1 Original Article

A method for identifying sensitivity of marine benthic invertebrates to ocean acidification through a biological traits approach

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9 Abstract

Ocean acidification poses a major threat to the structure and diversity of marine ecosystems. The 10 marine seabed sustains important ecosystem functions and so understanding the sensitivity to 11 12 increased pCO_2 within benthic invertebrates is critical for informing future management strategies. 13 Here, we explore a traits-based approach for estimating sensitivity of benthic taxa to ocean acidification, using data from the western area of the North Sea. We selected 56 taxa across 11 14 15 taxonomic groups representative of the various habitats found in the region. Biological traits considered sensitive to elevated pCO₂ were identified from literature review and the taxa were scored 16 for each trait to produce a total relative sensitivity (TRS) index. We investigated differences in 17 18 sensitivity between the taxa and across habitats and explored whether sensitivity was spatially aggregated. Our analyses indicated that benthic species are sensitive to acidification, with 51 % of 19 20 the taxa scoring in the top three TRS bands. overall with hot spots of sensitivity distributed across 21 the region but corresponding 'cold spots' (low sensitivity) more spatially restricted. The opportunities 22 and limitations of the approach are discussed.

23 Introduction

Since the industrial revolution our planet has been transformed with an extent and pace not seen since the Cretaceous-Paleogene mass-extinction (Zeebe *et al.*, 2016). One manifestation of this is the change to ocean chemistry driven by anthropogenic increases in atmospheric CO₂, with a 30% increase in surface ocean acidity since the pre-industrial era, and projections of a pH of between 8.05 and 7.75 by 2100 (EEA, 2019). Such drastic changes to ocean chemistry are expected to cause significant and lasting damage to marine ecosystems, particularly among benthic invertebrates (Kroeker *et al.*, 2011).

Benthic organisms play a vital role in maintaining the marine environment, contributing to 31 sedimentary organic matter cycling through bioturbation (Aller & Cochran, 2019), and inorganic 32 nutrient exchange with the water column (Griffiths et al., 2017). Benthic invertebrates are also a key 33 34 component of the diet of many pelagic organisms, notably during the juvenile life stage of commercially important fish species (Hüssy et al., 1997). In the event of the realisation of ocean 35 acidification (OA) projections, it is likely that the ability of some benthic invertebrates to continue 36 37 these roles will be compromised, potentially further disrupting marine ecosystems and the vital 38 services they provide (Barbier, 2017). The ecological importance and likely vulnerability of these organisms are why it is critical to determine the spatial distribution of specific groups likely to be most 39 sensitive to OA, to identify areas in need of specific conservation efforts (Degraer et al., 2019). As 40 41 highlighted by Degraer et al., the North Sea is one such area which could benefit from further 42 analysis.

Biological traits analysis (BTA) is a well-established methodology for assessing the impact of stressors on organisms, populations, and ecosystems (Bremner, 2008; Foden *et al.*, 2013; Kenny *et al.*, 2018). Organisms are considered according to attributes of their biology and ecology rather than phylogeny, allowing for easier mechanistic links between impacts and traits, rather than impacts and

species. The transferrable nature of this analysis is particularly appropriate for first-stage 47 assessments of the sensitivity of marine taxa to OA, since it utilises available natural history data 48 and does not require the impractical and laborious task of laboratory testing on multiple groups of 49 50 species. Traits-based approaches have been used by others to estimate sensitivity to climate change (Foden et al., 2013; Hare et al., 2016), though they do not always consider OA in detail or provide 51 empirical justification for the traits selected as measures of sensitivity. Here, we adapted these 52 climate vulnerability approaches to develop a sensitivity metric that considers OA specifically and 53 focuses on traits for which the link between the trait and OA-sensitivity is evidenced in the literature. 54

55 To explore how our metric performs in a spatial context, we used a large dataset from the UK North Sea: an aggregation of publicly available benthic data which have already been utilised in research 56 aimed at informing and influencing management of the North Sea seabed (Cooper et al., 2019; 57 Cooper & Barry, 2017; Cooper & Barry, 2020). Our underlying premise is that some benthic species 58 exhibit traits that make them particularly susceptible to OA, and by evaluating and ranking species 59 across these traits we can calculate a metric of sensitivity that is applicable across groups and 60 61 geographies. For the metric to be useful in this context, it should be able to separate species across a sensitivity gradient and to detect spatial variability in sensitivity between benthic communities. We 62 thus explored these two properties, investigating differences in sensitivity between taxonomic 63 64 groups, across habitats and in terms of spatial clustering.

65 Methodology

66 Species Data Preparation

67 The OneBenthic database (https://openscience.cefas.co.uk/ob_obdetgc/) held, at the time of data extraction in April 2020, approximately 650,000 samples from the seas surrounding the UK, with 68 collection dates ranging from 1969-2015. Many samples within this database were initially intended 69 70 for use in surveys relating to infrastructure and extraction projects, such as oil and gas exploitation or aggregates extraction. Entries are heavily clustered due to this, leading to a highly variable 71 72 coverage across the North Sea. Extensive cleaning of the data was required to remove unsuitable or superfluous information. This involved removing samples outside of our study area, ensuring 73 uniformity of collection methods (sieve of 1 mm mesh size and quantitative sampling gears), and to 74 75 further reduce bias in the analysis we also removed all samples collected before 01/01/2000. The extents of the North Sea were taken as defined by the International Hydrographic Organisation 76 77 (1953), and a shapefile sourced from Marineregions.org.

78 After these cleaning stages, 2832 taxa remained. Biological traits analysis can be a time-consuming 79 process for very large taxon lists, particularly when traits are scored for the first time. Such a large 80 number of taxa was not practical for the aims of the study, which was to understand the potential of the metric. We thus focussed on the most abundant taxa. The data were sorted by EUNIS habitat 81 82 class (https://www.eea.europa.eu/data-and-maps/data/eunis-habitat-classification-1), with the 30 83 most abundant taxa from each habitat retained. Records that could not be identified to at least genus 84 level were removed from the dataset (because it is difficult to assign biological traits to mixed groups), 85 and the next most abundant taxon was chosen to replace it. The final database used for further analysis was composed of a list of 56 taxa and 76,036 entries. 86

87 Biological Traits Identification, Collation & Scoring

88 Traits that the scientific community consider to be relevant to species' responses to elevated ocean pCO₂ were identified through a literature review, using terms such as 'benthic invertebrates', 'ocean 89 90 acidification', and 'sensitivity', using the Google Scholar search engine. Traits listed or described in the records returned by the literature review were collated, with the criterion that they must be 91 92 evidentially linked to a response to the stressor, and discrete. We selected six traits that have been found to mediate the species level of risk to a change in pCO_2 in the environment, four of which were 93 different from our final list. A major concern at this stage was the likelihood of finding high-quality 94 sources confirming these traits for each taxon catalogued. With this in mind, two traits (lifespan, 95

- 96 infaunal/epifaunal adult stage) were removed, and a further two were adapted (length of larval stage
- 97 > larval stage type and feeding; exoskeleton and chemical composition > acid sensitive structures).

The final traits chosen related to body structural integrity (the presence of acid-sensitive structures),

99 capacity to buffer against elevated internal pCO₂ (mobility & reproductive strategy) and the

- persistence of future generations (reproductive strategy and larval stage) (Table 1; Supplementary material: Appendix 1). Each trait was divided into three categories that describe the different ways a
- 102 trait is expressed for a given taxon and the categories were scored from 3 (most sensitive) to 1 (least
- sensitive) for their relative sensitivity to CO_2 based on the evidence from the literature review, giving
- a relative sensitivity score for each trait (RSS, Table 1).

ranging from more sensitive to increased pCO_2 (3) to least sensitive to increased pCO_2 (1). *The 1-4 scale within the Adult Mobility 107 category refers to the scale used to define movement of benthic invertebrates by Quierós et al. (2013).

Biological Trait		RSS	Justification
Acid-Sensitive Structures	External	3	Taxa with chitinous (Long <i>et al.,</i> 2019) and calcareous (Byrne & Fitzer, 2020) exoskeletons have been found to exhibit severe deformities when exposed to
	Internal	2	near-future CO ₂ concentrations, negatively impacting feeding, reproduction, movement, and mechanical protection. Calcification rates decrease in waters with elevated pCO ₂ , leading to reductions in exoskeleton size by as much as 50% (Díaz-Castañeda <i>et al.</i> , 2019) Differences in relative sensitivity scores between external and internal acid- sensitive structures was determined by assessing the likelihood of exposure to increased acidity in the surrounding medium
	None	1	Theoretical absence of negative effects associated with absence of acid- sensitive structures.
Adult Mobility (Quierós <i>et al.</i> Scores)	Sessile (*1&2)	3	Organisms capable of higher levels of activity are subject to more anaerobic respiration than slow moving or sessile taxa. Organisms that regularly respire anaerobically are more capable of buffering increased pCO ₂ to prevent acidosis and are more efficient at removing excess internal CO ₂ (Melzner <i>et al.</i> , 2009), with those displaying a more hypometabolic mode of life between and within taxonomic groups capable of less acid-base regulation (Pörnter, 2008). More active taxa have been found to be broadly more resistant to rapid increases in oceanic pCO ₂ throughout other mass extinction events (Knoll <i>et al.</i> , 1996), and to be less susceptible to ocean conditions projected by the year 2100 (Kelly & Hofmann, 2012).
	Slow (*3)	2	
	Free (*4)	1	
Reproduction Strategy	Broadcast	3	Evidence that increased ocean pCO ₂ can decrease fertilisation success, as a reduction in sperm velocity and intracellular Ca ²⁺ oscillations decrease the probability of gamete fusion per collision (Shi <i>et al.</i> , 2017; Colen <i>et al.</i> , 2012).
	External Brooding	2	Brooding species have been found to outcompete non-brooding species in environments with elevated pCO ₂ and show a higher ability to adapt to environmental changes (Lucey <i>et al.</i> , 2015). Brooders are often exposed to elevated pCO ₂ due to a concentration of respiratory waste, with this fluctuating particularly intensely in internal brood chambers, where parental-stress-induced isolation can result in the pH of the chamber falling as low as 7.46 (Cole <i>et al.</i> , 2016). The requirement of brooding organisms to withstand high acidity during this stage is thought to allow for greater tolerance to high pCO ₂ later in life (Gray <i>et al.</i> , 2019).
	Internal Brooding	1	
Larval Stage	Planktotrophic	3	Evidence that taxa with pelagic larval stages are more sensitive to increases in ocean pCO ₂ than direct developing counterparts (Lucey <i>et al.</i> , 2015), with the transition from pelagic larvae to the benthic adult form being energetically taxing and creating a survivability bottleneck (Díaz-Castañeda <i>et al.</i> , 2019). Algae exposed to projected levels of ocean pCO ₂ have less nutritional value, with lower protein and organic contents, requiring compensatory feeding (Duarte <i>et al.</i> , 2016). This increases the risk of larval malnutrition, potentially causing severe latent effects (Pechenik & Tyrell, 2015).
	Lecithotrophic	2	Lecithotrophic larvae either derive nutrition from maternal yolk reserves or are non-feeding at this stage (Kempf & Hadfield, 1985), and so are not subject to feeding-related stress. Studies have found lecithotrophs to both benefit (Dupont <i>et al.</i> , 2010) and be severely disadvantaged (Verkaik <i>et al.</i> , 2016) by exposure to low pH conditions.
	Direct Development or Non-Pelagic	1	Not subject to the stresses of planktonic development.

¹⁰⁵ Table 1: Biological traits that influence benthic species' sensitivity to ocean acidification. RSS = relative sensitivity score (RSS),

Each taxon was then given a score for each trait according to the RSS scoring system, using traits 109 information generated from a literature search in Google Scholar. Information was obtained from 110 more than 70 unique sources, the vast majority of which (89%) being peer-reviewed (Supplemental 111 material: Appendix 2). Where information of this standard was not available for taxa, details were 112 inferred from close relatives or gathered from reputable non-peer-reviewed sources, such as 113 Encyclopaedia Britannica. Each entry in the resulting traits data matrix was prescribed a subjective 114 confidence level to reflect the quality of the evidence (criteria can be found in supplemental material: 115 Appendix 2). While these confidence scores were not formally included in our analysis, we suggest 116 117 they are a useful tool for identifying which entries could be improved in the future.

The scores for each taxon were summed across the four traits to give a total relative sensitivity 118 (TRS). The TRS ranges from a theoretical minimum of 4 (no sensitive structures, free-moving and 119 insensitive reproduction) to a maximum of 12 (sensitive structures, sessile, sensitive reproduction). 120 All traits were weighted equally, as confidence in the differences in sensitivity between traits was not 121 sufficient for different weightings to be considered valid and justified. Data management for taxa and 122 123 samples was undertaken using the R statistical and programming environment and the package tidyverse (Wickham et al., 2019, R Core Team 2021). Data were originally split by individual taxa for 124 each sample. To prepare for spatial analysis, single scores were calculated for samples based on 125 the mean TRS of all catalogued taxa included, resulting in a final list of 11.845 samples. This has 126 the advantage of producing a single, spatially defined figure for TRS that is easier to process and 127 analyse, however, has the potential to exaggerate differences through its sensitivity to outliers. 128

129 Data Analysis

Data were imported into ArcMap 10.6.1 and displayed through the creation of a shapefile feature 130 class containing sensitivity information. To identify any clusters of significantly higher or lower 131 sensitivity scores, the ArcMap tool Hot Spot Analysis (Getis-Ord Gi*) was utilised, with the Zone of 132 Indifference method of conceptualisation of spatial relationships used (Esri, 2021). This tool analyses 133 134 features within the context of surrounding features, where the local sum of a feature and its neighbours are compared proportionally to the sum of all features. If the local sum is significantly 135 higher, then the feature is considered to be within a hot spot. Identifying such areas assists in 136 establishing which locations may have a higher proportion of benthic invertebrates that are 137 potentially more vulnerable to increased ocean pCO₂. 138

139 **Results**

140 Sensitivities Amongst Taxa

The benthic taxa TRS scores ranged from 6-12, out of the theoretical scoring range of 4-12. Sensitivity was generally high across taxa, with 55 % of the taxa scoring 10 or higher. Amphipods, bivalves and polychaetes differed in their TRS (K-W H = 19.219, p < 0.05), with Bivalvia having a higher mean TRS than both Polychaeta (Dunn-Bonferroni x = 16.118, p = 0.02) and Amphipoda (Dunn-Bonferroni x = -26.171, p < 0.01), while Amphipoda and Polychaeta were not significantly different (Dunn-Bonferroni x = -10.054, p = 0.19; Figure 1).

To delve deeper into the drivers of the sensitivity scores, Figure 2 details the spread of relative 147 148 sensitivity scores (RSS) across the four traits. High (3) RSS were dominant across all four traits. Both reproduction-based traits (reproduction strategy and larval stage) were similar in the 149 proportions of RSS. Within the adult mobility trait, 98 % of taxa catalogued had high and medium (2) 150 RSS, with only one taxon, Scalibregma inflatum, being awarded a low score. The acid-sensitive body 151 structure category was near-evenly split between high (3) and low (1) sensitivity, with similarly only 152 one taxon receiving a medium score. Three groups (Amphipoda, Bivalvia, Polychaeta) were 153 represented to a much greater extent than the others, so much of the information in the figure was 154 155 driven by these taxa. Further analysis was conducted to reveal which traits drove these differences in sensitivity; there were no significant differences in scoring for the adult mobility trait (KW H = 0.971. 156 p = 0.615), however, the acid-sensitive structures (K-W H = 25.553, p < 0.01), larval stage (K-1 H = 157

20.708, p < 0.01), and reproduction strategy (K-W H = 18.130, p < 0.01) traits each demonstrated 158 significant differences. The polychaetes were significantly less sensitive than the other two groups 159 in terms of acid-sensitive structures (Dunn-Bonferroni x = 17.679, p < 0.01) and amphipods had a 160 significantly lower sensitivity than polychaetes and bivalves in terms of reproductive strategy 161 (Amphipods-Polychaetes: Dunn-Bonferroni x = -17.571, p < 0.01) (Amphipods-Bivalves: Dunn-162 Bonferroni x = 18.3, p = 0.01) and larval stages (Amphipods-Polychaetes: Dunn-Bonferroni x = -163

- 20.571, p < 0.01) (Amphipods-Bivalves: Dunn-Bonferroni x = 23.4, p < 0.01; Figure 3). 164
- 165

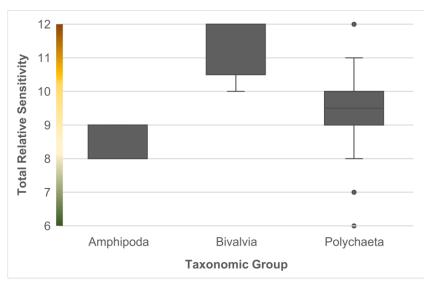


Figure 1: Total Relative Sensitivity of the three best-represented taxonomic groups: Amphipoda (N =7), Bivalvia (N=10), and 166 Polychaeta (N=28). Graduated colours on the Y axis represent high (brown), medium (yellow), and low (green) relative sensitivity values.

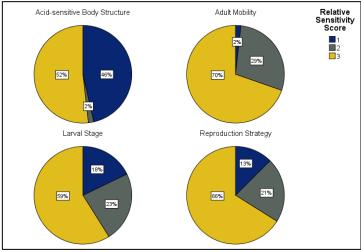


Figure 2: Total Relative Sensitivity of the three best-represented taxonomic groups: Amphipoda (N =7), Bivalvia (N=10), and Polychaeta (N=28). Graduated colours on the Y axis represent high (brown), medium (yellow), and low (greenre 2:

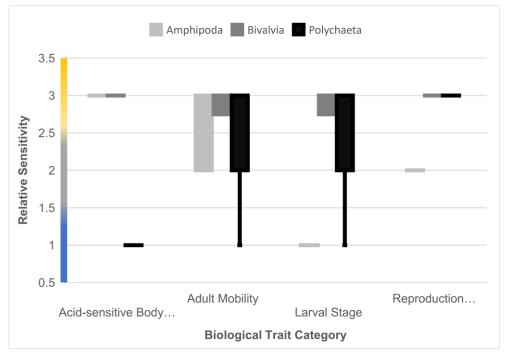


Figure 3: Relative sensitivity scores for each biological trait category across the three most well-represented taxonomic groups. Graduated colours on the Y axis represent high (amber), medium (grey), and low (blue) relative sensitivity values.

167 Differences Between Habitats

Figure 4 summarises the spread of mean TRS across EUNIS habitat types. There was a significant 168 difference in mean TRS between many of the habitats (Kruskal-Wallis H = 663.055, p < 0.01), with 169 Dunn-Bonferroni pairwise comparisons showing that A5.37 (deep circalittoral mud) had significantly 170 higher sensitivity than 8 of the other 11 EUNIS habitats (Figure 1). Notably, none of the significant 171 pairings included the two least represented EUNIS habitat types, A5.33/34 and A5.43 (infralittoral 172 sandy mud and infralittoral mixed sediments, respectively), possibly due to the few samples available 173 for analysis between these and the better-represented habitat type; the most and least represented 174 175 habitat types, A5.15 (deep circalittoral coarse sediments) at 3120 samples and A5.33/34 at 22 samples respectively, differed by two orders of magnitude. 176

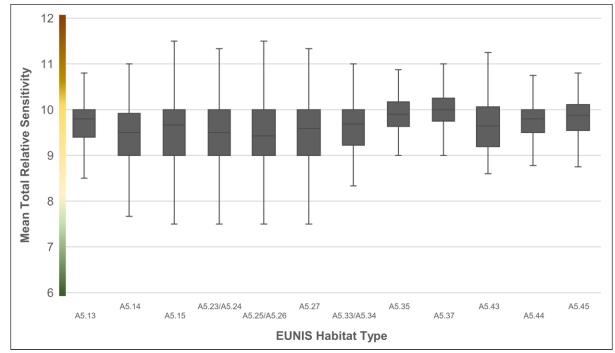


Figure 4: Boxplot of Mean Total Relative Sen4: Boxplot of Mean Total Types. Graduated colours on the Y axis represent high (amber), medium (grey), and low (blue) relative sensitivity values.

177 Spatial clustering

Figure 5 presents the geographic spread of mean TRS across the North Sea case study region, giving an overview of the distribution of samples (i.e., benthic communities) and their respective mean TRS across the community at each location. Some areas were relatively homogenous in terms of sensitivity (e.g. the moderately sensitive areas off the northeast English coast), while other regions such as England's south-east coast hosted a heterogenous mix of high, medium, and low sensitivity

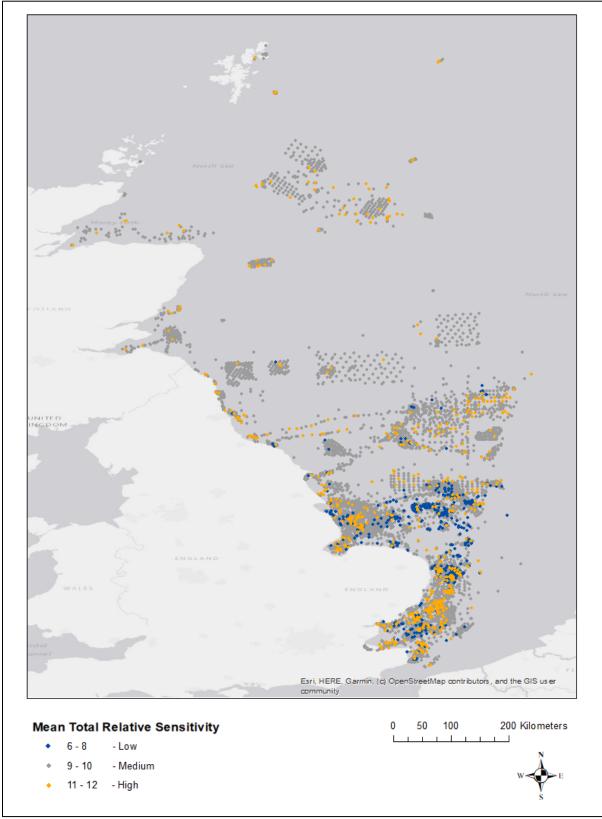


Figure 5: Samples plotted within the North Sea, with symbol colour indicating benthic invertebrate mean relative sensitivity.

communities. There was evidence of spatial clustering of sensitivity; Hot Spot Analysis (Getis-Ord Gi*) identified a large clustering of sensitivity hotspots along the Essex and Suffolk coast (Figure 6, A), the Humber Estuary (Figure 6, B), the eastern part of Dogger Bank (Figure 6, C), south of the Shetland Islands (Figure 6, D), and across the east coast of Northern England and Scotland (Figure 6). Statistically significant cold spots (low-sensitivity communities) were largely confined to the Norfolk coast seawards (Figure 6, E), extending as far as the south of Dogger Bank. There were

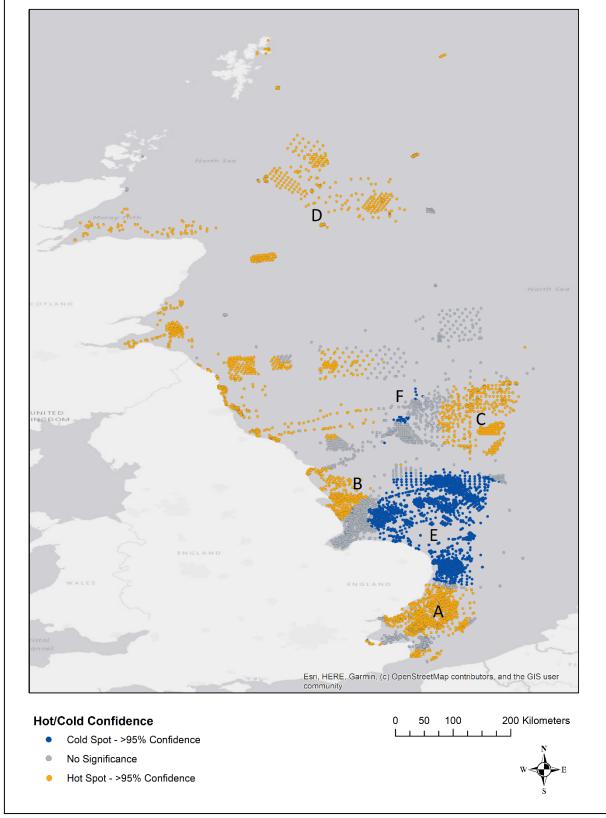


Figure 6: Output of a Hot Spot Analysis (Getis-Ord Gi*) of mean relative sensitivity of samples within the North Sea, displaying the locations of statistically significant clusters of high and low mean relative sensitivity values (p < 0.05).

several scattered cold spots at the western edge of Dogger Bank, but none in the northern UK NorthSea.

191 Discussion

Using a biological traits-driven sensitivity assessment methodology has allowed us to estimate 192 vulnerability to OA across a range of benthic invertebrate taxa within a spatial framework that 193 identifies locations of increased or decreased sensitivity. Application to a subset of taxa from the 194 western North Sea has shown the potential of the approach for identifying risks. Within our dataset. 195 sensitivity was estimated to be highest in bivalves and in communities inhabiting deep circalittoral 196 muds - a habitat typically dominated by polychaetes, with some bivalves also present (EEA, 2022). 197 Sensitivity was often spatially heterogenous at a local scale, with locations of high, medium and low 198 sensitivity found in proximity to each other, but larger-scale homogeneity (hotspots and coldspots) 199 were suggested across the region. The case study application used a subset of the total invertebrate 200 complement of North Sea benthos communities and as such is not being proposed as a definitve 201 answer on the sensitivity of the North Sea benthos; here it serves to illustate the power of the 202 approach. 203

204 A biological traits-based approach to assessing sensitivity across a large and diverse set of taxa provides a way of simplifying the complexities that come with ecological analysis on such a scale. 205 206 This method strikes a good compromise between practicality and the applicability of results to realworld ecological communities. However, this kind of analysis is heavily dependent upon high quality, 207 208 detailed traits information being available for many species (Degen et al., 2018). This was mitigated 209 somewhat by studying a subset of taxa, a strategy used in similar research (Hare et al., 2016). Moreobscure taxa may lack a precise description of the target attributes, as was found while conducting 210 this study. In cases such as these, biological traits may be inferred through the traits of closely related 211 species, such as within the same genus or family. This presents its own set of challenges, however, 212 213 as some traits may be inconsistent between even closely related species. While the uncertainty of 214 the precision or accuracy of sources can be quantified somewhat through confidence scores, there is yet no clear method for incorporating these into analyses. Despite this, it is a way of dealing with 215 the use of biological traits which is well established (Bolam et al., 2014). 216

One of the key drivers for choosing the biological traits categories was to attempt to capture the 217 nuances of sensitivity to increased ocean pCO_2 , by cataloguing and scoring a diverse set of 218 219 biological attributes linked to sensitivity to this change (see Table 1). The adult mobility trait, while being of sound scientific reasoning, did not in the end assist in capturing this nuance. Compared to 220 other traits used it is relatively homogeneous, with a large majority (70 %) of catalogued taxa falling 221 into the sessile category, and only one taxon assessed to be completely free moving (Figure 2). The 222 223 link between mobility of a marine organism and their ability to maintain internal pCO_2 is established (Knoll et al., 1996, Melzner et al., 2009; Table 1). However, as demonstrated by the results of this 224 study, levels of mobility within some benthic invertebrates are not particularly diverse, and so may 225 not be subject to many of the benefits theorised from high levels of exercise-induced anaerobic 226 227 respiration. This could, however, be a result of the particular cross-section of taxa found for this study: almost half of which were polychaetes - annelid worms often confined within tube structures. 228 With a wider selection of taxa, and better representation within other taxonomic groups, it is possible 229 230 that a more even spread would be seen for this trait category, yielding results that are more useful 231 in quantifying overall sensitivity. We propose that future studies based upon this general methodology would benefit from investigating a larger pool of biological traits, and incorporating a 232 'traits estimation' stage for selected taxa, where expected traits for each selected taxon are compiled 233 234 to assess diversity at an earlier stage.

Of the 11 taxonomic groups catalogued and included in wider analysis, eight were represented by three or fewer taxa, with six of these including only one taxon each (Supplemental material: Appendix 1). While the array of taxa was chosen through analysing abundance data for each EUNIS habitat classification, the final group was lacking in diversity. Although this still may be representative of these habitats, it allows for only a superficial understanding of taxonomic groups other than the three most represented here. Analysis of the three most catalogued taxonomic groups found that bivalves were likely to have the highest sensitivity to increases in ocean pCO₂. Knowledge that this group could be subject to more-extreme deleterious effects than other benthic groups could inform management strategies; while individual management actions at the local scale cannot mitigate widespread OA, there may be opportunities to take actions to reduce other pressures acting on the species that can be locally controlled or mitigated.

246 In our study dataset, deep circalitoral muddy habitat was estimated to contain a significantly larger proportion of taxa with biological traits sensitive to increases in ocean pCO₂. According to the 247 European Environment Agency (2022), this habitat is usually dominated by polychaetes, and home 248 249 to a high number of bivalves; in our dataset the top 10 taxa were predominantly polychaetes. This prevalence of sensitive taxa gives an indication of what the future could bring for these habitats. 250 Evidence is emerging on the importance of muddy sediments as carbon stores and of disruption to 251 the seabed as a potential source of further CO₂ emissions (Legge et al., 2020); OA impacts on 252 253 sensitive species have the potential to disrupt ecological functioning in the deep muds since benthic invertebrates are important contributors to carbon cycling through bioturbation and other activities 254 255 (Aller & Cochran, 2019). OA thus has the potential to alter how carbon is incorporated into or released from seabed sediments, the nature of which requires further investigation into its wider 256 257 impacts.

258 Conclusion

We believe that the calculation of the TRS methodology shows strong potential for application in 259 future studies of a similar nature. Clear separation of likely sensitive and non-sensitive taxa can be 260 achieved through the selection of traits that are relatively evenly shared through selected species -261 an aspect for improvement in this study, which could be remedied by a 'traits estimation' stage 262 before the finalisation of trait selection. Clusters of significantly higher and lower relative sensitivity 263 264 were found. Knowledge of the locations of hotspots could be used to inform future conservation 265 efforts, with the aims of managing other anthropogenic stressors that could exacerbate acidityrelated impacts on the ecosystem. High overall sensitivity among catalogued taxa indicates a high 266 risk of impact on the most abundant benthic organisms if projections of future ocean pCO_2 are 267 268 realised. A useful next step in the research would be to acquire a deeper view of sensitivity within different taxonomic groups in marine benthic invertebrate communities through an expansion of the 269 number and diversity of taxa catalogued, enabling enhanced accuracy of sensitivity estimation. 270 271 This methodology is well placed for performing similar analyses for different stressors, areas, and populations, which could include comparisons of locations that already experience higher levels of 272 acidification such as upwelling zones, methane seeps or hydrothermal vents. 273

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