McMullan M, Weese DJ, Cable J. 2007.** Selection by parasites in spate conditions in wild Trinidadian guppies (*Poecilia reticulata*). *International Journal of Parasitology* **37**: 805–812.
- Paterson S, Vogwill T, Buckling A, Benmayor R, Spiers AJ, Thomson NR, Quail M, Smith F, Walker D, Libberton B, Fenton A, Hall N, Brockhurst MA. 2010.** Antagonistic coevolution accelerates molecular evolution. *Nature* **464**: 275.
- Petit E, Excoffier L, Mayer F. 1999.** No evidence of bottleneck in the postglacial recolonization of Europe by the noctule bat (*Nyctalus noctula*). *Evolution* **53**: 1247–1258.

- Räsänen K, Hendry AP. 2008.** Disentangling interactions between adaptive divergence and gene flow when ecology drives diversification. *Ecology Letters* **11**: 624–636.
- Raymond B, Sayyed AH, Wright DJ. 2005.** Genes and environment interact to determine the fitness costs of resistance to *Bacillus thuringiensis*. *Proceedings of the Royal Society B: Biological Sciences* **272**: 1519–1524.
- Rozas J, Sánchez-DeIbarrio JC, Messeguer X, Rozas R. 2003.** DnaSP, DNA polymorphism analysis by the coalescent and other methods. *Bioinformatics* **19**: 2496–2497.
- Sala-Bozano M, Ketmaier V, Mariani S. 2009.** Contrasting signals from multiple markers illuminate population connectivity in a marine fish. *Molecular Ecology* **18**: 4811–4826.
- Sala-Bozano M, Mariani S. 2011.** Life history variation in a marine teleost across a heterogeneous seascape. *Estuarine Coastal and Shelf Science* **92**: 555–563.
- Sala-Bozano M, Tsalavouta M, Mariani S. 2008.** Isolation and characterisation of new polymorphic microsatellite markers for the striped sea bream (*Lithognathus mormyrus*). *Conservation Genetics* **10**: 1507–1509.
- Sandland GH, Minchella DJ. 2003.** Cost of immune defence: an enigma wrapped in an environmental cloak? *Trends in Parasitology* **19**: 571–574.
- Schierup MH, Mikkelsen AM, Hein J. 2001.** Recombination, balancing selection and phylogenies in MHC and self-incompatibility genes. *Genetics* **159**: 1833–1844.
- Schioedte JC, Meinert FR. 1883.** Symbolae ad Monographiam Cymothoarum Crustaceorum Isopodum familiae, III: Saophridae, IV: Ceratohoinae. *Natuhistorisk Tidsskrift* **13**: 281–378.
- Schwarzenbach GA, Ward PI. 2006.** Responses to selection on phenoloxidase activity in yellow dung flies. *Evolution* **60**: 1612–1621.
- Simkova A, Lafond T, Ondrackova M, Jurajda P, Ottova E, Morand S. 2008.** Parasitism, life history traits and immune defence in cyprinid fish from Central Europe. *BMC Evolutionary Biology* **8**: 29.
- Suau P. 1970.** Contribución al estudio de la biología de *Lithognathus (=Pagellus) mormyrus* L. (Peces espáridos). *Investigación Pesquera* **34**: 237–265.
- Tamura K, Dudley J, Nei M, Kumar S. 2007.** MEGA4: molecular evolutionary genetics analysis (MEGA) software version 4.0. *Molecular Biology and Evolution* **24**: 1596–1599.
- Thompson JN. 1999.** The evolution of species interactions. *Science* **284**: 2116–2118.
- Thompson JN, Cunningham BM. 2002.** Geographic structure and dynamics of coevolutionary selection. *Nature* **417**: 735–738.
- Wild G, Costain G, Day T. 2007.** An epidemiological context for the consequences of phenotypic plasticity in host–pathogen interactions. *Evolutionary Ecology Research* **9**: 221–238.
- Wolinska J, King KC. 2009.** Environment can alter selection in host–parasite interactions. *Trends in Parasitology* **25**: 236–244.
- Zibiden M, Haag R, Ebert D. 2008.** Experimental evolution of field populations of *Daphnia magna* in response to parasite treatment. *Journal of Evolutionary Biology* **21**: 1068–1078.

SUPPORTING INFORMATION

Additional Supporting Information may be found in the online version of this article:

Figure S1. Network representing cytochrome oxidase subunit I (COI) sequences for *Ceratohoa italica*. The size of the pies is proportional to the frequency of the haplotype they represent: black, samples from the Balearic (BA) population; grey, samples from the Tyrrhenian (TYR) population. Small black circles represent the number of mutational steps between haplotypes.

Table S1. General linear model (GLM) of hepatosomatic index (HSI) with age, sex and population of origin ('Pop') as explanatory variables. This model was used to calculate the residuals of HSI, which were then regressed against the parasite load.

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