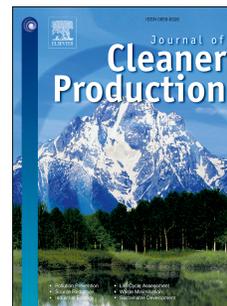


Accepted Manuscript

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PII: S0959-6526(17)31287-8

DOI: [10.1016/j.jclepro.2017.06.117](https://doi.org/10.1016/j.jclepro.2017.06.117)

Reference: JCLP 9867

To appear in: *Journal of Cleaner Production*

Received Date: 8 May 2016

Revised Date: 14 June 2017

Accepted Date: 14 June 2017

Please cite this article as: Shao L, Guan D, Wu Z, Wang P, Chen GQ, Multi-scale input-output analysis of consumption-based water resources: Method and application, *Journal of Cleaner Production* (2017), doi: 10.1016/j.jclepro.2017.06.117.

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1 **Multi-scale input-output analysis of consumption-based water**
2 **resources: method and application**

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17 **ABSTRACT**

18 This work develops a method of multi-scale input-output analysis for the embodied
19 water accounting of an economy. This method can distinguish between the different
20 virtual water contents of imported and local products and is therefore capable of
21 accurately estimating the virtual water that is embodied in trade. As a simplified
22 model rather than a multi-regional input-output analysis, this method substantially

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23 minimizes the data requirements. With the support of averaged Eora global embodied
24 water intensity databases for the world and Chinese economies, a three-scale
25 embodied water input-output analysis of the Beijing economy in 2007 has been
26 conducted. Dozens of virtual water flows that relate to the Beijing economy have been
27 identified and analyzed. Only 15% of the total water resources embodied in Beijing's
28 local final demand were from local water withdrawal; 85% were from domestically
29 and internationally imported products. The virtual water import is revealed to play a
30 more important role than physical water transfer in easing Beijing's water shortage.
31 Since the average water use efficiency of the Beijing economy is much higher than
32 that of the Chinese economy but somewhat lower than that of the rest of the world, Beijing
33 is suggested to shifting its imports to foreign countries to optimize global water use.
34 The method developed can be useful for water saving strategies for multiple
35 responsible entities holding different opinions, and it can be easily applied to the
36 embodied water accounting of a sub-national or even smaller economic community.

37

38 **Keywords**

39 Multi-scale input-output analysis; Consumption-based water resources; Water
40 resources embodied in trade; Virtual water strategy

41

42 **1. Introduction**

43 With the acceleration of globalization, cooperation, and specialization in the world
44 economy, ever-increasing international and interregional trade have brought more

45 trade imbalance issues concerning water resources. The water resources that are
46 embodied in export, import, and bilateral trade have therefore attracted a significant
47 amount of attention (Dalin et al., 2014; Hoekstra and Mekonnen, 2012). Among these
48 studies, input-output analysis (IOA), and increasingly multi-region input–output
49 (MRIO) analysis, plays an important role in tracing consumption-based water
50 resource usage and in determining responsibility (Deng et al., 2015; Guan et al., 2014).
51 Some water indicators, for example, the Water Stress Index (WSI) and the Water
52 Exploitation Index (WEI) (Lenzen et al., 2013a; Pfister et al., 2009; Zhao et al., 2015)
53 have also been attached to IOA to assist in the policy making.

54 Because the data support on a national level is relatively complete, the
55 consumption-based water resources of countries around the world and the virtual
56 water that is embodied in international trade have been successfully analyzed and
57 estimated by MRIO analyses (Chen and Chen, 2013; Lenzen et al., 2013a). Based on
58 available statistic trade data, some studies have made great contribution by connecting
59 a national MRIO model to a global MRIO model with limited number of country
60 groups to analyze regional virtual water flows (Cazcarro et al., 2016; Feng et al., 2011;
61 Yu et al., 2010). However, for other sub-national or ever smaller regional economies
62 where the detailed input-output data are not available, a complete water resource
63 MRIO still presents difficulties. For example, the virtual water MRIO studies of
64 Chinese regions either ignored international imported virtual water (Jiang et al., 2015;
65 Zhao et al., 2015) or assumed that imported products and the locally produced
66 products are homogenous (Deng et al., 2016).

67 Because of the lack of available data, some IOA based studies used the
68 single-region input-output (SRIO) model to estimate the virtual water flow of an
69 economy. The SRIO-based studies have either ignored imported virtual water (Zhao
70 et al., 2016) or have made the water intensity for imports equal to that of local product
71 (Guo et al., 2016; Renault, 2003; Zhao et al., 2009). Therefore, the SRIO can only
72 estimate how much water is saved by importing a product instead of producing
73 domestically, i.e., the water resources use that is avoided by imports (WAI). And it
74 prevents estimation of the real virtual water transfer and the design of oriented virtual
75 water strategies.

76 First presented by Chen and his colleagues, the method of multi-scale input-output
77 (MSIO) was proposed to construct carbon emission intensity database and to support
78 carbon emission accounting of a building (Chen et al., 2011). The present work
79 contributes a universal multi-scale input-output (MSIO) method to analyze embodied
80 water flows of an economy. It can distinguish between the different virtual water
81 contents of the same products from different economies, which is superior to SRIO. In
82 addition, because the averaged embodied water intensity databases for the world and
83 for national economies have been used to estimate the virtual water that is embodied
84 in trade, this method requires much less data than a complete MRIO analysis does.
85 The MSIO can be easily applied to account the resource use of a sub-national or even
86 a smaller economy according to individual need.

87 Beijing, as a megacity and the capital of China, has long suffered from water
88 shortage because of a series of both natural and social economic factors, such as an

89 unbalanced spatial distribution, heavy water pollution, a large population and rapid
90 economic growth. It is reported that the groundwater reserves of Beijing (1.74 billion
91 m³ in 2015) has been declining over the past a few decades, the amount of which has
92 been reduced by 9.48 billion m³ since 1980 (BWA, 2015). The rivers with relative
93 heavy water pollution (lower than V class water standard defined by Ministry of
94 Environmental Protection of China) account for 39.7% of the total river length in
95 2015 (BWA, 2015). Aiming to demonstrate the MSIO method of water resources and
96 assist water saving strategy design of Beijing, a case study has been conducted for a
97 three-scale embodied water input-output analysis of the Beijing economy. The rest of
98 the paper is organized as follows: Methodology and data sources are elaborated in
99 Section 2. Section 3 presents the results of embodied water accounting of the Beijing
100 economy. Section 4 discusses the advantages of the MSIO and contributes some
101 virtual water strategies for Beijing to save water resources withdrawal. Finally,
102 conclusions are drawn in Section 5.

103

104 **2. Methods and materials**

105 **2.1 Multi-scale input-output (MSIO) analysis**

106 **2.1.1 Multi-scale ecological input-output table of an economy**

107 To distinguish between the virtual water contents of local and imported products, a
108 non-competitive ecological extended input-output table is considered as the basis of
109 the MSIO analysis. Table 1 shows the general form of a multi-scale ecological
110 input-output table of an economic system.

111 The illustrated system is divided into n_0 industrial sectors, which are
 112 simultaneously sustained by local and imported products. The imports are assumed to
 113 be from m external economic systems, and the economic system in the S_h scale is
 114 divided into n_h industrial sectors. $z_{i,j}^L$ and $z_{i,j}^{S_h}$ represent local products and
 115 imported product of the S_h scale from Sector i that are used as intermediate inputs for
 116 local Sector j . y_i^L and $y_i^{S_h}$ represent local products and imported products of the S_h
 117 scale from Sector i to satisfy the local final use. The local products from local Sector i
 118 that satisfy external use in the external economic system are represented as exports,
 119 e_i^L . Imported products are also re-exported to the external economic systems (see
 120 $e_i^{S_h}$). Finally, x_i represents the total output of the i -th local sector, and $F_{k,j}$ represents
 121 the k -th water resources flow that is consumed by the j -th local sector.

122

123 Table 1

124 Multi-scale ecological input-output table of an economic system¹

Input	Output	Intermediate use			Final use		Output
		Sector 1	...	Sector n_0	Local final consumption	Export	
Local inputs	Sector 1	$z_{1,1}^L$...	z_{1,n_0}^L	y_1^L	e_1^L	x_1
	⋮	⋮	...	⋮	⋮	⋮	⋮
	Sector n_0	$z_{n_0,1}^L$...	z_{n_0,n_0}^L	$y_{n_0}^L$	$e_{n_0}^L$	x_{n_0}
Imported inputs from S_1	Sector 1	$z_{1,1}^{S_1}$...	$z_{1,n_0}^{S_1}$	$y_1^{S_1}$	$e_1^{S_1}$	
	⋮	⋮	...	⋮	⋮	⋮	
	Sector n_1	$z_{n_1,1}^{S_1}$...	$z_{n_1,n_0}^{S_1}$	$y_{n_1}^{S_1}$	$e_{n_1}^{S_1}$	
⋮	⋮	⋮	⋮	⋮	⋮		

	Sector 1	$z_{1,1}^{S_m}$...	$z_{1,n_0}^{S_m}$	$y_1^{S_m}$	$e_1^{S_m}$
Imported inputs from S_m	\vdots	\vdots	...	\vdots	\vdots	\vdots
	Sector n_m	$z_{n_m,1}^{S_m}$...	$z_{n_m,n_0}^{S_m}$	$y_{n_m}^{S_m}$	$e_{n_m}^{S_m}$
Direct water resources withdrawal	Water type 1	$F_{1,1}$...	F_{1,n_0}		
	\vdots	\vdots	...	\vdots		
	Water type p	$F_{p,1}$...	F_{p,n_0}		

125 ¹ The added value part is omitted as it is not involved in the basic formulation.

126

127 2.1.2 The basis formulas of multi-scale input-output analysis

128 Figure 1 shows the water resources input and output flows of a typical industrial
 129 sector i in an economic system (only the k -th water flow is considered). $\varepsilon_{k,j}^L$ and
 130 $\varepsilon_{k,j}^{S_h}$ represent the k -th embodied water resource intensities of local products from
 131 Sector j and imported products from Sector j of the economic system in the S_h scale.
 132 Here, the embodied water intensity is principally defined as the marginal water cost
 133 when the total amount of a product changes by an incremental unit.

134

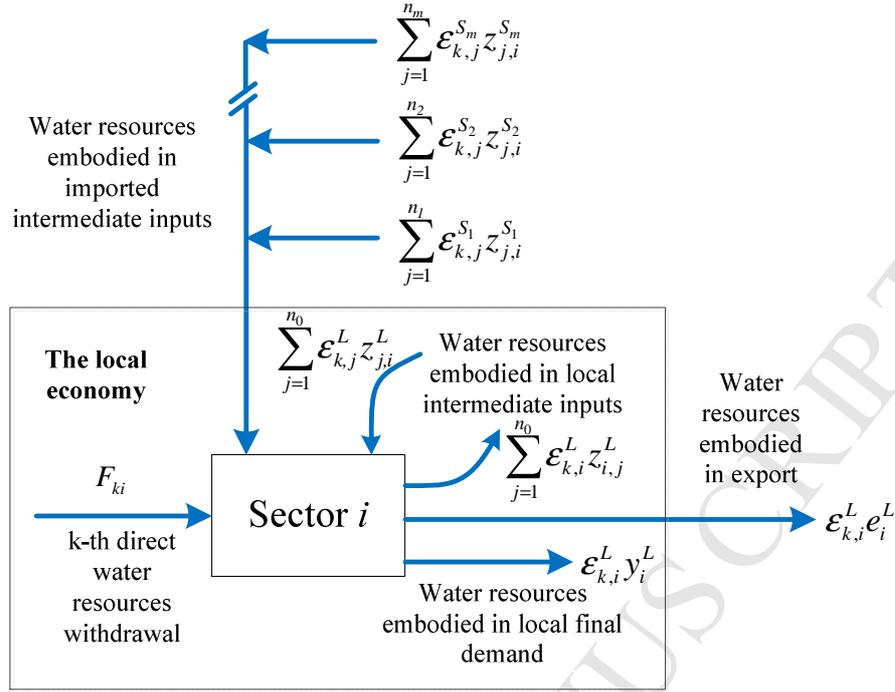


Figure 1. The virtual water balance of an industrial sector in an economic system

According to conservation law, the physical embodied water balance equation for

local Sector i can be written as

$$F_{k,i} + \sum_{j=1}^{n_0} (\epsilon_{k,j}^L z_{j,i}^L) + \sum_{h=1}^m \sum_{j=1}^{n_h} (\epsilon_{k,j}^{S_h} z_{j,i}^{S_h}) = \epsilon_{k,i}^L \left(\sum_{j=1}^{n_0} z_{i,j}^L + y_i^L + e_i^L \right). \quad (1)$$

Based on the basic economic balance of an input-output table (see Table 1), it is

understood that

$$x_i = \sum_{j=1}^{n_0} z_{i,j}^L + y_i^L + e_i^L. \quad (2)$$

For the entire economic system that is complemented by water resource flows of p types as an economic-ecological integration, an aggregate matrix equation can be

induced as

$$\mathbf{F} + \boldsymbol{\epsilon}^L \mathbf{Z}^L + \sum_{h=1}^m (\boldsymbol{\epsilon}^{S_h} \mathbf{Z}^{S_h}) = \boldsymbol{\epsilon}^L \mathbf{X}, \quad (3)$$

in which

$$\begin{aligned}
149 \quad & \mathbf{F} = [F_{k,i}]_{p \times n_0}, \\
150 \quad & \boldsymbol{\varepsilon}^L = [\boldsymbol{\varepsilon}_{k,i}^L]_{p \times n_0}, \quad \mathbf{Z}^L = [z_{i,j}^L]_{n_0 \times n_0}, \quad \boldsymbol{\varepsilon}^{S_h} = [\boldsymbol{\varepsilon}_{k,i}^{S_h}]_{p \times n_h}, \quad \mathbf{Z}^{S_h} = [z_{i,j}^{S_h}]_{n_h \times n_0}, \\
151 \quad & \mathbf{X} = [x_{i,j}]_{n_0 \times n_0}, \text{ where } x_{i,j} = x_i \text{ when } i=j, \text{ and } x_{i,j} = 0 \text{ when } i \neq j. \\
152
\end{aligned}$$

153 2.1.3 The embodied water intensity database based on the MSIO analysis

154 The embodied water intensity matrix for local products, as the basis for estimating
155 the virtual water resource flows, can be given based on Eq. (3) as

$$156 \quad \boldsymbol{\varepsilon}^L = (\mathbf{F} + \sum_{h=1}^m (\boldsymbol{\varepsilon}^{S_h} \mathbf{Z}^{S_h})) (\mathbf{X} - \mathbf{Z}^L)^{-1}. \quad (4)$$

157 It can be seen that the direct water withdrawal data, non-competitive economic
158 input-output data and the embodied water intensity matrixes for imported products are
159 the preconditions for obtaining the embodied water intensity database of local
160 products. As long as the embodied water intensity database of local products is
161 obtained, the water resources that are embodied in the final demand and the exports
162 can be calculated accordingly. After several deformations, Eq. (4) can also be
163 rewritten as the following to make it easier to understand:

$$164 \quad \boldsymbol{\varepsilon}^L = (\mathbf{F} + \sum_{h=1}^m (\boldsymbol{\varepsilon}^{S_h} \mathbf{Z}^{S_h})) \mathbf{X}^{-1} (\mathbf{I} - \mathbf{A}^L)^{-1}, \quad (5)$$

165 where $\mathbf{A}^L = \mathbf{Z}^L \mathbf{X}^{-1}$ is a coefficient matrix that describes the inputs in the production
166 of these sectors, and $(\mathbf{I} - \mathbf{A}^L)^{-1}$ is the so-called Leontief inverse (Miller and Blair,
167 2009). The equation is formally the same as the basis formula in universal
168 environmental input-output analysis (EIOA). In EIOA, only $\mathbf{F} \mathbf{X}^{-1}$ as direct water use
169 coefficients is applied to calculate the water resources embodied in final demand.

170 However, besides the local water withdrawal, some of the water resources that are

171 embodied in local products have originated from external economies by means of
 172 imported intermediate input. Therefore, in Eq. (5), the “direct” water use coefficients
 173 include not only the coefficients that originate from direct local water withdrawal
 174 ($\mathbf{F}\mathbf{X}^{-1}$) but also the coefficients that originate from external water imports
 175 ($\sum_{h=1}^m (\boldsymbol{\epsilon}^{S_h} \mathbf{Z}^{S_h}) \mathbf{X}^{-1}$). They together constitute the total water coefficients (internal and
 176 external) of local products. $\boldsymbol{\epsilon}^{L,L} (\mathbf{F}\mathbf{X}^{-1} (\mathbf{I} - \mathbf{A}^L)^{-1})$ and $\boldsymbol{\epsilon}^{L,S_h}$
 177 ($\sum_{h=1}^m (\boldsymbol{\epsilon}^{S_h} \mathbf{Z}^{S_h}) \mathbf{X}^{-1} (\mathbf{I} - \mathbf{A}^L)^{-1}$) are used to denote local (with direct water withdrawal)
 178 and external (with imported intermediate input in S_h scale) embodied water intensities
 179 of local products.

180 2.2 Data source

181 Based on the MSIO method illustrated in Section 2.1, the three-scale embodied
 182 water input-output analysis of the Beijing economy in 2007 is performed as a case
 183 study. There are three types of essential databases for MSIO analysis. The first
 184 database is a direct water use matrix for the Beijing economy in 2007. The ecological
 185 input-output analysis assumes that the water transfer in an economic system has been
 186 approximately covered by the input-output data among different industrial sectors.
 187 For example, the tap water fee indicating the water transfer from the water supply
 188 industry to other industries has been itemized as the intermediate output of the water
 189 supply industry in the statistics data. As the physical entry of water resources to
 190 human society, the direct water withdrawal data – rather than direct water
 191 consumption – are therefore concerned in this work. The statistical data that was

192 contributed by the Second National Economic Census of China has been referred to
 193 (see SI, Table S4) (CECY, 2010). Total water resource consumption has been divided
 194 into three types of water usage according to different withdrawal purposes.
 195 Agricultural water is for agricultural water requirements, mainly irrigation. Industrial
 196 water is for industrial production processes, for example, washing coal or cooling
 197 generator systems. Domestic water refers to daily water use by residents.

198 The second database is the economic input-output table. The official economic
 199 input-output table for the Beijing economy in 2007 has been adopted (BSB, 2009).
 200 This monetary input-output table is in units of Chinese Yuan (referred to as Yuan
 201 hereafter). There are 42 industrial sectors in the table, which can be found in Support
 202 Information (SI), Section S1.1. It is a competitive input-output table that does not
 203 distinguish between local and imported products. By assuming that the imported
 204 products have been distributed to intermediate input and final use with the same ratio
 205 of local products (Guan et al., 2014; Shao et al., 2016; Weber et al., 2008), it is
 206 transformed into a non-competitive input-output table. The intermediate input-output
 207 matrix and final use matrix of local products, domestic imported products, and foreign
 208 imported products can be obtained accordingly. For example, the foreign-imported
 209 intermediate input-output matrix, \mathbf{Z}^M , can be calculated as

$$210 \quad z_{i,j}^M = z_{i,j} \left(\frac{x_i^M}{x_i + x_i^D + x_i^M} \right) \quad (6)$$

211 where $z_{i,j}$ is the total intermediate input from Sector i to Sector j , x_i is the total output
 212 of Sector i , x_i^D is the domestically imported economic flow of Sector i , and x_i^M is

213 the foreign-imported economic flow of Sector i . Meanwhile, the final demand of
 214 Sector i from foreign imports, \mathbf{F}^M , is expressed as

$$215 \quad f_{i,k}^M = f_{i,k} \left(\frac{x_i^M}{x_i + x_i^D + x_i^M} \right) \quad (7)$$

216 where $f_{i,k}$ is the k -th final consumption of Sector i .

217 The last data source concerns the embodied water intensity databases for the
 218 domestic and foreign imported products of the Beijing economy. The results from an
 219 MRIO analysis to estimate the embodied water intensity database for 26 sectors of
 220 186 countries all over the world in 2007 on the basis of the Eora global MRIO
 221 database have been referred to (Lenzen et al., 2012; Lenzen et al., 2013b; Shao, 2014).
 222 According to the multi-scale input-output scheme, in principle, the databases should
 223 represent the rest of the world excluding the Chinese economy (because the products
 224 from the Chinese economy are considered to be domestic imported products) and the
 225 rest of China excluding the Beijing economy to estimate the virtual water that is
 226 embodied in foreign and domestic imports. With respect to the former, the average
 227 embodied water intensity for the 26 sectors in the other 185 nations of the world
 228 economy has been applied to estimate water resources embodied in foreign imported
 229 products (see SI, Table S2). However, regarding the latter, excluding the Beijing
 230 economy from the Chinese database is difficult. Therefore, the embodied water
 231 intensity for 26 sectors of the Chinese economy, as part of the results of the global
 232 MRIO analysis, has been applied to estimate to estimate water resources embodied in
 233 domestic imported products (see SI, Table S3). Because the share of the Beijing

234 economy is a small part of the Chinese economy (the GDP ratio is approximately 3%),

235 the deviation that the assumption presents can be ignored.

236

237 3. Results

238 3.1 The embodied water intensity of the Beijing economy

239 In 2007, Beijing has withdrawn 3.21 billion m³ water. Meanwhile, the
 240 consumption-based water resources of Beijing is estimated as 11.28 billion m³ water,
 241 approximately 3.51 times higher than the production-based water resources

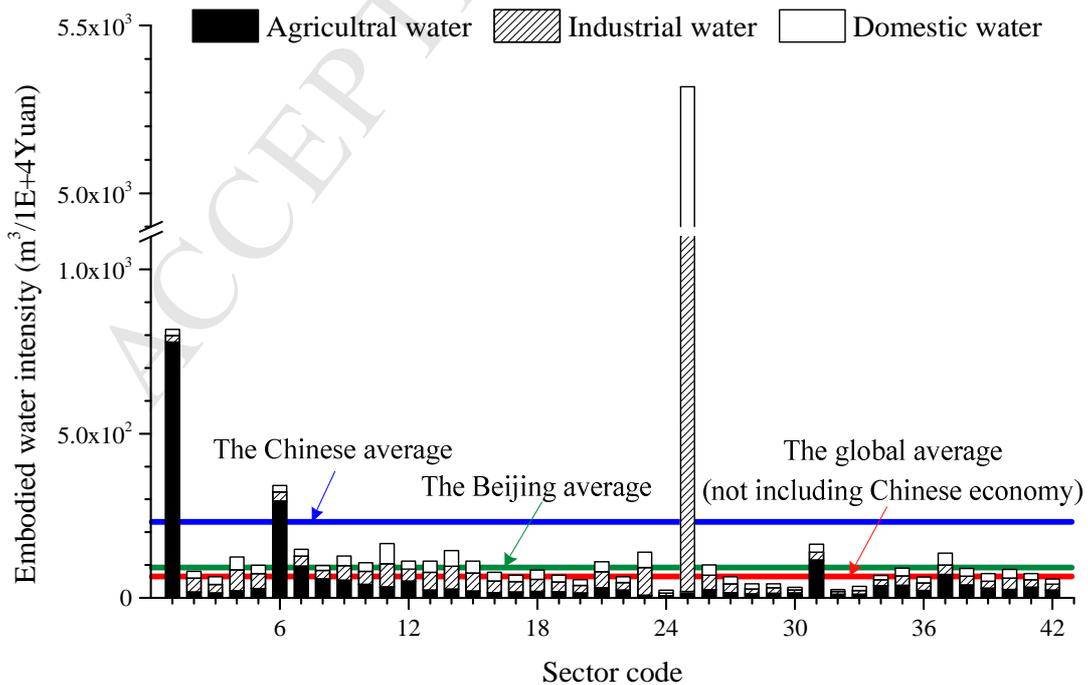
242 withdrawal. The embodied water intensities of all economic sectors in the Beijing

243 economy in 2007 in the form of water resource withdrawal initiated by the final

244 consumption of one unit product (this work uses the monetary unit) are shown in

245 Figure 2 (the values can be found in SI, Table S5).

246



247

248 Figure 2. The embodied water intensity of the Beijing economy in 2007

249

250 Among the 42 industrial sectors, Sector 25 (Production and Distribution of Water)
251 has the highest embodied water intensity, and Sector 1 (Agriculture, Forestry, Animal
252 Husbandry and Fishery) ranks second. It is because these two industries have
253 withdrawn abundant raw water resources, accounting for more than 85% of the total
254 freshwater withdrawal. The final products that are supplied by these sectors, such as
255 tap water, foods, and tobacco, are water-intensive products. Beijing can increase
256 imports of these products to save local water resources. Although many industries
257 seem irrelevant to agriculture, every sector has embodied agricultural water intensity.
258 The advantage of input-output analysis is that it can reveal the complex technical and
259 economic interactions of different industries.

260 A key theoretical premise of this study is that, because of the variation in
261 technological efficiency and the diversity of economic structure, the same type of
262 products that are produced in different communities have different embodied water
263 intensities, i.e., virtual water contents. The obtained results agree with this premise
264 (see the comparisons of Tables S2, S3, and S4 in SI). The average output-valued
265 embodied water intensity for the entire global economy (not including the Chinese
266 economy), the Chinese economy, and the Beijing economy in 2007 have been
267 estimated as $71.69 \text{ m}^3/1\text{E}+04 \text{ Yuan}$, $235.72 \text{ m}^3/1\text{E}+04 \text{ Yuan}$, and $91.97 \text{ m}^3/1\text{E}+04$
268 Yuan (the former two are calculated from the embodied water intensity database of
269 the Chinese economy and the rest of the global economy in 2007; see Figure 2). It can

270 be observed that the average water use efficiency of the Beijing economy is much
271 higher than that of the Chinese economy but somewhat lower than that of the rest of the
272 world. Thus, on average, China needs more than twice as many water resources as
273 Beijing to have the same economic output value.

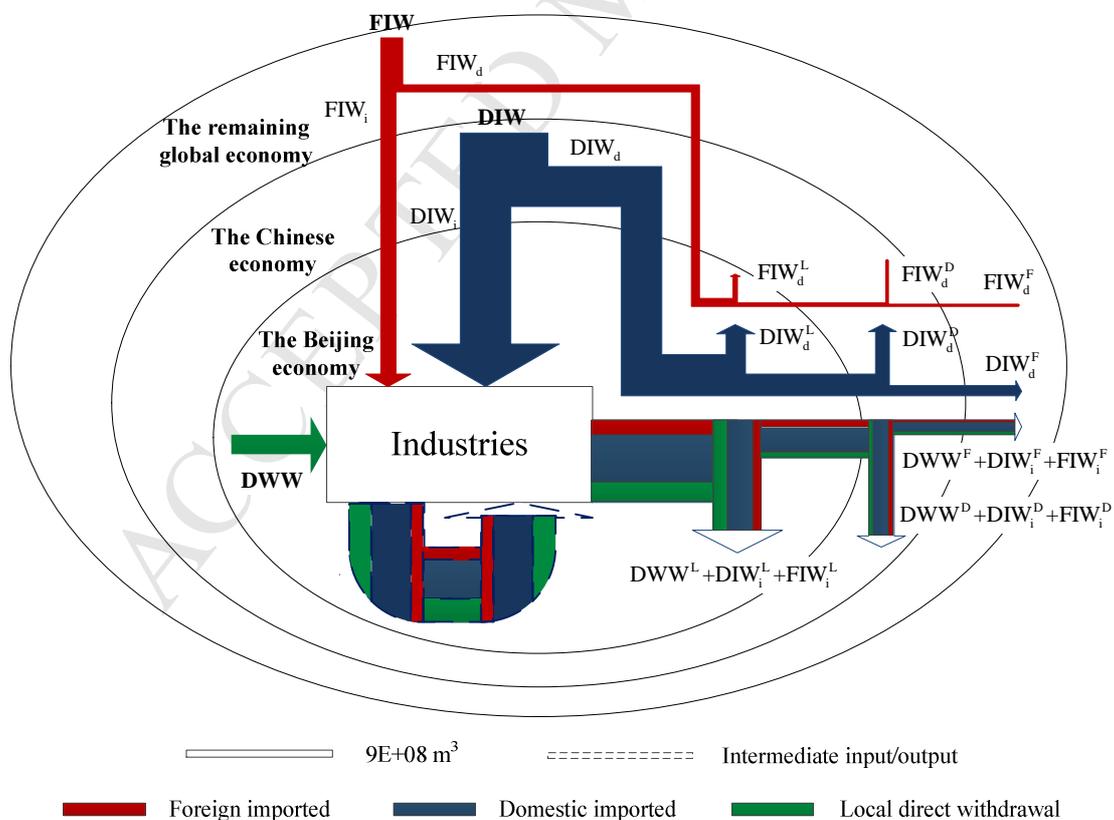
274 The significant gap between the Chinese and Beijing economies is mainly a
275 reflection of the embodied agricultural water intensity, where China's average value is
276 $161.59 \text{ m}^3/1\text{E}+04 \text{ Yuan}$ and Beijing's average value is only $41.04 \text{ m}^3/1\text{E}+04 \text{ Yuan}$
277 (which is even lower than the average value of the rest of the world). In 2012, over 88%
278 of the total agricultural area in Beijing had adopted water-efficient irrigation, and the
279 coefficient of irrigation agricultural water was as high as 0.70 in 2013 (Geng, 2014).
280 At the same time, agricultural water use in Beijing has continually dropped since 2001
281 (CSY, 2015). With regard to China, the coefficient was only 0.52 in 2013 (Geng,
282 2014). The Chinese government plans to increase this coefficient to over 0.60 in 2030
283 according to the national policy referred to as "the Strictest Water Resources
284 Management System" (SCC, 2012); the implementation of this policy was also
285 emphasized in a recently launched draft of the five-year-plan for China. Smooth
286 implementation of the plan will contribute to a lower embodied water intensity of
287 China.

288 **3.2 Water resources embodied in the final demand of the Beijing economy**

289 One of the most important core issues in ecological economics is that the use of
290 resources, such as water resource withdrawal, and environmental emissions, such as
291 greenhouse gas emissions, are driven by human final demand or final consumption.

292 All embodied water flows related to the Beijing economy have been calculated and
 293 are illustrated in Figure 3 (the definitions and values of the embodied flows are shown
 294 in SI, Table S6). Essentially, the final consumption of Beijing is simultaneously met
 295 by foreign imported, domestic imported and local products. However, local products,
 296 which were emphasized in Section 2.1, involve local, domestic, and global embodied
 297 water intensities through direct water withdrawal, domestic, and foreign imported
 298 intermediate inputs, respectively. Therefore, part of the embodied water of the local
 299 products has also originated from domestic and foreign imported embodied water.
 300 This is often ignored in previous studies, and the MSIO has made it possible to
 301 identify and calculate it in the present work.

302



303

304 Figure 3. The embodied water flows of the Beijing economy in 2007 (DWW: direct water
 305 withdrawal; FIW: foreign imported water resources; DIW: domestically imported water
 306 resources; Superscript L/D/F: water resources embodied in local final demand/domestic
 307 export/foreign export; subscript d: direct imported water resources; Subscript i: indirect
 308 imported water resources through imported intermediate inputs)

309

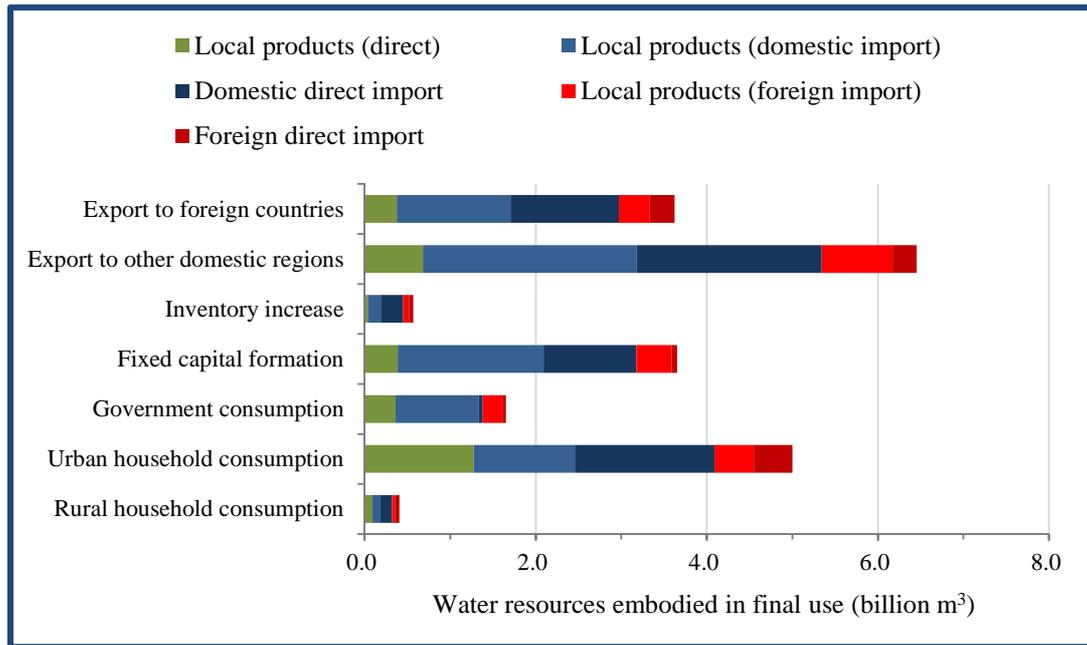
310 According to the results, less than 1/5 of the total water resources that are embodied
 311 in Beijing's local final demand have been met by direct local water withdrawal
 312 (DWW^L). Concerning domestic imported water (DIW^L), in addition to the 3.17 billion
 313 m^3 of ostensible embodied water that is supplied by direct domestic import (DIW_d^L),
 314 another 4.10 billion m^3 of water resources have also been supplied through
 315 intermediate industrial processes (DIW_i^L). The invisible or indirect embodied water of
 316 local products that originated from domestic import is even larger than the obvious or
 317 direct embodied water. Together, they supply more than approximately 3/5 of the total
 318 embodied water of Beijing's local final consumption. The situation for foreign
 319 imported water that is embodied in local final consumption is similar (FIW^L); the
 320 ostensible (FIW_d^L) and unapparent embodied water resources (FIW_i^L) are 0.65 billion
 321 m^3 and 1.20 billion m^3 , accounting for approximately 16% of the total embodied
 322 water of local final consumption.

323 Beijing's local final demand includes rural household consumption, urban
 324 household consumption, government consumption, fixed capital formation, and
 325 inventory increase. In addition, local products have also been exported to other

326 Chinese regions and foreign countries, and together, they form the final use of local
327 products (see Figure 4). The water resources that are embodied in urban household
328 consumption are approximately 12 times higher than the water resources of rural
329 household consumption. The ratio of the urban population to rural population is only
330 approximately 5.45, which reveals the significant gap between the standards of
331 resource use of urban and rural people.

332 Like local final consumption, the water resources that are embodied in total
333 domestic and foreign exports each constitute five components. Except for the
334 embodied water that is directly supplied by direct local water withdrawal (DWW^D
335 and DWW^F) and domestic (DIW_d^D and DIW_d^F) and foreign imported products (FIW_d^D
336 and FIW_d^F), a significant portion of the domestic and foreign exports' embodied water
337 that is supplied by local products also originates from domestic (DIW_i^D and DIW_i^F)
338 and foreign imported intermediate inputs (FIW_i^D and FIW_i^F). Local freshwater
339 withdrawal and domestic and foreign imports' embodied water account for 10.56%,
340 72.26%, and 17.18% of the total domestic export water. With respect to foreign
341 export water, the ratios are essentially the same. Nearly half (47%) of the 21.35
342 billion m^3 of virtual water resources related to Beijing (direct water withdrawal plus
343 domestic imported water plus domestic imported water, $DWW+DIW+FIW$) has been
344 re-exported or re-located to external economies.

345



346

347 Figure 4. Water resources embodied in the final use of the Beijing economy in 2007

348

349 **3.3 Embodied water trade balance of the Beijing economy**

350 Beijing has 14.53 billion m³ of domestic imported water from other Chinese
 351 regions, which is approximately 2.25 times greater than domestic exported embodied
 352 water. Regarding foreign trade, foreign exported embodied water is slightly higher
 353 (0.18%) than foreign imported embodied water. Domestic imported water is also 4
 354 times greater than foreign imported water. Concerning domestic and foreign exported
 355 water, the ratio is only approximately 1.78. Overall, total water resources embodied in
 356 domestic trade is 2.90 times greater than the total water resources of foreign trade.

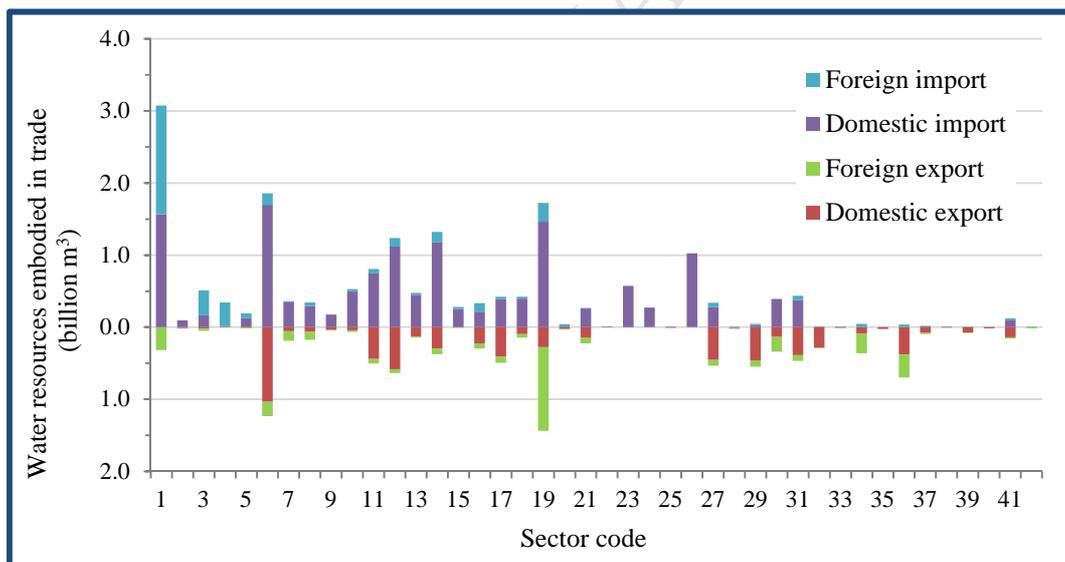
357 Figure 5 shows the water balance of each industry in the Beijing economy in 2007.

358 Overall, the primary industry and most secondary industries are embodied water net
 359 importers, and most tertiary industries are embodied water net exporters. Sector 1
 360 (Agriculture, Forestry, Animal Husbandry and Fishery) involves the largest embodied
 361 water trade and is also the largest embodied water importer. It may be because

362 agriculture is a water-intensive industry, and its embodied water intensity is the
 363 second largest after the water industry. This result also indicates that Beijing, as a
 364 megacity, is highly dependent on imported agricultural products.

365 Sector 19 (Manufacture of Communication Equipment, Computer and Other
 366 Electronic Equipment) is the second largest embodied water trader and also exports
 367 the most embodied water among all industries. Unlike the agriculture industry, Sector
 368 19's embodied water import and export are both very large, although the former is
 369 approximately 20% larger than the latter. The industry involves many high-tech
 370 products, whose trade deals have led to the massive embodied water flows.

371



372

373 Figure 5. The water trade balance of the Beijing economy in 2007

374

375 Sector 6 (Manufacture of Foods and Tobacco) ranks third and is also the second
 376 largest embodied water importer and exporter after Sectors 1 and 19, respectively. For
 377 most industries, the water that is embodied in domestic trade is greater than the water

378 that is embodied in foreign trade. Because there are typically more barriers in
379 international trade than in domestic trade, this result is reasonable. However, Sector
380 19 is the opposite. This contrast may be because the production advantage of sector
381 19 is sufficiently distinct to offset tariffs and the different preferences of domestic and
382 foreign customers.

383

384 **4. Discussion**

385 **4.1 Comparing this study to other related studies**

386 According to different measures to estimate the virtual water resources that are
387 embodied in imports, there are four IOA methods that can be applied to analyze the
388 consumption-based water resources use of an economy:

389 (1) SRIO (Single Regional Input-Output Analysis)

390 The water resources that are embodied in imports are ignored. Only the local water
391 resources are concerned.

392 (2) SRIO-r (Single Regional Input-Output Analysis-revised)

393 The water intensities for imports are assumed equal to that of local products. The
394 water resources use that is avoided by imports (WAI) rather than that is embodied in
395 imports is calculated.

396 (3) MRIO (Multi-Region Input-Output Analysis)

397 The same products from different countries are treated as completely different
398 types of products. The water resources that are embodied in imported products for
399 meeting local final demand are calculated.

400 (4) MSIO (Multi-Scale Input-Output Analysis)

401 The output-averaged embodied water intensity databases for the external
402 economies are applied to estimate the water resources that are embodied in imports.

403 In order to illustrate the differences between these IOA methods, they have been
404 applied separately to analyze the consumption-based water resources of the Beijing
405 economy in 2007. The results are compared in Table 2 (since a complete MRIO
406 analysis needs massive data to perform, here this paper only compares the results of
407 the former three methods; the differences between MSIO and MRIO are discussed
408 later in theory).

409 Table 2

410 Water resources accounting results based on different IOA methods

IOA method	Embodied water intensity (m ³ /1E+04 Yuan)	Water resources embodied in local final demand (billion m ³)	Water resources embodied in total final use (billion m ³)	Water resources embodied in domestic imports (billion m ³)	Water resources embodied in foreign imports (billion m ³)
SRIO	21.39	2.15	3.21	-	-
SRIO-r	21.39	2.15	3.21	1.06 (WAI)	2.26 (WAI)
MSIO	91.97	11.28	21.35	14.53	3.62

411

412 It can be seen that in general the results of SRIO and SRIO-r are much lower than
413 that of MSIO. It is because the SRIO and the SRIO-r have ignored and
414 underestimated the water resources that are embodied in imports. As indicated in Eq.
415 (5) and the related descriptions in Section 2.1, the embodied water intensity of Beijing
416 local products consists of local (with direct water withdrawal), domestic (with
417 domestic imported intermediate input), and global (with foreign imported

418 intermediate input) embodied water intensities (see SI, Section S2 and Figure S1).

419 The intensities obtained from SRIO and SRIO-r are equal, which is indeed the local
420 embodied water intensity of MSIO. Therefore the results of SRIO and SRIO-r are
421 only part of the results of MSIO, and the MSIO can provide us with additional
422 information on imported water resources that is embodied in local products.

423 Although in theory the MRIO is more accurate than the MSIO with respect to
424 regional resolution, the size of their basis data differs by orders of magnitude. For
425 example, in this study, on the basis of the existing world and Chinese embodied water
426 intensity datasets (which are typically drawn from previous MRIO analysis), 6,174
427 embodied water flows have been analyzed to obtain the final results. However, an
428 MRIO analysis that connects all Chinese provinces to all countries may involve over
429 70 million economic flows (assuming that each Chinese region has 30 sectors and 5
430 types of final demand and each country has 57 sectors and 3 types of final demand).
431 Nearly half of the data required, i.e., economic trade flows between Chinese region
432 and foreign countries, cannot be found in the existing databases or studies, and many
433 assumptions must be made.

434 Since the MSIO can distinguish between the different embodied water resources of
435 imported products and local products, it was proved to be superior to the SRIO.

436 However, MSIO is not a perfect method. And the present study also involves a few
437 uncertainties. For example, the aggregation error in connecting 26 sectors-based Eora
438 intensity database to 42 sectors-based Beijing input-output data. When the MRIO data
439 are not available or reliable, the MSIO can be served as a compromised but efficient

440 way to address consumption-based environmental issues. The advantage of MSIO
441 will be even more prominent when it needs to conduct an embodied water analysis for
442 a sub- national or even a small economy. The obtained embodied water intensity
443 database would be very useful for water footprint analysis of an engineering system,
444 e.g., a wastewater treatment plant, a building, too (Meng et al., 2014; Shao and Chen,
445 2013, 2016).

446 Some previous studies have analyzed the carbon emissions and embodied water of
447 the Beijing economy by MSIO (Chen et al., 2013; Guo et al., 2012; Han et al., 2015).
448 However, these studies have certain limitations. First, China's economy was not
449 excluded from the averaged database of the world economy. Second, the database for
450 the Chinese economy that was applied was derived from a single-region analysis that
451 did not distinguish between import products and local products. These limitations
452 created deviations in the final results.

453 **4.2 Virtual water strategy of the Beijing economy**

454 Beijing and other northern Chinese regions have long suffered from water shortage.
455 Given that the southern regions have more than enough water resources, the Chinese
456 government launched the nationwide "South-to-North Water Transfer Project"
457 (SNWTP) in 2002 after over fifty years of discussion and planning (Li, 2016). The
458 middle route of the SNWTP officially began operations in December 2014 after
459 twelve years of construction (Li, 2016). Approximately 1.9 billion m³ of freshwater
460 has been distributed to Beijing during the past two year, which has alleviated the local
461 water shortage and groundwater overdraft (Li, 2016). Over 70% of Beijing's tap water

462 is currently from the SNWTP (Li, 2016), and a rise in the groundwater level has been
463 observed for the first time in 16 years (STD, 2015).

464 According to the results, in 2007, Beijing domestically and internationally imported
465 14.53 and 3.62 billion m³ of embodied or virtual water. The net import is
466 approximately 8 billion m³ of virtual water (all from domestic imports). It can be seen
467 that virtual water transfer plays a more important role than physical water transfer in
468 easing Beijing's water shortage. This conclusion is consistent with that of Zhao et al.
469 (Zhao et al., 2015). Improving local water use efficiency can be considered the most
470 fundamental and sustainable measure to reduce direct water withdrawal of Beijing.
471 Since Beijing relies heavily on domestic imported water resources and the Chinese
472 average embodied water intensity is much higher than the Beijing average and the
473 global average, Beijing would also benefit a lot from the improvement of water use
474 efficiency on the domestic level. Beijing can invest in or spread water-saving
475 technology in other Chinese regions to reduce the Chinese average embodied water
476 intensity.

477 The concept of virtual water was first proposed by Allan in 1998 and has been
478 considered essential to a strategic solution for water-scarce regions (Allan, 1998).
479 Because all products (mainly agricultural products in early research) consume water
480 resources during production in the form of virtual water or embodied water, the
481 water-scarce problem can be settled through the commercial trade of water-intensive
482 products. Since the cost of physical transport of water resources is very high, virtual
483 water strategy can be considered as an alternative and better way to alleviate Beijing's

484 water shortage problem. The MSIO-based results can be very helpful in the design of
485 water saving strategies for multiple responsible entities holding different opinions.
486 Concerning reducing Beijing's local water resource withdrawal only, Beijing should
487 enlarge its imports of Sector 1 (agricultural products), 6 (foods and tobacco), 4 (metal
488 ores), and 7 (textile products). It is because these products have the largest local
489 embodied water intensities, and they can be easily substituted by imported like
490 products.

491 However, for the entirely different matter of global water saving, the production of
492 high-intensity products should be allocated to economies with high water production
493 efficiency or low water intensity. Whether Beijing should import or where it should
494 import from should also depend on water use efficiency. Since the average water use
495 efficiency of the Beijing economy is much higher than that of the Chinese economy
496 but somewhat lower than that of the rest of the world, Beijing should enlarge its
497 international imports and reduce its domestic imports.

498

499 **5. Conclusions**

500 Virtual water accounting has been a research hotspot in water resource
501 management field. This work contributed a multi-scale input-output (MSIO) analysis
502 method to account the consumption-based water resources of an economy. It can
503 distinguish the different virtual water contents of imported products and local
504 products. Compared with the single region input-output analysis (SRIO) method, the

505 MSIO can improve the accounting accuracy. Meanwhile, it requires much less data
506 than a complete MRIO analysis does.

507 The framework and algorithm of the MSIO method for water resources accounting
508 of an economy were presented in detail in Section 2.1. In order to illustrate the MSIO
509 method, a three-scale embodied water input-output analysis of the Beijing economy in
510 2007 was performed as a case study. On the basis of an Eora global MRIO analysis,
511 the output-weighting averaged embodied water resources intensity database for the 26
512 sectors in the other 185 nations (excluding China) of the world economy was applied
513 to estimate the water resources embodied in the foreign imports of Beijing.

514 Meanwhile, the database for the Chinese economy as a part of the MRIO analysis
515 results was applied to estimate the embodied water resources of Beijing's domestic
516 imports.

517 The consumption-based water resources consumption of Beijing was estimated as
518 11.28 billion m^3 , which was 3.51 times higher than the production-based water
519 resources withdrawal in 2007 (3.21 billion m^3). It is revealed that the average water
520 use efficiency of the Beijing economy (109 Yuan/ m^3) is much higher than that of the
521 Chinese economy (42 Yuan/ m^3) but somewhat lower than that of the rest of the world (139
522 Yuan/ m^3). Among the total amount of 21.35 billion m^3 water resources that were
523 related to Beijing, 15% were from direct water withdrawal, 68% were from
524 domestically imported virtual water resources, and 17% were from international
525 imported virtual water resources. Over 1/2 of these water resources have been

526 embodied in Beijing's local final demand, only 30% have been exported to other
527 regions in China, and 17% have been exported to other countries.

528 In 2007, Beijing imported 14.53 billion m³ water resources from other Chinese
529 regions, which was approximately 2.25 times greater than domestic exported
530 embodied water (6.45 billion m³). Meanwhile, the foreign imported embodied water
531 (3.62 billion m³) was more or less the same as foreign exported embodied water (3.62
532 billion m³). It is revealed that the primary industry and most secondary industries
533 were embodied water net importers, and most tertiary industries were embodied water
534 net exporters.

535 The differences between different IOA methods are discussed in the last section.
536 According to the results of this paper, the net virtual water import of Beijing was
537 approximately 8 billion m³ (all from domestic imports) in 2007. It is much larger than
538 the physical water transfer of Beijing during 2015-2016 (approximately 1.9 billion m³)
539 from the nationwide "South-to-North Water Transfer Project". The virtual water
540 transfer is revealed to play an important role in easing Beijing's water shortage.
541 According to the results, in order to save local water resources withdrawal, Beijing
542 should enlarge its imports of Sector 1 (agricultural products), 6 (foods and tobacco), 4
543 (metal ores), and 7 (textile products). On the other hand, in order to improve the water
544 use efficiency worldwide and maximize water conservation, Beijing should enlarge its
545 international imports and reduce its domestic imports.

546

547 **Acknowledgements**

548 This work is supported by the Natural Science Foundation of China (Grant Nos.
549 71503236 and 11272012), the Social Science Foundation of Beijing (No. 16LJC013),
550 the State Key Program for Basic Research of China (973 Program, No.
551 2013CB430402), and the Fundamental Research Funds for the Central Universities
552 (2652015151). Additional materials as noted in the text can be found in Supporting
553 Information.

554

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- A multi-scale input-output analysis method (MSIO) has been contributed.
- The MSIO can greatly minimize the data requirements compared to the MRIO method.
- Water resources embodied in final demand and trade of Beijing have been analyzed.
- Averaged embodied water databases for the world and Chinese economies were applied.
- More than 3/5 of local final demand's embodied water were domestically imported.